# FINAL STUDY REPORT CHARACTERIZATION OF DOWNSTREAM AQUATIC COMMUNITIES <br> RSP 3.18 

## CONOWINGO HYDROELECTRIC PROJECT

FERC PROJECT NUMBER 405


Prepared for:

## Exelon.

Prepared by:

Normandeau Associates, Inc.
Gomez and Sullivan Engineers, P.C

## EXECUTIVE SUMMARY

Exelon Generation Company, LLC (Exelon) has initiated with the Federal Energy Regulatory Commission (FERC) the process of relicensing the 573-megawatt Conowingo Hydroelectric Project (Conowingo Project). The current license for the Conowingo Project was issued on August 14, 1980 and expires on September 1, 2014. FERC issued the final study plan determination for the Conowingo Project on February 4, 2010, approving the revised study plan with certain modifications.

The final study plan determination required Exelon to conduct a literature-based study to provide a characterization of the current aquatic community below Conowingo Dam, which is the subject of this report. The objectives of this study are to utilize existing data to: 1) characterize resident fish abundance, size structure, condition, and reproductive success below Conowingo Dam-the existing data includes fish lift catches as well as results from other common fisheries gear types such as electrofishing, gill nets, and ichthyoplankton nets; 2) describe the benthic macroinvertebrate communities below Conowingo Dam collected by various common collection gears; and 3) provide updated information on these communities available through studies focused on other objectives.

An initial study report (ISR) was filed on February 22, 2011, containing Exelon's 2010 study findings. An initial study report meeting was held on March 9, 10 and 11, 2011 with resource agencies and interested members of the public. Formal comments on the ISR including requested study plan modifications were filed with FERC on April 27, 2011 by Commission Staff, several resource agencies and interested members of the public. Exelon filed responses to the ISR comments with FERC on May 27, 2011. On June 24, 2011, FERC issued a study plan modification determination order. The order specified what, if any, modifications to the ISRs should be made. For this study, FERC's June 24, 2011 order required no modifications to the original study plan. An updated study report (USR) was filed on January 23, 2012 addressing comments from stakeholders received at the March ISR meeting, those comments addressed by Exelon in the May 27, 2011 responses to ISR comments, as well as editorial and minor text changes. This final study report is being filed with the Final License Application for the Project.

From 1980 through 1991 a series of quantitative benthic studies were conducted in the non-tidal area of the Lower Susquehanna River below the Conowingo Dam to determine a release schedule sufficient to maintain healthy fish and macroinvertebrate communities. A cumulative total of 71 macroinvertebrate taxa were collected and were identified to the genus level. This result was used as a basis to characterize the invertebrate community as moderately rich. The community was generally comprised of facultative or
tolerant warm-water genera. Most abundant were: Chironomidae in the genera Cricotopus, Dicrotendipes, and Polypedilum; caddisflies in the genera Cheumatopsyche and Hydroptila; Asiatic clams in the genus Corbicula; flatworms in the genus Dugesia; a crustacean in the genus Gammarus; a snail in the genus Goniobasis; and aquatic worms in the genera Manayunkia, and Nais. The most important food items in the stomach contents of eight fish species examined were Chironomidae, Cheumatopsyche, and Gammarus.

Thirty-eight years of annual collections at the West Fish Lift (WFL) and 19 years at the East Fish Lift (EFL) facilitate description of year-to-year and long term fluctuations in catch and proportional abundance in key resident and migratory species of the Lower Susquehanna River fish assemblage from relicensing studies were conducted by Exelon from 1982 to 1987. These included electrofishing, gill net and ichthyoplankton sampling efforts from Conowingo Dam downstream to the tidal waters at Havre de Grace, Maryland. These data augment the fish lift collections by providing a more detailed spatially and temporally diverse characterization of the downstream fish populations in regards to species assemblage, condition, food habits and habitat use. Data collected in the 2010 fish stranding summer surveys in the spillway reach below Conowingo Dam also supplement the fish lift catches for a season not typically sampled by the lifts.

The dominant species documented by the macroinvertebrate and fish community studies downstream of Conowingo Dam to the tidal zone, are summarized in Table ES-1 and discussed below.

TABLE ES-1: SUMMARY OF FISH COMMUNITY AND DOMINANT SPECIES DOCUMENTED TO OCCUR DOWNSTREAM OF CONOWINGO DAM BASED ON HISTORICAL STUDIES, 1972-2010

| DATA SOURCE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | WFL | EFL | Ichthyoplankton sampling | Electrofishing | Gill Nets | Stranding Surveys |
|  | Years | $\begin{gathered} 1972 \text { to } \\ 2009 \end{gathered}$ | 1997-2009 | 1982-1984 | 1982-1987 | $\begin{aligned} & \hline 1981- \\ & 1984 \end{aligned}$ | 2010 |
|  | Total fish | 31,533,545 | 16,411,728 | $\begin{gathered} 275,710 \\ \text { (eggs) } \\ \hline \end{gathered}$ | 235,458 | 4,054 |  |
|  | Taxa | 72 | 59 | 27 | 66 | 28 | 14 |
|  | verage CPUE | 1,148/lift | 1,289/lift | n/a | 953/hr | 34/net | n/a |
|  | Gizzard Shad | 75\% | 87\% | 3\% | 49\% | 22\% | 57\% |
|  | American Eel |  |  |  | 11\% |  |  |
|  | American shad | <1\% | 8\% |  |  |  |  |
|  | White perch | 12\% | <1\% | 72\% | 12\% | 23\% |  |
|  | Blueback herring | 4\% | 4\% | 24\% (incl. alewife) |  |  |  |
|  | Channel catfish | 3\% |  |  | 4\% | 42\% |  |
|  | Banded killfish |  |  |  |  |  | 23\% |
|  | Sunfish |  |  |  | 7\% |  | 11\% |
|  | Largemouth bass |  |  |  |  |  | 4\% |
|  | Yellow Perch |  |  |  | 6\% |  |  |
|  | Other species | $\begin{gathered} \hline 5 \%(67 \\ \text { taxa) } \\ \hline \end{gathered}$ | $\begin{gathered} <1 \%(5 \\ \text { taxa) } \end{gathered}$ | $\begin{gathered} 1 \%(24 \\ \text { taxa) } \\ \hline \end{gathered}$ | 11\% (60 taxa) | $\begin{gathered} \hline 13 \%(24 \\ \text { taxa) } \end{gathered}$ | $\begin{aligned} & \hline 5 \%(10 \\ & \text { taxa) } \end{aligned}$ |

The WFL has been operated during anadromous spawning migrations since 1972 as part of a cooperative private, state, and Federal effort to restore American shad to the upper Susquehanna River. From 1972 to 2009 the WFL collected 31,533,545 fish of 72 different taxa from 27,481 lifts. The overall catch per unit effort (CPUE) of the WFL during the 38 years of operation was 1,148 fish per lift. From 1972 to 2009 gizzard shad accounted for $75 \%$ of the overall CPUE at the WFL. Gizzard shad became the dominant species in 1977 and retained its dominance over the next three decades. Overall, American shad comprised $1 \%$ of the overall collection at the West Fish Lift from 1972 to 2009. American shad CPUE at the West Fish Lift remained low throughout the 1970's but began to increase in the late 1980's and continued to increase through the 1990's; after 1996, American shad was routinely one of the top five most frequently collected species. The highest CPUE for American shad occurred in 2003 at 27 fish per lift. White perch ( $12 \%$ ), blueback herring (4\%) and channel catfish (3\%) were other species
proportionally abundant in CPUE. Sixty-seven other taxa combined for a total of $5 \%$ of the overall CPUE 1972 to 2009.

Pursuant to a settlement agreement on water quality and fish passage approved by FERC on January 24, 1989, Exelon was required to construct facilities for the protection of fish. The EFL began operation in 1991 as the cornerstone of the agreement. Beginning in 1997 all fish were lifted to the exit channel for continued volitional upstream movement following the construction and operations of fish lifts at the upstream Holtwood and Safe Harbor dams. From 1991 to 2009 the EFL collected 16,411,728 fish of 59 different taxa from 12,733 lifts. The overall CPUE from the 19 years of lift operation was 1,289 fish per lift. Gizzard shad have dominated the catch at the EFL, accounting for $86 \%$ of all fish collected since 1991. Other species proportionally abundant in CPUE at the EFL included American shad (8\%) blueback herring (4\%) and white perch (1\%). Fifty-five other taxa combined for a total approximately $1 \%$ of the overall CPUE 1991 to 2009.

From 1982 to 1984 ichthyoplankton sampling was generally performed from late March to late June to characterize the use of the lower Susquehanna River downstream of Conowingo Dam by resident and anadromous fishes as a spawning and nursery area. A total of 275,710 eggs, larvae or young from 27 taxa was collected from 1,322 icthyoplankton samples. During each year eggs were the most abundant life stage collected. White perch was the most frequently collected species, constituting $72 \%$ of all icthyoplankton observed. Alosa spp. (alewife and blueback herring) comprised $24 \%$ and gizzard shad comprised $3 \%$ of the ichthyoplankton collected. Other less commonly collected species included American shad and carp; these along with 24 other species accounted for less than $1 \%$ of the specimens collected.

Over the course of the 1982 to 1987 electrofishing series, 247 hours of effort were spent sampling the Conowingo tailrace, the riverine areas consisting of Lee's Ferry and the Pool, and the tidal zone. A representative collection was completed in every month of the year except February. A total of 235,458 fish of 66 taxa was collected in the lower Susquehanna River electrofishing program from May 1982 to October 1987. Overall, the CPUE from 1982 to 1987 was 953 fish/hr. Gizzard shad was the most proportionally abundant species comprising $49 \%$ of all fish collected, followed by white perch (12\%) and American eel (11\%). Species assemblage and proportional abundance varied seasonally in the Conowingo tailrace. Gizzard shad dominated the catch from September through December and were relatively high in proportion in April through June. Comely shiners were relatively abundant in both December and March collections. Carp were most prevalent in spring. A relatively high rate of collection for American eel was maintained throughout all months of sampling. White perch were also relatively abundant May through

October. Lee's Ferry and the Pool had very little month to month variability and white perch was the most frequently collected fish in all years. Other species frequently collected at Lee's Ferry and the Pool included: American eel, channel catfish, carp, gizzard shad, shorthead redhorse and yellow perch. At the tidal zone, the most frequently collected species varied more from year to year than the other survey areas. Gizzard shad were the most frequently collected species in 1983 and 1985. Redbreast sunfish were the most frequently collected species in 1986 and 1987. Yellow perch was the most frequently collected species in 1982 and 1984. Other species common in the tidal zone included: American eel, carp, pumpkinseed, bluegill, channel catfish, smallmouth bass, white sucker, spottail shiner and comely shiner.

Gill netting was conducted monthly from July through November in 1981 to 1984 in the same general areas as sampled by electrofishing below Conowingo Dam. A total of 4,054 fish of 28 taxa was collected from 118 gill net sets. Channel catfish was the most proportionally abundant species in CPUE (42\%) followed by white perch ( $23 \%$ ) and gizzard shad (22\%) in all areas across all years. Twenty-four taxa comprised the remaining $13 \%$ of fish collected. White perch exhibited several seasonal trends; at the Conowingo tailrace white perch catches were higher in July and August than in other months, and at Lee's Ferry white perch CPUE's were significantly higher in September. For gizzard shad the highest CPUEs consistently occurred in September and October.

Various food habits studies conducted 1982 to 1985 depict fishes below Conowingo Dam opportunistically utilizing food resources available from lotic habitats below the dam as well as forage produced in Conowingo Pond upstream. Detailed stomach analyses of individual white perch, channel catfish and yellow perch taken by electrofishing in the tailrace below Conowingo Dam July through December reported small zooplankters were abundant in white perch stomachs, but caddisfly larva (Cheumatopsyche) and chironomid larva were more important on a frequency basis, with caddis larvae most important based on percent of the biomass eaten (Weisberg and Janicki 1985). Chironomids were most important to channel catfish numerically and on a frequency basis. However, similar to white perch, caddis larva formed most of the diet biomass. The amphipod Gammarus was the most important food of yellow perch. Food resources recruited from Conowingo Pond and available to downstream fishes include zooplankton, important to both resident fishes and young of larger predators, and also young gizzard shad which sustain many of the larger resident and migratory predators below the dam.

Smallmouth bass have historically been and remain a highly sought after recreational species below Conowingo Dam. Smallmouth bass age and growth below Conowingo Dam were evaluated over a 4-year period from 1980 to 1983 (RMC 1985a). Mean fork length data depict a typical growth pattern. Based on mean FL attained by Age $4(366 \mathrm{~mm})$, most smallmouth bass were recruited to the harvestable population
below Conowingo Dam ( $\sim 305 \mathrm{~mm}$ TL) during their 4th year of life. Growth of smallmouth bass below Conowingo Dam was similar to or greater than that reported for several waters in PA and MD (RMC 1985a).

Length frequency data were collected and summarized for several species from electrofishing, gill nets and fish lifts 1982 to 1984 (RMC 1985a, $\underline{\text { b , ce }}$ ). Differences between years in length frequencies of collected fishes were likely due to varied growth or recruitment of particular year classes. In electrofishing samples, a majority of white perch in each sampling area were 141 to 170 mm , and electrofishing samples taken late in the spawning run in the spring reflected the size distribution of transient white perch adult spawners. Nearly $50 \%$ of channel catfish collected from the riverine reach and tidal zone were 151 to 200 mm . In the tailrace, the catch of channel catfish was almost exclusively juvenile ( $<250 \mathrm{~mm}$ ). The size distribution of gizzard shad collected by electrofishing in the tailrace and riverine reach was dominated by individuals exceeding 200 mm , whereas the tidal zone was dominated by smaller ( $101-150 \mathrm{~mm}$ ) gizzard shad. During electrofishing collections a majority of yellow perch collected throughout the study reach were 141 to 180 mm Electrofishing and fish lift collections for smallmouth bass exhibited seasonal variability in length as a majority of spring and summer collections consisted of larger fish ( $>151 \mathrm{~mm}$ ) while increasing numbers of small fish ( $<150 \mathrm{~mm}$ ) were collected in the fall. Striped bass collected during electrofishing in 1982 were distinctly adult or juvenile in age class as indicated by the sizes collected. The 1982 year class was dominant in subsequent collections in 1983 and 1984. The electrofishing and fish lift collections of walleye were comprised of larger, adult fish.

The length weight relationship as expressed by slope of the regression equation based on data collected in 2010 indicates that lengths and weights of selected species collected at the WFL were similar to those collected 1982 to 1987 (Table 9.1-2). Walleye and channel catfish slopes from 2010 were near the median ranges discovered in the 1980's. Smallmouth bass and largemouth bass slopes from 2010 were slightly lower ( $2 \%$ ) than the lower portion of range discovered in the 1980 's. Though the comparison is limited to one year from the 1980's, yellow perch from 2010 are lower (17\%) than the slope estimated in 1983. Both the 1980's fish and those collected in 2010 were within the ranges presented in Carlander (1969, 1977, 1997).

The relative robustness of a fish can often be described via fish condition ( $\mathrm{K}=$ weight/(length) ${ }^{3}$ ). Fish condition may express the relative nutritional state as ' K ' greater or less than a usual weight at a particular length. Condition factor may also vary with stage of development, maturation and sex in some species. Length weight data were collected for several species at the WFL of the Conowingo Dam in 2010 including channel catfish, redbreast sunfish, bluegill, smallmouth bass, largemouth bass, yellow perch and
walleye. Fish conditions for species collected at the WFL in 2010 were within the normal range of means presented from various populations of the same species in Carlander $\underline{1969}$ and $\underline{1977}$.

Data collected in the 2010 fish stranding summer surveys in the spillway reach below Conowingo Dam supplemented the fish lift catches for a season not typically sampled by the lifts. The most commonly observed fish species were similar on both the east and west sides of the spillway reach. Gizzard shad, banded killifish and sunfish were the three most observed species comprising $57 \%, 23 \%$ and $11 \%$ of the total observations, respectively. Largemouth bass comprised an additional $4 \%$ of the observations. The remaining $5 \%$ of the observations consisted of American eel, carp, minnows, quillback, catfishes, white perch, smallmouth bass, sunfish (Lepomis spp.), walleye, and darters. Blue crab were also observed.

Over the sampling duration 1972 to 2009, the fish lift catch data captured year to year variability and long term trends in fish assemblage of species vulnverable to the lift. Fish lift catches at both the EFL and WFL were robust and provide a baseline indicator of the dominant species in the lower Susquehanna River. Generally, species that were routinely dominant in the fish lift collections (gizzard shad, channel catfish, carp and white perch) were also dominant in 1980's electrofishing, gill net and ichthyoplankton collections. Electrofishing and gill net collections augment the data provided by the fish lift, adding detail to temporal and spatial description of the downstream fish populations. Fish recruitment to these other gear types investigated the various habitats from the Cononwingo tailrace to the tidal zone below Spencer Island throughout all seasons. Icthyoplankton samples provided detail on the reproduction and utilization of the lower Susquehanna habitats by earlier life stages of fish. Data collected in the 2010 fish stranding summer surveys along with recent fish lift catches depict the most current assemblage of species in the tailrace and the lower Susquehanna. Supplemental analyses on condition and length weight relationships describe fish health, benthic macro invertebrate studies and food habits studies explore relationships in predator prey interaction and community ecology. The fish lift, electrofishing, gill net, icthyoplankton studies and supplemental analyses collectively provide a thorough characterization of the fish community and habitat use.

Although several species have increased or declined in abundance, the fish species assemblage has remained diverse below Conowingo Dam with the same core group of species as was observed in the 1980's. A core assemblage consisting of gizzard shad, white perch, common carp, quillback, comely shiner, channel catfish, walleye, smallmouth and largemouth bass along with seasonal migrants like American shad, blueback herring, alewife, sea lamprey and striped bass form the primary group of inhabitants. Condition factor and length weight relationships of representative common fish species
downstream of Conowingo Dam are comparable to those from other normal, natural populations and are indicative of relatively favorable conditions and habitats in the lower Susquehanna River.

## TABLE OF CONTENTS

1.0 INTRODUCTION ..... 1
2.0 BACKGROUND ..... 3
3.0 STUDY AREA ..... 5
4.0 BENTHIC MACRO INVERTEBRATES ..... 6
4.1 STUDY YEAR 1980 ..... 6
4.2 Study Year 1982 ..... 9
4.3 Study Years 1983 and 1984 ..... 12
4.4 Study Year 1984 ..... 13
4.5 Study Year 1988-1989 ..... 14
4.6 Study Year 1989-1990 ..... 14
4.7 Study Year 1990-1991 ..... 15
4.8 Trend Analysis ..... 15
4.9 Summary Discussion and Conclusions ..... 16
5.0 FISH LIFTS ..... 35
5.1 West Fish Lift (WFL) Operations ..... 35
5.2 East Fish Lift (EFL) Operations ..... 36
5.3 Fish Lift Catch Per Unit Effort ..... 38
5.3.1 West Fish Lift ..... 38
5.3.2 East Fish Lift ..... 41
5.4 Discussion ..... 43
6.0 ICHTHYOPLANKTON, GILL NETS, AND ELECTROFISHING ..... 63
6.1 Icthyoplankton Sample Methods ..... 63
6.2 Electrofishing Sample Methods ..... 65
6.3 Gill Net Sample Methods ..... 66
6.4 Icthyoplankton Results ..... 66
6.4.1 Alosa spp ..... 67
6.4.2 American shad ..... 68
6.4.3 Gizzard shad ..... 69
6.4.4 White Perch ..... 69
6.4.5 Other Fishes ..... 70
6.5 Electrofishing Results ..... 70
6.6 Gill Net Results ..... 73
7.0 SUPPLEMENTAL STUDIES OF THE FISH COMMUNITY ..... 104
7.1 Fish Food Habits Below Conowingo Dam ..... 104
7.2 Age and Growth of Smallmouth Bass ..... 105
8.0 LENGTH FREQUENCY ..... 111
8.1 White Perch ..... 111
8.2 Channel Catfish ..... 111
8.3 Gizzard shad ..... 112
8.4 Yellow Perch ..... 113
8.5 Smallmouth Bass ..... 113
8.6 Striped bass ..... 114
8.7 Walleye 1984 ..... 115
9.0 LENGTH WEIGHT RELATIONSHIPS AND FISH CONDITION. ..... 117
9.1 Length Weight Relationships ..... 117
9.2 Fish Condition ..... 118
10.0 FISH STRANDING ..... 121
10.1 Methods ..... 121
10.2 Results ..... 121
10.3 Discussion ..... 122
11.0 CONCLUSION ..... 124
12.0 REFERENCES ..... 126
LIST OF TABLES
TABLE ES-1: SUMMARY OF FISH COMMUNITY AND DOMINANT SPECIES DOCUMENTED TO OCCUR DOWNSTREAM OF CONOWINGO DAM BASED ON HISTORICAL STUDIES, 1972-2010 ..... III
TABLE 4.1-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY ROCK BASKET IN 1980 (JANICKI AND ROSS) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 18
TABLE 4.2-1: BENTHIC MACROINVERTEBRATES COLLECTED BY DRIFT NET IN 1982 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 19
TABLE 4.2-2: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER IN 1982 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 20
TABLE 4.2-3 : COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T- SAMPLER IN 1982 (WESIBERG AND JANICKI) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 22
TABLE 4.3-1: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER SAMPLER 1983 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 24
TABLE 4.3-2: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T- SAMPLER IN 1983-1984 (WEISBERG AND JANICKI) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 26
TABLE 4.4-1: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER SAMPLER IN 1984 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 27TABLE 4.5-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T-SAMPLER 1988-1989 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGODAM. 29
TABLE 4.6-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED 1989-1990 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 30
TABLE 4.7-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T- SAMPLER IN 1990-1991 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 31
TABLE 5.3.1-1: TOTAL ANNUAL CATCH OF FISHES PER YEAR AT THE CONOWINGO DAM WEST FISH LIFT 1972-2009. ..... 44
TABLE 5.3.1-2: TOP TEN TOTALS OF ANNUAL CATCH OF FISHES BY YEAR, CPUE (FISH PER LIFT) AND PROPORTIONAL ABUNDANCE AT THE CONOWINGO DAM EAST AND WEST FISH LIFTS:YEARS 1972-2009. ..... 46
TABLE 5.3.2-1: TOTALS OF ANNUAL CATCH OF FISHES AT THE CONOWINGO DAM, EAST FISH LIFT 1991-2009. ..... 54
TABLE 5.3.2-2: TOP TEN TOTALS OF ANNUAL CATCH OF FISHES BY YEAR, CPUE (FISH PER LIFT) AND PROPORTIONAL ABUNDANCE AT THE CONOWINGO DAM EAST FISH LIFT:YEARS 1991-2009. ..... 55
TABLE 6.4-1: TOTAL NUMBER ICHTHYOPLANKTON COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUEHANNA RIVER 1982 TO 1984 ..... 76
TABLE 6.4-2: SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M ${ }^{3}$ ) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, APRIL THROUGH JUNE 1982. ..... 76
TABLE 6.4-3:SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M ${ }^{3}$ ) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, 30 MARCH THROUGH 28 JUNE 1983. 77
TABLE 6.4-4: SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M ${ }^{3}$ ) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, 3 APRIL THROUGH 28 JUNE 1984.77
TABLE 6.4-5: MONTHLY SUMMARY OF WHITE PERCH (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984. ..... 78
TABLE 6.4-6: MONTHLY SUMMARY OF ALOSA SPP. (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984. ..... 78
TABLE 6.4.3-1: MONTHLY SUMMARY OF GIZZARD SHAD (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984. ..... 79
TABLE 6.5-1:TOTAL NUMBER HOURS OF ELECTROFISHING LOWER SUSQUEHANNA RIVER: ALL LOCATIONS COMBINED 1982-1987. ..... 80
TABLE 6.5-2: MONTHLY TOTAL HOURS OF ELECTROFISHING LOWER SUSQUEHANNA RIVER: ALL LOCATIONS 1982-1987 ..... 80
TABLE 6.5-3: ELECTROFISHING LOWER SUSQUEHANNA RIVER ANNUAL L HOURS OF EFFORT, TOTAL FISH AND CPUE (FISH/HR) 1982-1987. ..... 80
TABLE 6.5-4: ELECTROFISHING LOWER SUSQUEHANNA RIVER OVERALL ANNUAL CPUE (FISH/HR) BY LOCATION 1982-1987 ..... 81
TABLE 6.5-5: ELECTROFISHING LOWER SUSQUEHANNA RIVER ANNUAL TOTAL FISH AND COMBINED OVERALL CPUE (FISH/HR) BY LOCATION 1982-1987 ..... 81
TABLE 6.5-6: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY CPUE BY AREA 1982-1987 ..... 82
TABLE 6.5-7: ELECTROFISHING CONOWINGO TAILRACE LOWER SUSQUEHANNA ANNUAL TOP TEN SPECIES, CPUES (FISH/HR) ..... 84
TABLE 6.5-8: ELECTROFISHING LOWER SUSQUEHANNA MONTHLY PROPORTIONAL ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT THE CONOWINGO TAILRACE.. 85
TABLE 6.5-9: ELECTROFISHING LEES FERRY LOWER SUSQUEHANNA ANNUAL TOP FIVE SPECIES, CPUES (FISH/HR) ..... 86
TABLE 6.5-10: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY RELATIVE ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT LEES FERRY ..... 86
TABLE 6.5-11 ELECTROFISHING THE POOL LOWER SUSQUEHANNA ANNUAL TOP FIVE SPECIES, CPUES ..... 87
TABLE 6.5-12: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY RELATIVE ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT THE POOL ..... 87
TABLE 6.5-13 ELECTROFISHING TIDAL ZONE LOWER SUSQUEHANNA ANNUAL TOP TEN SPECIES, CPUES ..... 88
TABLE 6.5-14: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY PROPORTIONAL ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT THE TIDAL ZONE 89
TABLE 6.6-1: GILL NETS LOWER SUSQUEHANNA RIVER ANNUAL TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984 ..... 90
TABLE 6.6-2: GILL NETS TAILRACE TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984 ..... 92
TABLE 6.6-3: GILL NETS LEES FERRY TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984 ..... 94
TABLE 6.6-4: GILL NETS THE POOL TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984 ..... 95
TABLE 6.6-5: GILL NETS TIDAL ZONE TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984 ..... 96
TABLE 6.6-6: COMPARISON OF THE MONTHLY CPUE (FISH/NET-NIGHT) OF WHITE PERCH COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM. ..... 97
TABLE 6.6-7: COMPARISON OF MONTHLY CPUE (FISH/NET-NIGHT) OF GIZZARD SHAD COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM 1982-1984 ..... 98
TABLE 6.6-8: COMPARISON COMBINED MONTHLY CPUE (FISH/NET-NIGHT) OF CHANNEL CATFISH COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM ..... 99
TABLE 6.6-9: COMPARISON OF THE MONTHLY CPUE(FISH/NET-NIGHT) OF STRIPED BASS COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM. ..... 100
TABLE 7.1-1: INDEX OF RELATIVE IMPORTANCE (IRI) BY SIZE GROUPS FOR STOMACH CONTENTS OF SMALLMOUTH BASS FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM, 1982 ..... 107

TABLE 7.1-2: SUMMARY OF FOOD ITEMS CONSUMED BY 181 YOUNG-OF-THE-YEAR STRIPED BASS FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM, 1982.. 107

TABLE 7.1-3: AGE GROUPS REPRESENTED IN SCALE SAMPLES FROM SMALLMOUTH BASS COLLECTIONS BELOW CONOWINGO DAM, 1980-1983.

TABLE 7.1-4: LENGTH DATA (FL-MM) FOR AGE GROUPS OF SMALLMOUTH BASS COLLECTED BELOW CONOWINGO DAM, 1980-1983. TOTAL FISH = 1,577.

TABLE 7.1-5: COMPARISON OF MEAN TOTAL LENGTH (MM) BY AGE ATTAINED BY SMALLMOUTH BASS AT THE END OF THE GROWING SEASON COLLECTED FROM MARYLAND WATERS AND THE SUSQUEHANNA RIVER DRAINAGE IN PENNSYLVANIA. 110

TABLE 9.1-1: RANGE OF SLOPE OF LENGTH WEIGHT REGRESSION IN 1980'S AND 2010 AND REFERENCE RANGE

TABLE 9.1-2: LENGTH WEIGHT RELATIONSHIPS BY SPECIES AND YEAR 119

TABLE 9.2-1: 2010 CONDITION FACTOR 120

TABLE 10.1-1: SUMMARY OF SUMMER STRANDING STUDIES, JUNE-SEPTEMBER 2010123

## LIST OF FIGURES

FIGURE 4.1: SAMPLING LOCATIONS FOR DRIFT ORGANISMS AND BENTHIC
MACROINVERTEBRATES BELOW CONOWINGO DAM, 1980-1991......................................... 32

FIGURE 4.2-1: COMMUNITY DENSITY AND DENSITY OF COMMON INVERTEBRATE TAXA COLLECTED WITH A SURBER SAMPLER FROM THE CONOWINGO DAM TAILRACE DURING 198-1984.33

FIGURE 4.2-2: COMMUNITY DENSITY AND DENSITY OF COMMON INVERTEBRATE
TAXA COLLECTED WITH A T-SAMPLER FROM THE CONOWINGO DAM TAILRACE
DURING 1982-1991 ..... 34
FIGURE 5.1-1: WEST FISH LIFT SCHEMATIC DRAWING ..... 59
FIGURE 5.1-2: EAST FISH LIFT SCHEMATIC DRAWING ..... 60
FIGURE 5.3.1-1: WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES (FISH PER LIFT) 1972-2009. ..... 61
FIGURE 5.3.2-1: EAST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES (FISH PER LIFT) 1991-2009. ..... 62
FIGURE 6.1-1: GENERAL LOCATIONS FOR ELECTROFISHING AND GILL NET SAMPLING BELOW CONOWINGO DAM, 1980-1991 ..... 101
FIGURE 6.1-2: LOCATIONS FOR ELECTROFISHING, GILL NET AND ICTHYOPLANKTON SAMPLING BELOW CONOWINGO DAM, 1980-1991 ..... 102
FIGURE 6.4-1: MONTHLY TOTAL ICHTHYOPLANKTON COLLECTED 1983 AND 1984 FOR ALOSA SPP., WHITE PERCH AND GIZZARD SHAD. ..... 103

## LIST OF APPENDICES

```
APPENDIX A - WEST FISH LIFT DATA
APPENDIX B-EAST FISH LIFT DATA
APPENDIX C-ICHTHYOPLANKTON
APPENDIX D-ELECTROFISHING
APPENDIX E-GILL NETS
APPENDIX F-AGE AND GROWTH PLOTS FROM RMC 1985A, B, C. AND LENGTH
FREQUENCY CHARTS FROM THE WEST FISH LIFT }201
```


## LIST OF ABBREVIATIONS

Agencies

| EFL | East Fish Lift |
| :--- | :--- |
| FERC | Federal Energy Regulatory Commission |
| ISR | Initial Study Report |
| MDE | Maryland Department of Environment |
| MDNR | Maryland Department of Natural Resources |
| MGS | Maryland Geological Survey |
| NMFS | National Marine Fisheries Service |
| NOAA | National Oceanic and Atmospheric Administration |
| NPS | National Park Service |
| PADEP | Pennsylvania Department of Environmental Protection |
| PADCNR | Pennsylvania Department of Conservation and Natural Resources |
| PEPCo | PECO Energy Power Company |
| PFBC | Pennsylvania Fish and Boat Commission |
| PGC | Pennsylvania Game Commission |
| PGS | Pennsylvania Geological Survey |
| SPCo | Susquehanna Power Company |
| SRBC | Susquehanna River Basin Commission |
| USACE | United States Army Corps of Engineers |
| USEPA | United States Environmental Protection Agency |
| USGS | United States Geological Survey |
| USFWS | United States Fish and Wildlife Service |
| USR | Updated Study Report |
| WFL | West Fish Lift |

## Units of Measure

C Celsius, Centigrade
cfs cubic feet per second
cm centimeter
El. elevation
F Fahrenheit
ft feet
fps feet per second
h hour
L liter
m meter
$\mathrm{mm} \quad$ millimeter
MW megawatt

## Miscellaneous

| DC | Direct Current |
| :--- | :--- |
| EA | Environmental Assessment |
| EIS | Environmental Impact Statement |
| FL | Fork Length |
| FPA | Federal Power Act |

### 1.0 INTRODUCTION

Exelon Generation Company, LLC (Exelon) has initiated with the Federal Energy Regulatory Commission (FERC) the process of relicensing the 573-megawatt (MW) Conowingo Hydroelectric Project (Project). Exelon is applying for a new license using the FERC's Integrated Licensing Process (ILP). The current license for the Conowingo Project was issued on August 14, 1980 and expires on September 1, 2014.

Exelon filed its Pre-Application Document (PAD) and Notice of Intent (NOI) with FERC on March 12, 2009. On June 11 and 12, 2009, a site visit and two scoping meetings were held at the Project for resource agencies and interested members of the public. Following these meetings, formal study requests were filed with FERC by several resource agencies. Many of these study requests were included in Exelon's Proposed Study Plan (PSP), which was filed on August 24, 2009. On September 22 and 23, 2009, Exelon held a meeting with resource agencies and interested members of the public to discuss the PSP.

Formal comments on the PSP were filed with FERC on November 22, 2009 by Commission staff and several resource agencies. Exelon filed a Revised Study Plan (RSP) for the Project on December 22, 2009. FERC issued the final study plan determination for the Project on February 4, 2010, approving the RSP with certain modifications.

The final study plan determination required Exelon to conduct a literature-based study to provide a characterization of the current aquatic community below Conowingo Dam. The report should include data beyond that provided by annual fish lift catches, including information on fish size, age, condition factor, and abundance indexed by CPUE. The objective of this report is to utilize existing data to: 1) characterize resident fish abundance, size structure, condition, and reproductive success below Conowingo Dam --the existing data includes fish lift catches as well as results from other common fisheries gear types such as electrofishing, gill nets, and ichthyoplankton nets; 2) describe the benthic macroinvertebrate communities below Conowingo Dam collected by various common collection gears; and 3) provide updated information on these communities available through 2010 studies focused on other objectives. The freshwater mussel community, part of the overall invertebrate community, was examined separately and results are reported in Conowingo Project Study 3.19-Freshwater Mussel Characterization Study below Conowingo Dam.

An initial study report (ISR) was filed on February 22, 2011, containing Exelon's 2010 study findings. An initial study report meeting was held on March 9, 10 and 11, 2011 with resource agencies and interested members of the public. Formal comments on the ISR including requested study plan
modifications were filed with FERC on April 27, 2011 by Commission Staff, several resource agencies and interested members of the public. Exelon filed responses to the ISR comments with FERC on May 27, 2011. On June 24, 2011, FERC issued a study plan modification determination order. The order specified what, if any, modifications to the ISRs should be made. For this study, FERC's June 24, 2011 order required no modifications to the original study plan. An updated study report (USR) was filed on January 23, 2012 addressing comments from stakeholders received at the March ISR meeting, those comments addressed by Exelon in the May 27, 2011 responses to ISR comments, as well as editorial and minor text changes. This final study report is being filed with the Final License Application for the Project.

### 2.0 BACKGROUND

The Susquehanna River below Conowingo Dam flows approximately 10 miles before entering Chesapeake Bay. The non-tidal portion of the Susquehanna River encompasses approximately 3.5 miles of river length from Conowingo Dam downstream to the mouth of Deer Creek (a tributary), which is the approximate natural upstream limit of tidal influence. The Chesapeake Bay stretches about 200 miles from the Susquehanna River in the north to the Atlantic Ocean in the south. A broad, shallow area called the Susquehanna Flats, south of Havre de Grace, Maryland, represents the point where the Susquehanna River flows into upper Chesapeake Bay.

The Conowingo Project uses limited active storage within Conowingo Pond for generation purposes. Maximum hydraulic capacity of the Conowingo powerhouse is $86,000 \mathrm{cfs}$. The current minimum flow regime below Conowingo Dam was formally established with the signing of a settlement agreement in 1989 between the Project owners and several federal and state resource agencies (46 FERC 961,063) (FERC 1989). The established minimum flow regime below Conowingo Dam is the following:

| March 1 - March 31 | $3,500 \mathrm{cfs}$ or natural river flow |
| :---: | :---: |
| April 1 - April 30 | $10,000 \mathrm{cfs}$ or natural river flow, whichever is less |
| May 1 - May 31 | 7,500 cfs or natural river flow, whichever is less |
| June 1 - September 14 | $5,000 \mathrm{cfs}$ or natural river flow, whichever is less |
| September 15 - November 30 | 3,500 cfs or natural river flow, whichever is less |
| December 1 - February 28 | 3,500 cfs intermittent (maximum six hours off followed by equal amount on) |

The downstream discharge must equal these values or the discharge measured at the Susquehanna River at the Marietta, Pennsylvania United States Geological Survey (USGS) gage (No. 01576000), whichever is less. The Marietta USGS gage is located approximately 35 miles upstream of Conowingo Dam above the Safe Harbor Dam.

The Lower Susquehanna River below the Conowingo Dam contains a variety of freshwater habitats that support communities of resident and migratory fishes. The river in this reach consists of lotic, riverine habitat conditions extending 9 miles to the upper Chesapeake Bay. The upper 3.5 miles are non-tidal and the lower 6.5 miles to Havre de Grace, Maryland exhibit daily tidal cycles.

Thirty-eight years of annual collections at the West Fish Lift (WFL) and 19 years at the East Fish Lift (EFL) facilitate description of year to year and long term fluctuations in catch and proportional abundance in key resident and migratory species of the Lower Susquehanna River fish assemblage. In addition to the fish lift data, fish distribution and abundance surveys from previous relicensing studies were conducted by Exelon from 1982 to 1987. These data augment the fish lift collections in providing a more detailed spatially and temporally diverse characterization of the downstream fish populations in regards to species assemblage, condition, and habitat use. These studies included electrofishing, gillnet and ichthyoplankton sampling efforts from Conowingo Dam to the tidal zone just below Spencer Island.

Benthic macroinvertebrate communities were also examined as part of the these studies. In parallel and extending through the 1980's into the 1990's, benthic community studies were performed by Maryland Power Plant Siting (now Power Plant Research Program, PPRP) towards eventual settlement in 1989 of the minimum flow issue. Resolution of the minimum flow issue yielded the seasonal flow release schedule described above.

### 3.0 STUDY AREA

The lower river is the 10 mile stretch of the Susquehanna River from Conowingo Dam to the Chesapeake Bay. Within this 10 miles the character of the river varies considerably. The tailrace extends from Conowingo Dam to below Rowland Island. The maximum depth at zero generation is 23 ft . The main flow of water occupies the deep channel between the west bank and Rowland Island. In the riverine reach below the tailrace, the water flow spreads out toward the east shore. This is a typical shallow bedrock river with numerous boulders and bedrock outcroppings. At the head of this reach is Lee's Ferry. This is a shallow, rocky pool with depths of 6.5 ft at zero generation. From Lee's Ferry water flows through shallow riffles, runs, and pools to a larger area called The Pool, which has a depth of 10 ft at zero generation. The non-tidal portion of the lower river is composed of the tailrace and the riverine reach. The lower limit to the riverine reach is the de facto tidal limit at Deer Creek. The tidal portion of the lower river extends from just upstream of Deer Creek to the mouth of the river at the head of the Susquehanna Flats. The upper end of the tidal area is rocky and shallow similar to the riverine reach. Below Spencer Island the river deepens and broadens. Maximum depth is approximately 39 ft . The tidal amplitude in this reach is normally 2 ft .

### 4.0 BENTHIC MACRO INVERTEBRATES

Several studies provide the data that will be used here to characterize the benthic macroinvertebrate community below Conowingo Dam. These include:

- Janicki and Ross (1982) (study year 1980)
- Janicki and Weisberg (1983) (study year 1982)
- Weisberg and Janicki (1985) (study years 1982-1984)
- RMC (1985 a; $\underline{b} ; \underline{c})$ (study years 1982-1984)
- Scott. (1991) (study years 1988-1991)

Several hundred samples and tens of thousands of specimens were quantitatively collected with multiple gear types during the referenced studies. Samples were taken from habitats within the non-tidal reach below Conowingo Dam, extending downstream about 3.5 miles (Figure 4.1). The individual studies above were performed during the initial decade of the current license period. A number of comprehensive benthic macroinvertebrate studies were completed between 1980 and 1991 to address the minimum flow issue, which contributed information towards a Settlement Agreement between the Maryland Department of Natural Resources (MDNR) and the Philadelphia Electric Company (now Exelon). The Settlement Agreement resulted in the flow release schedule identified above (see Section 2.0, Background).

Descriptions of the resident benthic communities are made through an evaluation of the most abundant taxa collected. This report provides an ecological characterization of the invertebrate communities as they existed during the study years: 1980, 1982, 1983, 1984, 1988, 1989, and 1991. To assemble it a literature search was conducted and augmented by communications with MDNR and the Susquehanna River Basin Commission (SRBC).

### 4.1 Study Year 1980

Samples were collected in 1980 by Janicki and Ross (1982) with rock basket samplers. Each sampler consisted of a wire basket 16 cm in diameter, 25 cm long, and filled with stones 5 to 20 cm on the longest axis. Ten baskets were placed in series along each of four transects (Figure 4.1) and secured with cable. Following a colonization period of three weeks they were removed and replaced with new ones. Five of the ten baskets from each transect were randomly selected for processing. The collection effort spanned a
time period of June through October and included three habitat types: channel, isolated pool, and exposed substrate.

Specimen identifications in the original study were made to genus. A Tolerance Index, ranging from 0 to 10 (where low numbers indicate pollution sensitivity) was applied to each taxon for the present characterization. The Indices were taken from lists published by MDNR (2005) and PADEP (2009). Tolerance Indices from 0 to 3 are considered to represent sensitive forms, whereas values between 4 and 6 are considered facultative and values from 7 to 10 represent tolerant taxa. Taxa listed in the tables in bold represent $10 \%$ or more of the cumulative total.

Mean number of specimems per basket ranged between 311 and 911 for channel and isolated pool habitats. The numbers of specimens taken from exposed habitat were lower by one or two orders of magnitude, so most of the data summarized in Table 4.1-1 are from wetted portions of the river bottom. Omitting rare and uncommon taxa, the invertebrates in Table 4.1-1 combined represent $90.1 \%$ of all specimens collected. The most abundant taxa were Gammarus (32.2\%), Dugesia (25.1\%), Chironomidae $(16.1 \%)$, and Nais (5.6\%). All of these are tolerant forms found in warm-water habitats.

Gammarus is a relatively mobile crustacean found in both lotic (flowing) and lentic (still) habitats. They tend to prefer shallow areas (less than one meter) of slow to moderate current and do best in well oxygenated water. The most common species in the Susquehanna River is G. fasciatus, a generalist found in a variety of rivers, streams, lakes, and ponds at densities that can exceed 10,000 per square meter. They are more or less confined to the upper surface of the substrate (are epi-benthic) but have the capacity to swim in a haphazard fashion that is the basis for the common names given to them: scuds or side-swimmers. Their mobility allows them to adapt to fluctuating water levels characteristic of the tailwaters below Conowingo Dam. Omnivorous, they are noted in Pennak (1989) as voracious feeders that tend to be most active at night. They readily consume a wide variety of plant and animal matter and are also known to scavenge. Gammarus are a fairly large invertebrate that attains an adult size approaching 10 mm making them an energy laden food source for resident fish. Stomach content analysis of eight fish species conducted by Weisberg and Janicki (1985) during 1982 and 1983 showed Gammarus to be a major component (expressed as number and/or biomass) in the diets of channel catfish, white catfish, brown bullhead, yellow perch, white perch, bluegill, and pumpkinseed. They were also common but somewhat less numerous in the stomachs of green sunfish.

The primitive flatworm Dugesia ranked second in abundance. By far the most common species in warm rivers is $D$. tigrina, a small worm usually less than five mm long and capable of only limited locomotion
on the surface of submerged substrates. This, in combination with a preference for still shallow water, would make them susceptible to water level fluctuation. Respiration is accomplished by gas exchange through the epidermis so they are able to persist in low oxygen environments. They have the capacity to reproduce either asexually or sexually but usually do so via fission, an adaptation that allows them to avoid loss of progeny due to egg destruction. To facilitate locomotion Dugesia secrete a distasteful slime through their body wall, so they are seldom a major component in the diet of fishes. They were not found in the stomach contents of any of the eight fish species noted above.

Chironomidae (midges) are a complex and species-rich family of two-winged (true) flies that are found in all types of fresh water. Because of the numerous ecological roles they play as both consumers and prey and the degree of resource partitioning they represent, chironomids enhance the stability of aquatic systems. Collections typically included 10 to 20 genera. Most midges are either facultative or pollution tolerant. The family is so versatile that it is difficult to provide a description applicable to the majority of genera commonly identified from large data sets. Some are epi-benthic, living in cases, while many others are adept burrowers. Burrowing forms would be favored in the non-tidal reach because they would be able to avoid desiccation (or freezing in the winter) on periodically exposed substrates by migrating downward into the still-moist subsurface. The genera listed in Table 4.1-1 fall within the subfamilies Chironominae and Orthocladiinae that include many burrowers. Midges are small, thin-bodied insects and usually less than 5 mm in length. On an individual basis they do not provide much energy when fed upon by large predators. However, they are normally so numerous that when taken together they form an important component of the diet of almost all of the omnivorous and carnivorous fish found in the nontidal reach. Midges undergo what is termed a complete metamorphosis, meaning a pupal stage is included. During metamorphosis, pupae migrate en mass from the stratum through the water column to the surface where they molt and emerge as terrestrial adults. It is at this (exposed) stage in their life history that they become particularly susceptible to predation. Numerically, chironomids were the most frequently encountered forage item in the stomach contents of all eight fish species investigated.

The most commonly identified chironomid was Dicrotendipes, one of the most tolerant genera within the family. They are included in a group referred to as the blood-red midges. Their color is produced by a hemoglobin-like compound called erythrocruorin (Pennak 1978) that acts as respiratory pigment, allowing them a selective advantage in low-oxygen environments. They have also been observed to be disproportionately thermal-tolerant in heated discharges.

Also common was a very small segmented worm in the genus Nais, a widely distributed genus often found associated with filamentous algae. Nais feed upon single celled algae and bacteria that form a film
on the surface of the filaments. Some species prefer soft sediments. Because only limited amounts of either habitat type are found in the tailrace, the presence of Nais may be due in part to passage downstream from Conowingo Pond. They reproduce asexually via budding. Respiration is through the body membrane. Segmented worms (identified to the Class Oligochaeta) were present in the stomach contents of all eight fish species investigated by Weisberg and Janicki (1985) but were not numerous.

### 4.2 Study Year 1982

During 1982 data were collected by RMC (RMC 1985a) and Weisberg and Janicki (1985) using three quantitative gear types: a drift sampler, a Surber sampler, and a T-Sampler.

Drift is a dispersal mechanism through which invertebrates, both actively or passively, enter the water column and are transported by the current downstream. When this occurs they become available as prey for resident fishes. RMC collected samples at up to seven locations from July through September using a $0.5-\mathrm{m}$ diameter drift net with $1-\mathrm{mm}$ mesh (Figure 4-1). The net was deployed from a boat and suspended in or towed through the current for a period of ten minutes. Following retrieval, specimens were washed into a cod-end of the net and preserved. The volume of water passing through the net was measured with a flow meter mounted to the inside. A cumulative total of $258 \mathrm{~m}^{3}$ of water was passed through the net during the course of study.

RMC also collected benthic samples at seven locations with a Surber sampler. The Surber consists of a metal frame that encloses a $0.092 \mathrm{~m}^{2}$ area of substrate. Samples were collected by placing the Surber on cobble/gravel substrate in an area of current. The enclosed substrate within the frame was then disturbed by hand, allowing the river flow to transport specimens into a capture net attached to the downstream end of the device. The Surber is designed to collect samples from areas where water depth is one foot or less. Either one or two samples were collected at each station over a four-month period in the summer and autumn and the specimens identified to genus, except for the midges (left at Chironomidae) and segmented worms (left at Oligochaeta).

Weisberg and Janicki used a T-Sampler to collect specimens from gravel and bedrock substrates arrayed along a transect (Transect D) established downstream of the Octoraro Creek confluence (Figure 4-1). The T-Sampler is a suction device deployed by divers to collect benthic samples at greater depth than what is possible with the Surber. It is constructed of plastic drain pipes assembled in the shape of a "T" with a $15-\mathrm{cm}$ diameter neoprene flange attached to one end to allow a good seal with the substrate. The sampler, when deployed, encloses a $0.025 \mathrm{~m}^{2}$ area of substrate. Samples are collected by inserting a hand onto an opening leading to the flange and disturbing the substrate inside the enclosure in a similar fashion to the

Surber. A submersible pump was attached to one end of the device to force dislodged specimens into a sampling receptacle. Five replicates were collected at each location at approximately two-week intervals from July through December, yielding a total of 120 samples. Collections included both intermittently exposed and permanently wetted portions of the river channel.

Drift net densities varied from month to month between 91 individuals per $\mathrm{m}^{3}$ in July to 10 per $\mathrm{m}^{3}$ in September. The collection effort produced a total of 20 taxa (all stations/dates combined). However, only four taxa were numerous enough to represent at least $5 \%$ of the total during at least one of the months sampled (Table 4.2-1) including: Leptodora (70.6\%), Chironimidae (15.8\%), Hydroptila (2.8\%), and Cheumatopsyphe (1.6\%).

Leptodora kindti is a limnetic (found in the water column) crustacean in the order Cladocera that was particularly abundant ( $76 / \mathrm{m}^{3}: 82.6 \%$ of the total) in the July collections. A zooplankter adapted to lakes and ponds, they represent an out-migrant from Conowingo Pond, which serves as a transported food item. They are large for a planktonic organism, attaining a length of nearly $1-\mathrm{mm}$ when fully mature, predatory, and attract to aquatic vegetation. Stomach content analysis showed that they were preyed upon in large numbers by white perch.

The phantom midge (Chaoborus punctipennis) was another drift organism not as abundant and was found in the diet of white perch and bluegill. Chaoborus are a limnetic form present in large numbers in Conowingo Pond. Predatory, they are mobile swimmers that migrate vertically into the water column after dark, becoming subject to passage downstream during generation. Their presence along with $L$. kindti illustrates the contribution that organisms in flow releases from Conowingo Pond make to the forage base available to the fish found below the dam.

The Surber sampler study produced a total of 25 taxa in 1982 at a mean density of $2,065 \mathrm{per}^{2}$ (Table 4.2-2), indicating a fairly sparse community. Productive ecosystems typically produce invertebrate densities exceeding 10,000 per $\mathrm{m}^{2}$ (Hynes, 1970). Most abundant were Chironomidae (46.8\%), Cheumatopsyche (19.7\%), Gammarus (10.7\%), Corbicula (7.0\%), Hydroptila (4.4\%), Oligochaeta ( $2.7 \%$ ), and Goniobasis $(2.5 \%)$. Cheumatopsyche belong to a group referred to as the net-spinning caddisflies and attain a length of up to 15 mm as mature larvae. Often very numerous, Cheumatopsyche are an epi-benthic insect that construct protective retreats (casings) attached to the top and sides of medium to large stones. They also build silken capture-nets from which they passively filter food items from the water column. Cheumatopsyche are omnivores. They are relatively immobile, adapted (obligated) to lotic habitats and, as such, are susceptible to water level fluctuation, especially on bedrock
substrates. Respiration is both through the body surface and through external gills so they do best in welloxygenated environments. Cheumatopsyche mature through complete metamorphosis and frequently emerge synchronously when they become most available to predatory fishes. Both the larvae and pupae were a major component in the diet of the fish species examined, except for brown bullhead.

Hydroptila is the most common genus within a family referred to as the micro-caddisflies. They are very small, usually with a body length of 2 or 3 mm , preferring to live in backwaters and areas of relatively slow current. Given their limited size and adaptation to slow moving currents, they are likely susceptible to sudden rises in water level. Unless very numerous, they do not provide a large amount of energy to foraging fishes. As is true of most caddisflies, Hydroptila are case-makers, although the cases are not fixed as in Cheumatopsyche and they do not build them until their fifth (final) larval instar (stage). Thus, they are free-living throughout roughly half of their larval development. Hydroptila feed on filamentous algae and diatoms and prosper in well oxygenated habitats. They were present in low numbers in the stomach contents of six of the eight fish species examined, brown bullhead and channel catfish excepted.

Corbicula fluminea is a medium-sized clam native to Asia that, since its introduction in the Midwest, has become one of the most abundant mollusks found in warm-water environments. They attain a maximum shell diameter of 35 mm . They feed through a pair of siphons extended into the water column to "pump" unicellular algae and bacteria into their gills from where it is filtered and transported to their digestive systems. Habitual burrowers, Corbicula can easily retain moisture on de-watered substrates by closing their shells until flow is restored. They prefer gravel/cobble substrates to which they can attach themselves. Corbicula are most available as forage to bottom-feeding species but were not found in the stomachs of any of the catfishes inspected.

Goniobasis virginica is a relatively large, gilled snail common in rivers. Shell length in mature specimens reaches 20 mm and width at the aperture (opening) is about 5 mm . As is typical of most snails, Goniobasis move about by secreting a thin ribbon of mucus referred to as a "slime track" that acts a lubricant allowing locomotion. Slow movers, they can nonetheless cover a considerable distance given sufficient time. Gilled snails generally require more oxygen than their pulmonate (air breathing) counterparts but Goniobasis are relatively tolerant, able to survive low-oxygen environments for periods of time. They prefer slow-moving, shallow areas but are adept at attachment and can avoid displacement due to water level fluctuations. Gilled snails all possess a chitinous operculum dorsally on the body, that when retracted into the shell forms a seal protecting the individual from drying. Goniobasis are mostly preyed upon by bottom feeders. They were found in small numbers in the stomach contents of channel catfish and brown bullhead.

Oligochaete worms represented $8.4 \%$ of the sample totals from the November collection. Two subfamilies were represented: Naidinae and Tubificinae. Naid genera identified from collections included Dero, Nais, Pristina and Stylaria. Tubificid genera were Aulodrilus, Bothrioneureum, and Branchiura. The three additional Naid genera essentially occupy the same niche as Nais. The tubificines are burrowers that perform a similar ecological role as that of the more familiar terrestrial earthworms; aerating, mixing, and breaking down the sediment. Most (Branchiura excepted) are very small, less than 10 mm in length, thin-bodied, and all are very tolerant organisms. Their preference for soft (unstable) sediments makes them susceptible to water level fluctuation. Oligochaetes were found in small numbers in the stomachs of all species investigated except for green sunfish.

A total of 31 genera was collected with the T-Sampler at a noticeably higher density ( 12,270 per $\mathrm{m}^{2}$ ) than the Surber Sampler (Table 4.2-3). This higher density suggests that community density increased in deeper (more permanently wetted) areas. With one exception, the dominance hierarchy was very similar to that observed with either the Surber or rock basket samplers including: Chironomidae (30.0\%), Manayunkia (22.3\%), Dugesia (12.2\%), Cheumatopsyche (8.7\%), Oligochaeta (6.3\%), Gammarus (5.8\%), and Corbicula (5.7\%).

Manayunkia is part of the class Polychaeta, known as marine or bristle worms. Although Polychaera are found almost exclusively in saline bay and ocean environments, a limited number of genera have been able to colonize oligohaline and tidal freshwater areas. M. speciosa is a polychaete that is fully adapted to fresh non-tidal rivers. Manayunkia is a very small (less than 5 mm ), sedentary, tube-building organism that feeds with a pair of tentacular fan-like gills (lophophores). They live in soft (unstable) sediments and feed by extending their gills from the tubes to filter particulate matter from the water column. Manayunkia is seldom found at dissolved oxygen concentrations lower than $5 \mathrm{mg} / \mathrm{L}$, so the Tolerance Index of 10 given them may be erroneous, and more a function of the species ability to colonize freshwater environments. Small but very numerous, they have the potential to form a fairly substantial food source for resident fishes. However, Manayunkia were only found in small numbers in the stomachs of one species, white perch.

### 4.3 Study Years 1983 and 1984

Data sets were condensed from studies completed by RMC (1985b) and Weisberg and Janicki (1985). Gear types were the Surber sampler and the T-Sampler, each used to collect specimens from the same locations/transects as during 1982 (Figure 4-1). A total of 53 samples was obtained from quarterly collections during 1983 with the Surber, augmented with 160 samples collected over two-week intervals with the T-Sampler between July 1983 and February 1984.

Forty-one genera were collected by RMC by Surber sampler at a mean density of 2,303 per $\mathrm{m}^{2}$ (Table 4.3-1), Dominant taxon included Chironomidae (38.5\%), Oligochaeta (23.5\%), Cheumatopsyche (11.8\%), Gammarus (7.0\%), Corbicula (4.3\%), Hydroptila (3.4\%), and Goniobasis (2.6\%)

Except for Oligochaeta, the abundant taxa identified from the 1983 collections were taken in numbers similar to those observed during 1982. Densities exceeding 1,000 per $\mathrm{m}^{2}$ from at least one of the sample periods were recorded for Oligochaeta, Cheumatopsyche, and Chironomidae.

Analysis of the 1983-84 samples collected with the T-Sampler, illustrated in Table 4.3-2, produced 48 genera. Mean density was 13,635 individuals per $\mathrm{m}^{2}$, similar to the 1982 data. Most frequently collected taxon were Chironomidae (29.4\%), Gammarus (17.3\%), Manayunkia (14.4\%), Dugesia (5.6\%), Cheumatopsyche (5.2\%), Sphaeriidae (4.9\%), and Ferrissia (4.9\%).

Three additional taxa were abundant (relative to 1982): sphaeriid clams (genera Pisidium, or Musculium) and Ferrissia rivularis, a small snail.

The clams essentially occupy the same niche as do the Corbicula noted earlier although they are considerably smaller as adults, seldom exceeding a diameter of 10 mm . Sphaeriids were found in small numbers within the stomachs of white catfish, white perch, yellow perch, and pumpkinseed.

Ferrissia rivularis is a small, pulmonate (non-gilled) snail, referred to as a limpet, which feeds by scraping algae from rocky substrates in the shallows of running waters. They are one of a fairly small number of snail species found more commonly in areas of discernable current rather than in quiescent pool and backwater areas. F. rivularis are almost always less than 5 mm in length. The pulmonary cavity is vestigial so they breathe through the body surface. Due to the lack of a functioning pulmonary cavity, oxygen requirements for $F$. rivularis are greater than that of other pulmonates. Ferrissia rivularis were found in small numbers in the stomachs of channel catfish, yellow perch, pumpkinseed, and bluegill.

### 4.4 Study Year 1984

Data reported from 42 samples collected during the summer and fall of 1984 with a Surber sampler (RMC 1985 c ) are provided in Table 4.4-1. This effort produced 40 genera and a density of $2,017 \mathrm{per}^{2}$. Dominant taxon included: Chironomidae (31.3\%), Cheumatopsyche (27.2\%), Oligochaeta (12.9\%), Corbicula (7.3\%), Gammarus (6.2\%), Goniobasis (2.6\%), and Hydra (2.4\%).

Hydra is a very small, sessile form attached to hard substrates in either flowing or quiet waters at depths of 1.5 ft or less. They feed by immobilizing prey (e.g., zooplankton) that comes into contact with toxic
nematocysts incorporated within their tentacles. Reproduction is through budding. Hydras were not widely fed upon.

### 4.5 Study Year 1988 - 1989

These data (Table 4.5-1) are condensed from Scott (1991), who collected with a T-Sampler from Transect D below Octoraro Creek using the same methodology employed during the 1982 - 1984 studies by Weisberg and Janicki (1985). During September 1988 through March 1989, 144 samples were obtained from gravel and bedrock substrates at two-week intervals. Community Density averaged 11,925 per m², similar to earlier years. Dominant taxa were Corbicula (28\%), Polypedilum (16\%), Cheumatopsyche (9\%) Manayunkia (8\%), Oligochaeta (8\%) Dugesia (5\%) and Cricotopus (5\%).

These results indicated that taxonomic composition among those genera that represented the majority of the specimens collected had changed little during the time interval since the earlier collections in 1980 1984. The invertebrate community continued to be dominated by facultative or tolerant warm-water taxa. Two midge genera were identified, that were not to be especially numerous from the earlier datasets were Polypedilum (within the subfamily Chironominae) and Cricotopus (Orthocladiinae). Both genera likely provide ample forage to resident fish species.

Polypedilum occur in both still and flowing waters and on a variety of substrates. Many species are able to avoid exposure by burrowing, while others are case-makers found on hard substrates. During the Tsampler collections they were found in considerably greater numbers on bedrock substrates. Thus, Polypedilum appear well adapted to the swift currents produced during generation periods but may not be able to withstandstand desiccation when bedrock is exposed.

Tolerance indices for the common Cricotopus species range between 6 and 10. Most are 7 or higher, making Cricotopus one of the more tolerant genera within the family. Cricotopus tend to be more riverine in their habitat preference than are many other large chironomid genera, meaning they are more often found in swifter currents and coarser substrates. They are adept burrowers.

### 4.6 Study Year 1989 - 1990

Data reported by Scott (1991) were gathered from a total of 118 T-sampler collections from September 1989 to March 1990 (Table 4.6-1). Community Density increased relative to the earlier study years. An average of 20,225 per $\mathrm{m}^{2}$ was calculated. Commonly collected taxa were Manayunkia (30\%), Polypedilum (19\%), Cheumatopsyche (12\%), Corbicula (8\%), Oligochaeta (6\%) and Dugesia (5\%).

Taxomomic composition observed from the 1989 to 1990 data were similar to those collected during previous years, except for a higher percent abundance of Manayunkia speciosa.

### 4.7 Study Year 1990 - 1991

The data summarized below from Table 4.7-1 were assembled from 107 T-sampler collections between September 1990 and March 1991 (Scott, 1991). Community Density was again relatively high, 18,500 per $\mathrm{m}^{2}$. Dominant taxa included Corbicula (23\%), Chironomidae (19\%), Manayunkia (18\%), Cheumatopsyche (10\%), Oligochaeta (6\%), and Dugesia (5\%).

The dominance hierarchy from 1990 - 1991 was similar to those observed in earlier years. Thus, we conclude that the character of the benthic macroinvertebrate community in the Susquehanna River below Conowingo Dam changed little during the ten-year time frame encompassed by this review.

### 4.8 Trend Analysis

Figures 4.2-1 and 4.2-2 are provided to show temporal trends in community density for two gear types, the Surber and the T-sampler, respectively. To compile them, standing crop data were taken from RMC (1985) for the Surber samples collected during 1982 to 1984 . T-Sampler densities were taken from Weisberg and Janicki (1985) for the years 1982 to 1984. Others were derived (estimated to within 25 specimens per $\mathrm{m}^{2}$ ) from figures given in $\operatorname{Scott}(\underline{1991)}$ ) for 1988 to 1991. Also shown are population densities for seven common taxa (those abundant during most years), including Cheumatopsyche, Chironomidae, and Gammarus (those commonly identified from the stomach content analyses).

There are some inherent problems with this calculation because sampling from year to year occurred: during different seasons; from different habitats (cobble or bedrock); from exposed versus wetted substrates; from different locations; and during a number of flow regimes. However, with this in-mind some attention can be given to the results, including those from the latter years of study (after 1988) once the current flow regime was essentially in-place. Note that the density values on the figures are a cumulative mean of all of the samples collected each year.

As noted, the Surber data from 1982 to 1984 produced community density estimates that varied within a narrow range of 2,017 and 2,303 specimens per $\mathrm{m}^{2}$.

In comparison, the T-Sampler data from this same (earlier) time-frame produced community densities that were also consistent from year to year, ranging between 12,270 and $13,635 \mathrm{per}^{2}$. Higher densities from the T-Samples may have been due to placement of the device in deeper water (more permantly wetted locations). However, the results from both gear types were consistant in that they show little change in
the respective densities during this earlier three-year period. Community density from the 1988-89 year was 11,925 per $\mathrm{m}^{2}$.

Higher densities were observed from the T-Sampler data during 1989 to 1991. The result from the 198990 study year was $20,225 \mathrm{per} \mathrm{m}^{2}$, followed by $18,500 \mathrm{per} \mathrm{m}^{2}$ in 1990-91. This represents a near fifty percent increase in community density from the years 1982, 1983-84, and 1988-89. This increase may be cautiously interpreted as a positive response to the new flow regime. Individual taxa that displayed new maxima during one or both latter years were: Cheumatopsyche, Chironomidae, Corbicula, and Manayunkia. The greatest increase was observed for Manayunkia that attained a population density of 6,125 per m $^{2}$ in 1989-90.

### 4.9 Summary Discussion and Conclusions

During 1980 through 1991 a series of quantitative benthic studies was conducted in the Susquehanna River in the tailrace and non-tidal waters below Conowingo Dam to determine a release schedule sufficient to maintain healthy fish and macroinvertebrate communities. These data provide a descriptive characterization of the invertebrate community during the study years 1980, 1982, 1983, 1984, 1988, 1989, and 1991. The study area encompassed a 3.5 -mile non-tidal section of river channel bounded by two tributaries: Octoraro Creek (upstream near the tailrace) and Deer Creek (downstream near where the Susquehanna River becomes tidal). Gear types used were rock basket, drift net, Surber, and T-Samplers. Gravel and bedrock substrates were sampled, including both submerged and exposed areas.

A cumulative (all years combined) total of 71 taxa was collected and identified to the genus endpoint. The 1988 through 1991 study years produced a total of 115 invertebrate taxa identified to the genus/species level (Scott, 1991). Note that the identifications shown on Tables 4.5-1, 4.6-1 and 4.7-1 are to genus.

Community density estimates from the Surber collections were near 2,000 individuals per $\mathrm{m}^{2}$. Density estimates from the T-Samples (taken in deeper water) were higher, near 13,000 individuals per $\mathrm{m}^{2}$. These results were used as a basis to characterize the community as moderately dense. During the final two years of study density increased to near 18,000 and 20,000 per $\mathrm{m}^{2}$.

The community was generally comprised of facultative or tolerant warm-water genera. Most abundant were: Chironomidae (Cricotopus, Dicrotendipes, and Polypedilum), Cheumatopsyche, Corbicula, Dugesia, Gammarus, Goniobasis, Hydroptila, Manayunkia, and Oligochaeta (Nais). The most important food items in the stomach contents of eight fish species examined were Chironomidae, Cheumatopsyche, and Gammarus.

Most of the genera identified from the studies possess some adaptation to water level fluctuation and low dissolved oxygen concentrations. Review of the tolerance indices listed on the report tables shows only 8 of 71 genera with values of 3 or less, the range used herein to denote sensitive (intolerant) taxa. Although tolerance indices are assigned to invertebrate taxa more according to their ability to adapt to chemical degradation than for habitat instability caused by changes in water levels. In general invertebrate taxa resistant to reductions in water quality also tend to be resistant to habitat alteration. Twenty-eight genera were facultative $($ Tol. $=4-6)$ and the remainder $(35$ genera $)$ were tolerant $(T o l .=7-10)$.

The fishery below the dam described within subsequent sections of this report appears robust, suggesting that the invertebrate populations provide an adequate food base. The fish also appear be in good condition (see Section 9.0). The invertebrate data collected during the later years of the tailrace studies showed observable increases in community density, after much of the current release schedule had become operational. However, it seems unlikely that the community composition has changed appreciably, given the water quality and habitat constraints imposed upon it by impoundment.

TABLE 4.1-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY ROCK BASKET IN 1980 (JANICKI AND ROSS) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | Common Name | Tolerance Index | Percent Abundance (5) | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hydroida | Hydridae | Hydra | hydra | 4 | 2\% |  |
| Turbellaria | Planariidae | Dugesia | flat worm | 9 | 25\% | predator/sprawler |
| Oligochaeta | Naididae | Nais | naiad worm | 8 | 6\% | collector /burrower |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 3\% |  |
|  | Planorbidae | Gyraulis | orb snail | 8 | 4\% |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 32\% | shredder/sprawler |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 4\% |  |
| Diptera | Chironomidae ${ }^{(4)}$ (pupae) ${ }^{\text {Cricotopus }}$ |  | midge | 7 | 1\% | collector /burrower |
|  |  |  | midge | 8 | 3\% |  |
|  |  | Dicrotendipes | midge | 9 | 6\% | collector /burrower |
|  |  | Glyptotendipes | midge | 7 | 1\% |  |
|  |  | Polypedilum | midge | 6 | 2\% |  |
|  |  | Tanytarsus | midge | 5 | 3\% |  |
|  |  |  | Total |  | 90\% ${ }^{(3)}$ |  |

(1) Source: Janicki and Ross, 1982.
(2) Dominant taxa listed in bold print.
(3) The total does not include rare/uncommon taxa.
(4) Genera within the family Chironomidae represented 16.1 percent of the cumulative total.
(5) The mean number of specimens per Rock Basket, taken from channel and isolated pool habitats.
(6) Study Period: June through October 1980.
(7) Sample Type: Rock Basket.
(8) Pooled Data for transects A through D

## TABLE 4.2-1: BENTHIC MACROINVERTEBRATES COLLECTED BY DRIFT NET IN 1982 (RMC) FROM THE SUSQUEHANNA

 RIVER BELOW CONOWINGO DAM.
## Study Year 1982

| Order | Family or Genus | $\begin{aligned} & \text { Common } \\ & \text { Name } \end{aligned}$ | Tolerance Index | July |  | August |  | September |  | Annual Means ${ }^{(3)}$ |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\begin{gathered} \text { Density } \\ \left(\text { no. } / \mathbf{m}^{3}\right) \\ \hline \end{gathered}$ | Pct. | $\begin{gathered} \text { Density } \\ \left(\text { no. } / \mathbf{m}^{3}\right) \\ \hline \end{gathered}$ | Pct. | $\begin{gathered} \text { Density } \\ \left(\text { no. } / \mathbf{m}^{3}\right) \\ \hline \end{gathered}$ | Pct. | Density (no. $/ \mathrm{m}^{3}$ ) | Pct. |  |
| Hydroida | Hydra | hydra | 4 | 3 | 4\% | P | 0.1\% | P | 0.1\% | 1 | 3\% |  |
| Nematoda |  | Round worm | 9 | $\mathrm{P}^{(3)}$ | 0.1\% | P | P | P | 1\% | P | 0.2\% |  |
| Turbellaria | Dugesia | flat worm | 9 |  |  |  |  | P | 0.3\% | P | P |  |
| Oligochaeta |  | segmented worms | 10 |  |  |  |  | P | 0.2\% | P | P |  |
| Polychaeta | Manayunkia | fan worm | 10 |  | 1\% | P | 0.1\% | P | 3\% | P | 1\% |  |
| Bivalvia | Corbicula | Asiatic clam | 6 | P | P | P | 0.3\% | P | 0.8\% | P | P |  |
| Gastropoda | Physella | pouch snail | 7 | P | P |  |  |  |  | P | P |  |
|  | Planorbidae | orb snail | 8 | P | P |  |  | P | 0.2\% | P | 1\% |  |
|  | Ferrissia | limpet snail | 7 | P | P |  |  | P | 0.2\% | P | P |  |
| Amphipoda | Gammarus | side swimmer | 7 | 1 | 1 | P | 1\% | P | 4\% | P | 1\% |  |
| Cladocera | Leptodora | water flea | - | 76 | 83\% | P | 1\% | 1 | 4\% | 34 | 71\% | plankter |
| Copepoda | Argulus | fish louse | - | P | P |  |  | P | 0.3\% | P | P |  |
| Ephemeroptera | Baetis | mayfly | 4 | P | P |  |  | P | 0.1\% | P | P |  |
|  | Heptagenia | mayfly | 3 | P | P | P | 0.1\% | P | 0.2\% | P | P |  |
| Trichoptera | prob. Hydroptila | caddisfly | 6 |  | 0.1\% | P | 2\% | 5 | 35\% | 1 | 3\% | scraper/clinger |
|  | Cheumatopsyche | caddisfly | 6 | 2 | 2\% | 1 | 5\% | 2 | 19\% | 3 | 2\% | filterer/clinger |
|  | Leptoceridae | caddisfly | 4 |  |  | P | 0.1\% |  |  | P | P |  |
| Diptera | Chaoborus | phantom midge | 8 |  |  | P | 0.1\% | P | 1 | P | P |  |
|  | Chironomidae | midge | 7 | 9 | 9\% | 11 | 90\% | 2 | 16\% | 8 | 16\% | Collector /burrower |
|  | Tipula | cranefly | 7 |  |  |  |  | P | 0.1\% | P | P |  |
|  |  | Totals |  | 91 | 99\% | 12 | 99\% | 10 | 84\% | 47 | 100\% ${ }^{(4)}$ |  |

[^0](2) Dominant taxa ( $>10.0 \%$ of the total in one or more collections) are entered in bold print
(3) Different volumes of water were collected at the various stations/dates throughout the summer so the annual means are not cal culated by averaging the preceding columns.
(3) Different volumes of water were collected at the various stations/dates throughout the
(4) The totals for percent abundance do not include terrestrial invertebrates or fish larvae.
(5) The letter $P$ indicates taxa present at a density of less than 1 per cubic meter or that are less than 0.1 percent of the total.
(5) The letter P indicates taxa present at a density of less than 1 per cubic meter or that are less th
(7) Study year 1982 .
(7) Sample type: Drift Net (diameter $=0.5$ - meter, total volume sampled $=258.2$ cubic meters).
(8) Pooled Data: Stations 1 through 7.

TABLE 4.2-2: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER IN 1982 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Study Year: |  | 1982 |  |  |  |  |  |  |  |  |  |  |  | Dominant and Subdominant Taxon Feeding Group/Habita t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sample Type: |  | Surber Sampler (area $=\mathbf{0} .092$ square meter) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | Pooled Data: Stations 1 through 7 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | July |  | August |  | September |  | November |  | $\text { Annual Means }{ }^{(1)}$ |  |  |
|  |  | Common | Tol. | Density |  | Density |  | Density |  | Density |  | Density |  |  |
| Taxon: |  | Name | Index | (no./m²) | Pct. | (no./m²) | Pct. | (no./m) | Pct. | (no./m) | Pct. | (no./m ${ }^{2}$ ) | Pct. |  |
| Hydroida |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Hydra | hydra | 4 | 3 | 0.3\% |  |  | 27 | 1\% |  |  | 8 | 0.4\% |  |
| Nematoda |  | round worm | 9 |  |  |  |  | $\mathrm{P}^{(3)}$ | 0.1\% | P | 0.2\% | P | 0.1\% |  |
| Turbellaria |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Dugesia | flat worm | 8 | 48 | 4\% | 11 | 0.4\% | 85 | 2\% | P | 0.2\% | 38 | 2\% |  |
|  | Hirudinia | leech | 6 |  |  |  |  | P | P | P | 0.1\% | P | P |  |
| Oligochaeta |  | segmented worms | 10 | 8 | 1\% | 11 | 0.4\% | 102 | 3\% | 97 | 8\% | 57 | 3\% | collector / burrower |
| Bivalvia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Corbicula | Asiatic clam | 6 | 13 | 1\% | 22 | 1\% | 139 | 4\% | 386 | 34\% | 147 | 7\% | filterer / burrower |
| Gastropoda |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Physella | pouch snail | 7 | 3 | 0.3\% |  |  | P | P | 3 | 0.3\% | P | 0.1\% |  |
|  | Planorbida <br> e | orb snail | 8 |  |  |  |  | 12 | 0.3\% | 33 | 3\% | 12 | 1\% |  |
|  | prob. <br> Goniobasis | horn snail | 7 | 34 | 3\% | 34 | 1\% | 74 | 2\% | 66 | 6\% | 53 | 3\% | scraper/ climber |
|  | Ferrissia | limpet snail | 7 |  |  | 4 | 0.2\% | 37 | 1\% | 35 | 3\% | 20 | 1\% |  |
| Amphipoda |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Gammarus | side swimmer | 7 | 252 | 21\% | 46 | 2 | 199 | 6\% | 359 | 31\% | 224 | 11\% | shredder / sprawler |
| Cladocera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Leptodora | water flea | - |  |  |  |  | P | 0.1\% | P | 0.2\% | P | P |  |
| Isopoda |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Caecidotea | sow bug | 6 |  |  | 1 | 0.1\% | P | 0.1\% |  |  | P | P |  |
| Mysidacea |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Mysis | opossum shrimp | - |  |  |  |  | P | 0.1\% |  |  | P | P |  |
| Hydracarina |  | water mite | 7 |  |  |  |  | P | P |  |  | P | P |  |

Table 4.2-2 Cont.

| Study Year: |  | 1982 |  |  |  |  |  |  |  |  |  |  |  | Dominant <br> and <br> Subdominan <br> t Taxon <br> Feeding <br> Group/Habi <br> tat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sample Type: |  | Surber Sampler (area $=0.092$ square meter) |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | Pooled Data: Stations 1 through 7 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | July |  | August |  | September |  | November |  | $\text { Annual Means }{ }^{(1)}$ |  |  |
|  |  | Common | Tol. | Density |  | Density |  | Density |  | Density |  | Density |  |  |
| Taxon: |  | Name | Index | ( $\mathrm{no} . / \mathrm{m}^{2}$ ) | Pct. | ( $\mathrm{no} . / \mathrm{m}^{2}$ ) | Pct. | ( $\mathrm{no} . / \mathrm{m}^{2}$ ) | Pct. | ( $\mathrm{no} . / \mathrm{m}^{2}$ ) | Pct. | ( $\mathrm{no} . / \mathrm{m}^{2}$ ) | Pct. |  |
| Ephemeroptera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Baetis | mayfly | 4 |  |  | 1 | 0.1\% |  |  |  |  | P | P |  |
|  | Heptagenia | mayfly | 3 | 24 | 2\% | 26 | 1\% | 12 | 0.3\% | 10 | 1\% | 22 | 1\% |  |
|  | Tricorythodes | mayfly | 4 | 1 |  |  |  | 7 | 0.2\% |  |  | P | 0.1\% |  |
| Odonata |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Calopteryx | damselfly | 8 |  |  | 1 | 0.1\% |  |  |  |  | P | P |  |
| Megaloptera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Chauloides | alderfly | 1 |  |  |  |  |  |  | P | 0.1\% | P | P |  |
| Trichoptera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | prob. Hydroptila | caddisfly | 6 | 7 | 1\% | 261 | 10\% | 112 | 3\% | 25 | 2\% | 92 | 4\% | $\begin{gathered} \text { scraper / } \\ \text { clinger } \\ \hline \end{gathered}$ |
|  | Cheumatopsyche | caddisfly | 6 | 395 | 33\% | 509 | 20\% | 755 | 21\% | 16 | 1\% | 415 | 20\% | clinger |
|  | Neotrichia | caddisfly | 2 |  |  |  |  | P | P |  |  | P | P |  |
|  | Polycentropus | caddisfly | 6 |  |  |  |  | P | P |  |  | P | P |  |
| Lepidoptera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Petrophila | moth | 5 |  |  |  |  |  |  | P | 0.1\% | P | P |  |
| Diptera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Chironomidae | midge | 7 | 384 | 33\% | 1,634 | 62\% | 1,987 | 56\% | 46 | 4\% | 978 | 47\% | collector / burrower |
|  | Simulium | black fly | 6 |  |  | 1 | 0.1\% |  |  |  |  | P | P |  |
|  |  | Totals |  | 1,171 | 99\% | 2,561 | 99\% | 3,548 | 100\% | 1,075 | 94\% | 2,065 | 99\% ${ }^{(2)}$ |  |

(1) Source: RMC, 1985a
(2) Dominant taxa listed in bold print
(3) Different numbers of samples were collected at the various stations/dates throughout the year so the annual means are not calculated by averaging the preceding columns.
(4) The totals for Density and Percent Abundance do not include rare/uncommon taxa.
(5) The letter $P$ indicates taxa present at a density of less than 1 per square foot or that are less than 0.1 percent of the total (a cum. of 25 genera were collected)
(4) Dominant taxa ( $>10.0 \%$ of the total in one or more collections) are entered in bold print.
(6) Study year 1982.

TABLE 4.2-3 : COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T- SAMPLER IN 1982 (WESIBERG AND JANICKI) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | $\begin{aligned} & \text { Common } \\ & \text { Name } \end{aligned}$ | Tolerance Index | Frequency of occurrence | $\begin{aligned} & \text { Density } \\ & \text { (no. } / \mathrm{m}^{2} \text { ) } \end{aligned}$ | Percent Abundance | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Nematoda |  |  | round worm | 9 |  |  | $\mathrm{P}^{(3)}$ |  |
| Hoplonemertea | Tetrastemmatidae | Prostoma | proboscis worm | 7 |  |  | P |  |
| Turbellaria | Tricladida | Dugesia | flat worm | 9 | 99 | 1,552 | 12\% | predator/sprawler |
| Oligochaeta | Naididae | Dero | naiad worm | 8 | 96 | 776 | 6\% | collector/burrower |
|  |  | Nais | naiad worm | 8 |  |  | P |  |
|  |  | Pristina | naiad worm | 8 |  |  | P |  |
|  |  | Stylaria | naiad worm | 8 |  |  | P |  |
|  | Tubificinae | Aulodrilus | tube worm | 10 |  |  | P |  |
|  |  | Bothrioneureum | tube worm | 8 |  |  | P |  |
|  |  | Branchiura | tube worm | 8 |  |  | P |  |
|  |  | imm. Tubificid w/hair chaetae | tube worm | 8 |  |  | P |  |
| Hiridinea | Glossophoniidae | Batracobdella | leech | 6 |  |  | P |  |
|  |  | Helobdella | leech | 6 |  |  | P |  |
| Polychaeta | Sabellidae | Manyunkia | fan worm | 10 | 113 | 2,776 | 22\% | filterer/burrower |
| Bivalvia | Corbiculidae | Corbicula | Asiatic clam | 6 | 59 | 720 | 6\% | filterer/burrower |
|  | Sphaeriidae |  |  |  | 70 | 164 | 1\% |  |
|  |  | Musculium | fingernail clam | 6 |  |  | P |  |
|  |  | Pisidium | pill clam | 6 |  |  | P |  |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 74 | 323 | 3\% |  |
|  | Lymnaeidae | prob. Fossaria | pond snail | 7 |  |  |  |  |
|  | Physidae | Physella | pouch snail | 7 |  |  |  |  |
|  | Planorbidae | Gyraulus | orb snail | 8 |  |  |  |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 101 | 713 | 6\% | shredder/ sprawler |
| Isopoda | Asellidae | Caecidotea | sowbug | 6 |  |  | P |  |
| Hydrachnidia |  |  | water mite | 7 |  |  | P |  |

Table 4.2-3: Cont.

| Order | Family | Genus | Common Name | Tolerance Index | Frequency ${ }^{(2)}$ of occurrence | $\begin{aligned} & \text { Density } \\ & \left(\mathrm{no} . / \mathbf{m}^{2}\right) \end{aligned}$ | Percent Abundance | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ephemeroptera | Heptaganiidae | Heptagenia | mayfly | 3 |  |  | P |  |
|  |  | Maccaffertium | mayfly | 5 |  |  | P |  |
|  | Isonychiidae | Isonychia | mayfly | 2 |  |  | P |  |
| Odonata | Libellulidae |  | dragonfly | 9 |  |  | P |  |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 80 | 1,086 | 9\% | filterer/clinger |
|  | Polycentropodidae | Hydroptila | caddisfly | 6 |  |  | P |  |
|  |  | Cyrnellus | caddisfly | 8 |  |  |  |  |
| Diptera | Chironomidae |  | midge | 7 | 120 | 3,747 | 30\% | collector/burrower |
|  | Simuliidae | Simulium | black fly | 6 |  |  | P |  |
|  | Tabanidae |  | deer fly | 6 |  |  | P |  |
|  |  |  | Totals |  |  | 12,270 | 98\% |  |

(1) Source: Wesiberg and Janicki, 1983.
(2) Dominant taxa listed in bold print.
(3) Study year 1982 .
(4) Sample type: T-sampler (area $=0.025$ square meter)
(5) Pooled data transect D.
(6) The letter P indicates taxa present at a density of less than 1 per cubic meter or that are less than 0.1 percent of the total.

TABLE 4.3-1: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER SAMPLER 1983 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family or Genus | CommonName | Tolerance Index | Winter |  | Summer |  | September ${ }^{(5)}$ |  | Fall |  | $\underset{\text { Annual }}{ }$ <br> Means ${ }^{(2)}$ |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\begin{gathered} \text { Density } \\ \left(\mathrm{no} . / \mathrm{m}^{2}\right) \\ \hline \end{gathered}$ | Pct. | Density (no./m²) | Pct. | Density (no./m²) | Pct. | $\begin{gathered} \text { Density } \\ \left(\text { no. } / \mathbf{m}^{2}\right) \end{gathered}$ | Pct. | Density $\text { (no. } \left./ \mathrm{m}^{2}\right)$ | Pct. |  |
| Hydroida | Hydra | hydra | 4 | - | - | - | - | - | - | - | - | - | $\mathrm{P}^{(4)}$ |  |
| Nematoda |  | round worm | 9 | - | - | - | - | - | - | - | - | - | P |  |
| Turbellaria | Dugesia | flat worm | 8 | - | - | - | - | - | - | - | - | - | P |  |
|  | Hirudinia | leech | 6 | - | - | - | - | - | - | - | - | - | P |  |
| Oligochaeta |  | segmented worms | 10 | 1,341 | 82\% | 185 | 5\% | 793 | 30\% | 102 | 10\% | 567 | 24\% | collector/ burrower |
|  | Branchiobdellidae | commensal worm | 6 | - | - | - | - | - | - | - | - | - | P |  |
| Polychaeta | Manyunkia | fan worm | 10 | - | - | - | - | - | - | - | - | - | P |  |
| Bivalvia | Pyganodon | freshwater mussel | 4 | - | - |  | - | - | - | - | - | - | P |  |
|  | Corbicula | Asiatic clam | 6 | 17 | 1\% | 99 | 3\% | 95 | 40\% | 173 | 16\% | 105 | 4\% | filterer/burrower |
| Gastropoda | Bithynia | mud snail | 7 | - | - | - | - | - | - | - | - | - | P |  |
|  | Physella | pouch snail | 7 | - | - | - | - | - | - | - | - | - | P |  |
|  | Planorbidae | orb snail | 8 | - | - | - | - | - | - | - | - | - | P |  |
|  | prob. Goniobasis | horn snail | 7 | - | - | 48 | 1\% | 32 | 1\% | 177 | 17\% | 63 | 3\% | scraper/climber |
|  | Ferrissia | limpet snail | 7 | 2 | 0.1\% | 5 | 0.2\% | 41 | 2\% | 48 | 4\% | 32 | 1\% |  |
| Amphipoda | Gammarus | $\begin{gathered} \text { side } \\ \text { swimmer } \end{gathered}$ | 7 | 23 | 1\% | 288 | 8\% | 72 | 3\% | 336 | 31\% | 170 | 7\% | shredder/sprawler |
| Decapoda | Cambaridae | crayfish | 6 | - | - | - | - | - | - | - | - | - | P |  |
| Hydracarina |  | water mite | 7 | - | - | - | - | - | - | - | - | - | P |  |
| Isopoda | Caecidotea | sow bug | 6 | - | - | - | - | - | - | - | - | - | P |  |
| Ephemeroptera | Baetis | mayfly | 4 | - | - | - | - | - | - | - | - | - |  |  |
|  | Caenis | mayfly | 7 | 2 | 0.1\% | - | - | 90 | 3\% | - | - | 43 | 2\% |  |
|  | Heptagenia | mayfly | 3 | 2 | 0.1\% | - | 0.1\% | 41 | 2\% | 5 | 1\% | 23 | 1\% |  |
|  | Isonychia | mayfly | 2 | - | - | - | - | - | - | - | - | - | P |  |
|  | Maccaffertium | mayfly | 5 | - | - | - | - | - | - | - | - | - | P |  |
|  | Tricorythodes | mayfly | 4 | - | - | - | - | - | - | - | - | - | P |  |

Table 4.3-1: Cont.

| Order | Family or Genus | Common Name | Tolerance Index | Winter |  | Summer |  | September ${ }^{(5)}$ |  | Fall |  | Annual Means ${ }^{(2)}$ |  | DominantandSubdominant TaxonFeedingGroup/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Density $\left(\mathrm{no} . / \mathrm{m}^{2}\right)$ | Pct. | $\begin{aligned} & \text { Density } \\ & \text { (no./m²) } \\ & \hline \end{aligned}$ | Pct. | $\begin{gathered} \text { Density } \\ \left(\mathrm{no} . / \mathrm{m}^{2}\right) \\ \hline \hline \end{gathered}$ | Pct. | Density $\left(\mathrm{no} . / \mathrm{m}^{2}\right)$ | Pct. | $\begin{gathered} \text { Density } \\ \left(\mathrm{no} . / \mathbf{m}^{2}\right) \\ \hline \hline \end{gathered}$ | Pct. |  |
| Odonata | Erythemis | dragonfly | 7 | - | - | - | - | - | - | - | - | - | P |  |
| Megaloptera | Chauloides | alderfly | 1 | - | - | - | - | - | - | - | - | - | P |  |
| Trichoptera | prob. Hydroptila | caddisfly | 6 | 12 | 1\% | 120 | 3\% | 114 | 4\% | 7 | 1\% | 83 | 3 | scraper/ <br> clinger |
|  | Cheumatopsyche | caddisfly | 6 | 14 | 1\% | 1,060 | 30\% | 115 | 4\% | 17 | 2\% | 286 | 12\% | filterer/ clinger |
|  | Leptoceridae | caddisfly | 4 | - | - | - | - | - | - | - | - | - | P |  |
|  | Polycentropus | caddisfly | 6 | - | - | - | - | - | - | - | - | - | P |  |
| Lepidoptera | Petrophila | moth | 5 | - | - | - | - | - | - | - | - | - | P |  |
| Coleoptera | Stenelmis | riffle beetle | 5 | - | - | - | - | - | - | - | - | - | P |  |
|  | Psephenus | riffle beetle | 4 | - | - | - | - | - | - | - | - | - | P |  |
| Diptera | Chaoborus | phantom midge | 8 | - | - | - | - | - | - | - | - | - | P |  |
|  | Chironomidae | midge | 7 | 151 | 9\% | 1,624 | 46\% | 1,138 | 43\% | 103 | 10\% | 932 | 39\% | collector/ burrower |
|  | Culicidae | mosquito | 8 | - | - | - | - | - | - | - | - | - | P |  |
|  | Culicoides | biting midge | 10 | - | - | - | - | - | - | - | - | - | P |  |
|  | Hemerodromia | dance fly | 8 | - | - | - | - | - | - | - | - | - | P |  |
|  | Simulium | black fly | 6 | - | - | - | - | - | - | - | - | - | P |  |
|  | Tabanidae | deer fly | 6 | - | - | - | - | - | - | - | - | - | P |  |
|  | Tipulidae | crane fly | 5 | - | - | - | - | - | - | - | - | - | P |  |
|  |  | Totals |  | 1,565 | 95\% | 3,428 | 97\% | 2,530 | 95\% | 968 | 90\% | 2,303 | $\mathbf{9 5 \%}{ }^{(3)}$ |  |

(1) Source: RMC, 1985 b
(2)
Different numbers of samples were collected at the various stations/dates throughout the year so the annual means are not calculated by averaging the preceding columns.
(2) Different numbers of samples were collected at the various stations/dates throughou
(4) The letter P indicates taxa less than 1.0 percent of the annual total (a cum. of 41 genera were collected)
(5) Combination of two sample dates: 15 and 21 September (prior to and imm. after minimum flow release).
(6) Dominant taxa ( $>10.0 \%$ of the total in one or more collections) are entered in bold print.
(7) Combination of two sample dates: 15 and 21 September (prior to and imm. after minimum flow release).
(8)
(8) Study year 1983 .
(9) Gea type $=$ Surber sampler $($ area $=0.092$ square meter $)$
(10) Pooled Data: Stations 1 through 7 .

TABLE 4.3-2: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T-SAMPLER IN 1983-1984 (WEISBERG AND JANICKI) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | Common Name | Tolerance Index | Frequency ${ }^{(3)}$ of occurrence | Density (nos (no. $\mathrm{m}^{2}$ ) | Percent Abundance | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Turbellaria | Planariidae | Dugesia | flat worm | 9 | 109 | 778 | 6\% | predator/sprawler |
| Hoplonemertea | Tetrastemmatidae | Prostoma | proboscis worm | 7 | 71 | 71 | 1\% |  |
| Oligochaeta |  |  | segmented worms | 10 | 130 | 969 | 7\% |  |
| Polychaeta | Sabellidae | Manyunkia | fan worm | 10 | 118 | 2,020 | 14\% | filterer/burrower |
|  |  |  | clams | - | 24 | 133 | 1\% |  |
| Bivalvia | Corbiculidae | Corbicula | Asiatic clam | 6 | 76 | 304 | 2\% |  |
|  | Sphaeriidae | prob. Pisidium | pill clam | 6 | 107 | 685 | 5\% | filterer/burrower |
|  |  | or Musculium |  |  |  |  |  |  |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 91 | 689 | 5\% | scraper/climber |
|  | Pleuroceridae | prob. <br> Goniobasis | horn snail | 7 | 75 | 222 | 2\% |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 131 | 2,427 | 17\% | shredder/sprawler |
|  |  |  | caddisfly | 6 | 43 | 210 | 2\% |  |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 92 | 732 | 5\% | filterer/clinger |
|  | Polycentropodidae | Cyrnellus | caddisfly | 8 | 38 | 69 | 1\% |  |
|  | Hydroptilidae | prob. Hydroptila | caddisfly | 6 | 67 | 182 | 1\% |  |
| Diptera | Chironomidae |  | midge | 7 | 160 | 4,144 | 29\% | collector/burrower |
|  |  |  | Totals |  |  | 13,635 | 97\% |  |

(1) Source: Weisberg, S.B. and A. J. Janicki, 1985.
(2) Dominant taxa listed in bold print
(3) Number of samples out of a total of 160 in which this taxon was found
(4) The remaining 2.9 percent of the total benthic abundance included 33 taxa (genus/species identifications).
(5) The total number of taxa collected during 1982-1983 for the Project was 48
(6) Study period July 1983 through February 1984
(7) Sampler type: $T$ - Sampler (area $=0.025$ square meter).
(8) Pooled Data: Transects B and D

TABLE 4.4-1: BENTHIC MACROINVERTEBRATES COLLECTED BY SURBER SAMPLER IN 1984 (RMC) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family or Genus | $\begin{aligned} & \text { Common } \\ & \text { Name } \end{aligned}$ | Tolerance Index | Summer |  | September ${ }^{(6)}$ |  | Fall |  | Annual <br> Means ${ }^{(2)}$ |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\begin{aligned} & \hline \text { Density } \\ & \left(\mathrm{no} . / \mathbf{m}^{2}\right) \end{aligned}$ | Pct. | $\begin{aligned} & \hline \text { Density } \\ & \left(\text { no. } / \mathbf{m}^{2}\right) \end{aligned}$ | Pct. | $\begin{aligned} & \hline \text { Density } \\ & \left(\mathrm{no} . / \mathrm{m}^{2}\right) \\ & \hline \end{aligned}$ | Pct. | $\begin{gathered} \text { Density } \\ \left(\mathrm{no} . / \mathrm{m}^{2}\right) \end{gathered}$ | Pct. |  |
| Hydroida | Hydra | hydra | 4 | 8 | 0.3\% | 24 | 2\% | 175 | 8\% | 53 | 2\% | predator/clinger |
| Nematoda |  | round worm | 9 | - | - | - | - | - | - | - | $\mathrm{P}^{(4)}$ |  |
| Turbellaria | Dugesia | flat worm | 8 | - | - | - | - | - | - | - | P |  |
|  | Hirudinia | leech | 6 | - | - | - | - | - | - | - | P |  |
| Oligochaeta |  | segmented worms | 10 | 47 | 2\% | 123 | 5\% | 829 | 39\% | 279 | 13\% | collector/burrower |
| Polychaeta | Manyunkia | fan worm | 10 | - | - | - | - | - | - | - | P |  |
| Bivalvia | Corbicula | Asiatic clam | 6 | 100 | 4\% | 204 | 11\% | 220 | 10\% | 158 | 7\% | filterer/burrower |
| Gastropoda | Bithynia | mud snail | 7 | - | - | - | - | - | - | - | P |  |
|  | Physella | pouch snail | 7 | - | - | - | - | - | - | - | P |  |
|  | Planorbidae | orb snail | 8 | - | - | - | - | - | - | - | P |  |
|  | prob. <br> Goniobasis | horn snail | 7 | 71 | 3\% | 46 | 2\% | 38 | 2\% | 55 | 3\% | scraper/climber |
|  | Ferrissia | limpet snail | 7 | 47 | 2\% | 30 | 2\% | 30 | 2\% | 38 | 2\% |  |
| Amphipoda | Gammarus | side swimmer | 7 | 73 | 3\% | 96 | 4\% | 302 | 14\% | 135 | 6\% | shredder/sprawler |
| Cladocera | Leptodora | water flea | - | - | - | - | - | - | - | - | P |  |
| Hydracarina |  | water mite | 7 | - | - | - | - | - | - | - | P |  |
| Isopoda | Caecidotea | sow bug | 6 | - | - | - | - | - | - | - | P |  |
| Ephemeroptera | Anthopotamus | mayfly | 3 | - | - | - | - | - | - | - | P |  |
|  | Baetis | mayfly | 4 | - | - | - | - | - | - | - | P |  |
|  | Caenis | mayfly | 7 | 3 | 0.2\% | 25 | 2\% | - | - | - | P |  |
|  | Heptagenia | mayfly | 3 | 23 | 1\% | 35 | 2\% | 22 | 1\% | 26 | 1\% |  |
|  | Maccaffertium | mayfly | 5 | - | - | - | - | - | - | - | P |  |
|  | Siplonuridae | mayfly | 7 | - | - | - | - | - | - | - | P |  |
|  | Stenacron | mayfly | 2 | - | - | - | - | - | - | - | P |  |
|  | Tricorythodes | mayfly | 4 | - | - | - | - | - | - | - | P |  |

Table 4.4.-1 Cont.

| Order | Family or Genus | $\begin{aligned} & \text { Common } \\ & \text { Name } \end{aligned}$ | Tolerance Index | Summer |  | September ${ }^{(6)}$ |  | Fall |  | Annual Means ${ }^{(2)}$ |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Density (no. $/ \mathrm{m}^{2}$ ) | Pct. | Density (no. $/ \mathrm{m}^{2}$ ) | Pct. | Density (no./m ${ }^{2}$ ) | Pct. | $\begin{gathered} \text { Density } \\ \left(\text { no. } / \mathrm{m}^{2}\right) \\ \hline \end{gathered}$ | Pct. |  |
| Odonata | Didymops | dragonfly | 4 | - | - | - | - | - | - | - | P |  |
|  | Libellulidae | dragonfly | 9 | - | - | - | - | - | - | - | P |  |
|  | Macromia | dragonfly | 3 | - | - | - | - | - | - | - | P |  |
| Hemiptera | Corixidae | water boatman | 6 | - | - | - | - | - | - | - | P |  |
|  | Veliidae | water strider | 8 | - | - | - | - | - | - | - | P |  |
| Megaloptera | Chauloides | alderfly | 1 | - | - | - | - | - | - | - | P |  |
| Trichoptera | prob. Hydroptila | caddisfly | 6 | 14 | 1\% | 11 | 1\% | 1 | $\begin{gathered} \hline 0.1 \\ \% \end{gathered}$ | - | P |  |
|  | Cheumatopsyche | caddisfly | 6 | 803 | 35\% | 528 | 25\% | 184 | 9\% | 591 | 27\% | filterer/clinger |
|  | Leptoceridae | caddisfly | 4 | - | - | - | - | - | - | - | P |  |
|  | Polycentropus | caddisfly | 6 | - | - | - | - | - | - | - | P |  |
| Lepidoptera | Petrophila | moth | 5 | - | - | - | - | - | - | - | P |  |
| Coleoptera | Stenelmis | riffle beetle | 5 | - | - | - | - | - | - | - | P |  |
|  | Psephenus | riffle beetle | 4 | - | - | - | - | - | - | - | P |  |
| Diptera | Chaoborus | phantom midge | 8 | - | - | - | - | - | - | - | P |  |
|  | Chironomidae | midge | 7 | 984 | 42\% | 584 | 29\% | 179 | 9\% | 679 | 31\% | collector/burrower |
|  | Hemerodromia | dance fly | 8 | - | - | - | - | - | - | - | P |  |
|  | Simulium | black fly | 6 | 48 | 2 | - | - | - | - | 2 | 1\% |  |
|  |  | Totals |  | 2,247 | 96\% | 1,706 | 84\% | 1,980 | 94\% | 2,017 | 94\% |  |

(1) Source: RMC, 1985c
(2) Different numbers of samples were collected at the various stations/dates throughout the year so the annual means are not calculated by averaging the preceding columns.
(3) The totals for Density and Percent Abundance do not include rare/uncommon taxa.
(4) The letter P indicates taxa less than 1.0 percent of the annual total (a cum. of 40 genera were collected).
(5) Dominant taxa (>10.0\% of the total in one or more collections) are entered in bold print.
(6) Combination of two sample dates: 19 and 26-27 September (prior to and imm. after minimum flow release).
(7) Study year 1984.
(8) Sampler type surber sampler (area $=0.092$ square meter)
(9) Pooled Data: Stations 1 through 7

TABLE 4.5-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T- SAMPLER 1988-1989 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | $\begin{aligned} & \text { Common } \\ & \text { Name } \end{aligned}$ | Tolerance Index | Percent Abundance ${ }^{(2)}$ |  |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Gravel Substrate | Bedrock Substrate | Average |  |
| Turbellaria | Planariidae | Dugesia | flat worm | 9 | 9\% | 1\% | 5\% | predator/sprawler |
| Oligochaeta |  |  | segmented worms | 10 | 9\% | 7\% | 8\% | collector/burrower |
| Polychaeta | Sabellidae | Manyunkia | fan worm | 10 | 12\% | 4\% | 8\% | filterer/burrower |
| Bivalvia | Corbiculidae | Corbicula | Asiatic clam | 6 | 36\% | 20\% | 28\% | filterer/burrower |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 5\% | 1\% | 3\% |  |
|  | Planorbidae | Micromenetus | orb snail | 8 | 2\% | 1\% | 2\% |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 4\% | 1\% | 3\% |  |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 7\% | 11\% | 9\% | filterer/clinger |
|  |  |  | midge | 7 | 2\% | 8\% | 5\% |  |
| Diptera | Chironomidae | Cricotopus | midge | 8 | 1\% | 9\% | 5\% | shredder/burrower |
|  |  | Polypedilum | midge | 6 | 3\% | 29\% | 16\% | shredder/clinger |
|  |  |  | Totals |  | 90\% | 92\% | 91\% |  |

(1) Source: Scott, 1991
(2) Dominant taxa ( 10 pct. or more of the totals) are listed in bold print
(3) Study period 1988 through March 1989
(4) Sampler type T- sampler (area $=0.025$ square meters)
(5) Pooled Data: Transect D ${ }^{(4)}$

TABLE 4.6-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED 1989-1990 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | Common Name | Tolerance Index | Percent Abundance ${ }^{(2)}$ |  |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Gravel <br> Substrate | Bedrock <br> Substrate | Average |  |
| Turbellaria | Planariidae | Dugesia | flat worm | 9 | 8\% | 1\% | 5\% | predator/sprawler |
| Oligochaeta |  |  | segmented worms | 10 | 4\% | 7\% | 6\% | collector/burrower |
| Polychaeta | Sabellidae | Manyunkia | fan worm | 10 | 36\% | 23\% | 30\% | filterer/burrower |
| Bivalvia | Corbiculidae | Corbicula | Asiatic clam | 6 | 10\% | 5\% | 8\% | filterer/burrower |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 4\% | < $1 \%$ | 2\% |  |
|  | Planorbidae | Micromenetus | orb snail |  | 6\% | < $1 \%$ | 3\% |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 5\% | 1\% | 3\% |  |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 11\% | 13\% | 12\% | filterer/clinger |
| Diptera | Chironomidae ${ }^{(3)}$ |  | midge | 7 | 2\% | 6\% | 4\% |  |
|  |  | Cricotopus | midge | 8 | 1\% | 6\% | 4\% |  |
|  |  | Polypedilum | midge | 6 | 4\% | 34\% | 19\% | shredder/clinger |
|  |  |  | Totals ${ }^{(3)}$ |  | 91\% | 96\% | 94\% |  |

(1) Source: Scott, 1991
(2) Dominant taxa (10 pct. or more of the totals) are listed in bold print
(3) The totals do not include rare/uncommon taxa.
(4) A cumulative total of 115 macroinvertebrate taxa (genus/species) were collected over three winter sampling periods (1988, 1990, and 1991
(5) The Percent Abundance of all chironomid genera combined was 7 pct. from bedrock substrate and 46 pct. from bedrock substrate.
(6) The total number of samples collected was 118.
(7) Study period 1989 through March 1990 (Year 2).
(8) Sampler type T-sampler (area $=0.25$ square meters).
(9) Pooled data, transect D.

TABLE 4.7-1: COMMON BENTHIC MACROINVERTEBRATE TAXA COLLECTED BY T-SAMPLER IN 1990-1991 (SCOTT) FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Order | Family | Genus | CommonName | Tolerance Index | Percent Abundance ${ }^{(2)}$ |  |  | Dominant and Subdominant Taxon Feeding Group/Habitat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | (Gravel Substrate) | (Bedrock <br> Substrate) | Average |  |
| Turbellaria | Planariidae | Dugesia | flat worm | 9 | 5\% | 1\% | 3\% |  |
| Oligochaeta |  |  | segmented worms | 10 | 7\% | 4\% | 6\% | collector/burrower |
| Polychaeta | Sabellidae | Manyunkia | fan worm | 10 | 9\% | 26\% | 18\% | filterer/burrower |
| Bivalvia | Corbiculidae | Corbicula | Asiatic clam | 6 | 30\% | 15\% | 23\% | filterer/ burrower |
| Gastropoda | Ancylidae | Ferrissia | limpet snail | 7 | 2\% | > $0.1 \%$ | 1\% |  |
|  | Planorbidae | Micromenetus | orb snail | 8 | 1\% | > $0.1 \%$ | 1\% |  |
| Amphipoda | Gammaridae | Gammarus | side swimmer | 7 | 17\% | 4\% | 11\% | shredder/ sprawler |
| Trichoptera | Hydropsychidae | Cheumatopsyche | caddisfly | 6 | 7\% | 13\% | 10\% | filterer/clinger |
| Diptera | Chironomidae |  | midge | 7 | 8\% | 29\% | 19\% | collector/ burrower |
|  |  |  | Totals ${ }^{(2)}$ |  | 86\% | 92\% | 89\% |  |

(1) Source: Scott, 1991
(2) Dominant taxa ( 10 pct. or more of the totals) are listed in bold print.
(3) The totals do not include rare/uncommon taxa.
(4) A cumulative total of 115 macroinvertebrate taxa (genus/species) were collected over three winter sampling periods (1988, 1990, and 1991).
(5) The total number of samples collected was 107.
(6) Study period 1990 through March 1991 (Year 3)
(7) Sampler type T-sampler (area $=0.25$ square meters).
(8) Pooled data, transect D


CONOWINGO HYDROELECTRIC PROJECT PROJECT NO. 405


Figure 4.1:
Sampling Locations for Drift Organisms and Bethnic Macroinvertebrates below Conowingo Dam 1980-1991

FIGURE 4.2-1: COMMUNITY DENSITY AND DENSITY OF COMMON INVERTEBRATE TAXA COLLECTED WITH A SURBER SAMPLER FROM THE CONOWINGO DAM TAILRACE DURING 198-1984.


FIGURE 4.2-2: COMMUNITY DENSITY AND DENSITY OF COMMON INVERTEBRATE TAXA COLLECTED WITH A T-SAMPLER FROM THE CONOWINGO DAM TAILRACE DURING 1982-1991.

Figure 4.2-2 Community density and density of common invertebrate taxa collected with a T-Sampler from the Conowingo Dam Tailrace during 1982-1991.


### 5.0 FISH LIFTS

Data from both the West and East Fish lifts have been compiled in Susquehanna River Anadromous Fish Restoration Cooperative (SRAFRC)Annual Progress Reports (SRAFRC, 2006) up to 2006.

### 5.1 West Fish Lift (WFL) Operations

The Conowingo Dam Fish Passage Facility has been operated during anadromous spawning migrations since 1972 as part of a cooperative private, state, and Federal effort to restore American shad to the upper Susquehanna River. Annual operation of the WFL at Conowingo Dam has yielded a 38-year timeline of catch and proportional abundance of migratory and resident fish. Routine spring operations typically targeted the April to mid-June time period. WFL operations were expanded in some years to include some summer operation to monitor relative abundance of fishes other than alosids at low dissolved oxygen (DO) and high water temperatures conditions. Also, as part of the biologically-based study plan following the last relicensing, the WFL operations were further expanded in 1981 to include fall months as well. The WFL was used to acquire specimens of several species for various life history studies (e.g., food habits, age and growth, and movement). In 1982 and 1984 the fish lift was also operated intermittently on a monthly basis July through November to investigate seasonal variablitity in species composition and relative abundance.

The WFL is an elevator device that lifts the fish to a sorting tank or transport truck located on the west bank of the tailrace (Figure 5.1-1). There are two telescoping weir entrances to the facility, one facing approximately east, the other approximately south. There is a short holding channel (approximately 25 feet long), at the upper end of which rests the hopper. The hopper rests in a depression in the floor of the holding channel, and lifts the fish out of the facility to the tank above. Also located in the holding channel is a crowder which acts to partially trap the fish by the V-position of its gates. Attraction water is provided by the generation of Station Service Units 1 and 2. Part of this water is directed through a diffusion grating to the weir gates. The remainder passes beneath the holding channel, emerging upstream of the hopper to continue flowing through the holding channel and out the weir gates. Fish that are attracted to this flow, enter the facility through the weir gates, and swim to the upstream end of the holding channel near the hopper. When a lift is to be made, the crowder gates are closed; the crowder is moved upstream, sweeping all fish into the hopper area; then the hopper is lifted collecting all the trapped fish.

Since before 1985, most shad collected at the WFL have been sorted from the daily catch, placed into circular transport tanks, and stocked into suitable spawning waters upstream of Holtwood and Safe

Harbor dams, or utilized for a variety of other scientific purposes such as tank spawning trials. Resident fishes were returned to the river after their numbers were estimated (SRAFRC, 2006).

The current objectives of Conowingo WFL operations generally include: the collecting and enumerating of American shad, river herring and other migratory and resident fishes and obtaining American shad for an on-site tank spawning and shad egg collection program conducted at Conowingo Dam. WFL operational procedures adopted by the Susquehanna River Technical Committee (SRTC) in 1997 included limiting the period of operation to the peak six weeks of the run (late April through the first week in June) and limiting daily lift operations to 8 hours (1100-1900). Within these parameters the WFL maintained appropriate entrance velocities of 4 to 7 fps and curbed use of adjacent turbine units 1 and 2 whenever river flow dropped below $60,000 \mathrm{cfs}$.

### 5.2 East Fish Lift (EFL) Operations

Pursuant to a settlement agreement on water quality and fish passage approved by the FERC on January 24, 1989, Exelon was required to construct facilities for the protection of fish. The EFL began operation in 1991 as the cornerstone of the agreement.(46 FERC 961,063$)$ (FERC, 1989).

From 1991 through 1996, American shad and river herring were transported across Conowingo Dam and trucked to upstream release sites because the upstream dams were not equipped with upstream fishways. During the spring runs of 1991 through 1996, EFL catches were sorted as at the WFL. Beginning in 1997 all fish were lifted to the exit channel for continued volitional upstream movement following the construction and operations of fish lifts at Holtwood and Safe Harbor dams. Sorting and trucking operations from the EFL were stopped. The number of fish passed upstream at the EFL were estimated by a trained observer counting fish at the window in the viewing room. Annual operation of the EFL at Conowingo Dam has yielded a 19-year timeline of catch and proportional abundance of migratory and resident fish species. Routine spring operations typically targeted the April to mid June time period.

The EFL is located to the east of the Project's turbine units and adjacent to the deflection wall, which separates the tailrace from the spillway (Figure 5.1-2). The EFL consists of two functioning entrance channels with independent weir gates at the downstream end of each channel to regulate flow. The entrance channels are 14 ft high by 10 ft wide and can transport 300 cfs of supplemental attraction flow, creating velocities ranging from three to six fps inside the entrance gate. The specific entrance(s) used to attract fishes was dictated by the station discharge and which turbine units were operating. (SRAFRC 2006).

The EFL merges into a single crowder channel upstream of the entrance channels. Entering the crowder channel, the fish pass through the crowder gate. Once a number of fish have passed through the crowder gate, the gate is closed and the fish are trapped. The crowder screen upstream of the crowder gate is raised allowing the fish to move into the hopper area at the upstream end of the crowder channel. With the hopper sitting on the bottom of the channel, the crowder gate is moved forward, concentrating the fish into the area immediately over the hopper. The crowder screen is then lowered into position further corralling the fish, which are now ready to be lifted to the exit trough. As the 3,500 gallon hopper is raised to the exit trough, the crowder is returned downstream into its fishing position. When the hopper reaches the exit trough, a door to the hopper is opened and the fish and water within are released into the exit trough. The exit trough is 14 ft wide x 12 ft high x 190 ft long. On their own volition, the fish swim by a viewing window situated in a constricted area of the Exit Trough before heading upstream into Conowingo Pond.

The mechanical aspects of EFL operation are similar to those described in RMC (1992) and Normandeau Associates, Inc. (1999). Fishing time and/or lift frequency was determined by fish abundance, but the hopper was cycled at least hourly during daily operations. The method of lift operation was also influenced by fish abundance. In order to not overload the hopper or place undue stress on fish when many fish were in the fishing channel, the crowder was not operated; instead the crowder screen was raised and then lowered trapping fish over the hopper. This mode of operation, called "fast fish", involved leaving the crowder in the normal fishing position and raising the hopper frequently to remove fish that accumulated in the holding channel.

The fish lift data provides baseline account of the dominant species below Conowingo Dam prone to utilizing the fish lift. Numerous physical and biological factors influence the species and sizes of fish collected. These factors include: fish behavior and swimming ability, crowding, predator - prey interaction, water velocities, hydraulic patterns, turbulence, noise, smell and light intensity. Due to operational constraints, velocities and restrictions on the lower limits of screen mesh size, small fish and small-sized species are generally collected less effectively relative to larger fish (Larinier et al. 2002).

The fish lift facilities at the Conowingo Dam are designed to pass migratory fishes, particularly membersof the herring family (Clupeidae), including the American shad (Alosa sapidissima), alewife (Alosa pseudoharengus), blueback herring (Alosa aestivalis), and hickory shad (Alosa mediocris) and are primarily operated during the spring and early to summer migration period (April to early June).

### 5.3 Fish Lift Catch Per Unit Effort

From 1972 - 2009 variability in the annual totals of fish collected at the lifts was considerable. Operational methodologies were influenced by natural river flows, water temperatures, generation schedules and fish population numbers. Inter-annual variability in fish lift collections is likely due to a variety of natural and anthropogenic factors such as hydrological regimes, commercial and recreational fishing pressures, varying accessibility to habitats, yearly operational changes, and varying efforts of collection. Variability related to efforts of collection such as, scheduling, mechanical modes, hours of operation, fishing time (time crowder gates are open), number of lifts and number of days of fish lift operation impacted the overall year to year totals of fish collected. When consistently documented, collection efforts can be standardized to represent a unit of collection per effort. Based on the fish lift data presented in "Fish Lift Operation Reports"(RMC 1991 to 1996, NAI 1997 to 2010) and annual SRAFRC progress reports (SRAFRC 1982 to 2006), standardization of catch is best represented as fish per lift. Due to variability in notation and documentation of "fishing time" and "operational time", as well the EFL documenting only "estimated operational time" from 1997 through 2009 and the WFL documenting only "fishing time" in 2000 through 2009, "fish per lift" is the most practical choice for standardization of effort. As numerous operational changes have occurred over the years, fish per lift is the most consistent and finite normalization of effort that exists for the fish lift data for the purposes of characterizing the fish community.

### 5.3.1 West Fish Lift

From 1972 to 2009 the WFL collected 31,533,545 fish of 72 different taxa from 27,481 lifts (Table 5.3.11). The overall lift CPUE during the 38 years of operation was 1,148 fish per lift (Table 5.3.1-1). The highest overall catch was in 1987 (2,593,445 fish, Table 5.3.1-1) and the highest CPUE was in 1981 (2,762 fish per lift, Table 5.3.1-1). The lowest amount of fish per lift was collected in 2004 (249 fish per lift) along with the lowest total catch ( 37,589 fish, Table 5.3.1-1). Richness ranged from a high of 49 taxa collected in 1988 to a low of 30 taxa in both 2003 and 2004 (Table 5.3.1-1); the average number of taxa collected per year from 1972 to 2009 was 40 .

The 1970's fish lift collections present a relatively balanced fish assemblage (Table 5.3.1-2). Overall, the most commonly collected species 1972 to 1979 were white perch ( 534 fish per lift), gizzard shad (277 fish per lift), blueback herring ( 153 fish per lift), channel catfish ( 85 fish per lift), American eel ( 42 fish per lift), alewife ( 30 fish per lift), common carp ( 21 fish per lift) and quillback ( 14 fish per lift, Table 5.3.1-2). White perch was prevalent throughout the decade, ranging from $19 \%$ of the overall CPUE in 1977 to $56 \%$ in 1975 (Table 5.3.1-2). In 1974 white perch comprised the largest overall CPUE of the
decade for any species, 1,095 fish per lift (Table 5.3.1-2). Gizzard shad became increasingly common in the latter part of the decade. In 1973 gizzard shad represented only $4 \%$ ( 30 fish per lift) of the overall CPUE and increased to $64 \%$ (1,050 fish per lift) of the overall CPUE in 1977 (Table 5.3.1-2). Blueback herring was one of the most commonly collected species in the early 1970's representing $25 \%$ of the overall CPUE in 1973, but declined rapidly to $8 \%$ by 1975 (Table 5.3.1-2). The CPUE of blueback herring decreased from 415 fish per lift in 1974 to 8 fish per lift in 1979 (Table 5.3.1-2). Catches of channel catfish fluctuated throughout the 1970's. Channel catfish was $25 \%$ of the overall CPUE in 1972 ( 75 fish per lift), declining to $4 \%$ of the overall CPUE in 1976 ( 61 fish per lift) and increasing again to 19\% of the overall CPUE in 1979 (127 fish per lift, Table 5.3.1-2).

Amongst decades, the 1980's had the highest overall CPUE at 1,685 fish per lift (1980 to 1989, Table 5.3.1-1). The catch was dominated by gizzard shad every year, ranging from $61 \%$ of the overall CPUE in 1982 (1,692 fish per lift) to $96 \%$ of the overall CPUE in 1987 (1,760 fish per lift, Table 5.3.1-2). The highest CPUE for a species during the decade occurred in 1981 when gizzard shad were collected at a rate of 2,361 fish per lift Table 5.3.1-2. Overall, gizzard shad were collected at a CPUE of 1,540 fish per lift 1980 to 1989. Other species prevalent in the fish lift collections 1980 to 1989 included white perch ( 53 fish per lift), channel catfish ( 32 fish per lift), comely shiner ( 11 fish per lift), common carp ( 11 fish per lift), blueback herring ( 7 fish per lift) and shorthead redhorse ( 5 fish per lift, Table 5.3.1-2).

Relative abundance of fishes varied seasonally in 1982 and 1984 as did species diversity and was similar to the electrofishing catch (Section 6.5). In both years species diversity was highest in the spring and lowest in the fall, likely due to the addition of migrant clupeids in the spring and a limited sample size in the fall. Gizzard shad dominated the catch in all months (Appendix A-24). The relative abundance of gizzard shad was high during October and November due to recruitment of large numbers of young-of-the-year. Generally, channel catfish were most abundant in the spring (June) and fall (October, November). Smallmouth bass and white perch were most abundant during spring with the highest catches occurring in May and June. Walleyes were present in all months with the highest catches occurring in summer months. Juvenile striped bass were abundant during the summer months of July and August.

The 1990's had the lowest overall CPUE amongst decades at 765 fish per lift (Table 5.3.1-1). (As of 1997, the WFL operations were modified to an 8 hour day (from a 12 hour day) and a 6 week time frame.). Species relatively common throughout 1990 to 1999 included gizzard shad ( 663 fish per lift), blueback herring ( 31 fish per lift), white perch ( 29 fish per lift), American shad ( 11 fish per lift), channel catfish ( 7 fish per lift), comely shiner ( 6 fish per lift), and common carp ( 6 fish per lift, Table 5.3.1-2).

In all years except 1997, gizzard shad were the dominant species collected at the WFL. The catch of gizzard shad represented $87 \%$ of the overall CPUE 1990 to 1999 (Table 5.3.1-2). In 1997, blueback herring comprised $39 \%$ of the overall CPUE ( 218 fish per lift) while gizzard shad comprised $37 \%$ ( 207 fish per lift, Table 5.3.1-2). White perch constituted $17 \%$ of the overall CPUE ( 96 fish per lift, Table 5.3.1-2) in 1997.

Gizzard shad also dominated the collections at the WFL 2000 to 2009, ranging from $61 \%$ of the catch in 2004 ( 152 fish per lift) to $99 \%$ of the catch in 2008 ( 1,507 fish per lift, Table 5.3.1-2). White perch were more common in the early part of the decade. In 2002 white perch accounted for $16 \%$ of the overall CPUE ( 156 fish per lift); in 2006 the proportion of white perch declined to $<1 \%$ of the overall CPUE ( 3 fish per lift, Table 5.3.1-2). White perch had an overall CPUE of 50 fish per lift from 2000 to 2009. Although they comprised only $1 \%$ of the collection 2000 to 2009, channel catfish were the second most proportionally abundant fish in 2004, comprising $13 \%$ of the overall CPUE ( 32 fish per lift). Other species common 2000 to 2009 included, American shad (19 fish per lift), blueback herring ( 9 fish per lift), channel catfish ( 7 fish per lift) and alewife ( 5 fish per lift, Table 5.3.1-2).

From 1972 to 2009 gizzard shad accounted for $75 \%$ of the overall CPUE at the WFL. Gizzard shad became the dominant species for the first time in 1977 and retained its dominance over the next three decades (Figure 5.3.1-1). White perch (13\%), blueback herring (4\%), channel catfish (3\%) were other species relatively abundant in CPUE. Sixty-eight other taxa combined for a total of $5 \%$ of the overall CPUE 1972 - 2009 (Appendix A-25).

Clupeids, apart from hickory shad (gizzard shad, blueback herring, American shad, and alewife) were in the overall top ten species collected 1972 - 2009 and together comprised $80 \%$ of the number of fish collected at the WFL (Table 5.3.1-2). The Moronidae family (white perch and striped bass) were frequently in the yearly top 10 of species collected and combined, comprised $13 \%$ of the overall collection 1972 - 2009 (Table 5.3.1-2). Members of the Ictaluridae family including channel catfish, white catfish, brown bullhead, yellow bullhead, flathead catfish were well represented. Channel catfish were the fourth most frequently caught species and comprised 3\% of the overall collection 1972 to 2009. Members of the Catostomidae family including quillback, shorthead redhorse, northern hog sucker, creek chubsucker combined composed less than $1 \%$ of the overall collections; shorthead redhorse and quillback were often in the top ten annual species collected and were the tenth and eleventh most commonly collected species overall 1972 to 2009 (Appendix A-25). Centrarchids including redbreast sunfish, bluegill, white crappie, smallmouth bass, largemouth bass, pumpkinseed, and rock bass were frequently collected but combined, represented less than $1 \%$ of the total catch (Appendix A-25). Percidae including
walleye, yellow perch, tessellated darter and logperch were also frequently collected (Appendix A-25). Many different species of Cyprinids were collected (common carp, comely shiner, spotfin shiner, spottail shiner, golden shiner, bluntnose shiner, goldfish, rosyface shiner, swallowtail shiner). Common carp was the fifth most frequently collected species 1972 to 2009 , but overall comprised less than $1 \%$ of the overall total collected (Appendix A-25).

Several catadromous and estuarine species were collected at the WFL. American eel were the sixth most frequently collected species overall 1972 to 2009 (Table 5.3.1-2). Atlantic menhaden were collected in the 1970's ( 2,214 fish, Appendix A-25), were rare in the 1980's, and were not collected in the 1990's or 2000's. Sea lamprey were collected regularly (1,013 fish, Appendix A-25).

### 5.3.2 East Fish Lift

From 1991 to 2009 the EFL collected 16,411,728 fish of 63 different taxa from 12,733 lifts (Table 5.3.21). The overall catch per lift from the 19 years of operation was 1,289 fish per lift (Table 5.3.2-1). The highest overall catch ( $2,394,583$ fish ) and the highest overall CPUE ( 3,998 fish per lift) occurred in 1992 (Table 5.3.2-1). The lowest total catch occurred in 2005 (377,762 fish) and the lowest CPUE occurred in 1991 ( 557 fish per lift, Table 5.3.2-1). Richness was highest in 1992 ( 45 taxa), lowest in 2003 ( 25 taxa) and averaged 33 taxa 1991 to 2009 (Table 5.3.2-1).

Similar to catches at the WFL since 1977, gizzard shad have dominated the catch at the EFL since it began operation in 1991 (Figure 5.3.2-1), ranging from 98\% of the overall catch in 1992 to $47 \%$ of the overall catch in 2001 (Table 5.3.2-2). In all years combined (1991 to 2009), gizzard shad account for $86 \%$ of all fish collected (Table 5.3.2-2). In 1992, 2,351,351 gizzard shad were collected, the most of any species in any year (Table 5.3.2-2). In 1992, the highest CPUE for gizzard shad also occurred ( 3,925 fish per lift, Table 5.3.2-2).

Routinely, American shad were the second most frequently collected species at the EFL. From 1991 to 2009 American shad comprised $7 \%$ of the overall catch per lift (Table 5.3.2-2). The proportional abundance of American shad CPUE at the fish lift ranged from 31\% in 2000 to less than $1 \%$ in 1992 (Table 5.3.2-2). In 2001 the highest CPUE of American shad occurred (346 fish per lift, Table 5.3.2-2) and the lowest CPUE of American shad occurred in 1993 (10 fish per lift, Table 5.3.2-2).

From 1991 to 2009, blueback herring comprised 4\% of the overall CPUE at the EFL (Table 5.3.2-2). In 1997, 1999 and 2001 significant catches of blueback herring were made. In 2001, 510 herring per lift were collected (Table 5.3.2-2), the highest amount in any year and the second most proportionally abundant species that year after gizzard shad. Very few blueback herring have been collected since 2001
with none taken in 2006. Populations of blueback herring have been declining in the northeast due to a number of potential causes including habitat loss, targeted or by catch at sea via commercial fishing and increased numbers of striped bass and other types of predators (ASMFC, 2009).

Other species frequently collected and proportionally abundant at the EFL included: white perch ( 9 fish per lift, $<1 \%$ of overall CPUE), common carp ( 5 fish per lift, $<1 \%$ of overall CPUE), quillback ( 2 fish per lift, $<1 \%$ of overall CPUE), comely shiner ( 1 fish per lift, $<1 \%$ of overall CPUE), channel catfish ( 1 fish per lift, $<1 \%$ of overall CPUE), and walleye ( 1 fish per lift, $<1 \%$ of overall CPUE, Table 5.3.2-2).

Over the operating history at the EFL (1991 to 2009), members of the clupeid family including gizzard shad, American shad and blueback herring were the overall top three species collected and comprised $97 \%$ of all fish collected (Table 5.3.2-2). Alewife was the eleventh most frequently collected species and in 2001 had an overall CPUE of 13 fish per lift (Table 5.3.2-2), the highest of any year. Another clupeid, hickory shad, was collected occasionally (Appendix B-23). The Moronidae family, represented by white perch and striped bass, were frequently in the yearly top 10 of species collected though combined, were less than $1 \%$ of the total fish. Striped bass was the tenth most frequently collected species and had an overall CPUE of less than1 fish per lift (Table 5.3.2-2). The highest CPUE for striped bass occurred in 1998 (2 fish per lift, Table 5.3.2-2).

Many different species of Cyprinids were collected (common carp, comely shiner, spotfin shiner, spottail shiner, golden shiner, blacknose dace, creek chub, longnose dace, bluntnose minnow, Appendix B-23). Common carp was the fifth most frequently collected species overall, and populations have trended downward since the early 1990's. In 1991 a high CPUE of 20 fish per lift were collected, in 2006 a low of less than 1 fish per lift was collected (Table 5.3.2-2). Comely shiner was the seventh most frequently collected species with a CPUE of 1 fish per lift 1991 to 2009 (Table 5.3.2-2). The Catostomidae family (quillback, shorthead redhorse, northern hog sucker, creek chubsucker) combined composed less than $1 \%$ of the overall collections, but shorthead redhorse and quillback were often in the top ten annual species collected (Appendix B-23). Quillback was the sixth most frequently collected species, with an overall CPUE of 1 fish per lift 1991 to 2009 (Table 5.3.2-2).

Species from the Ictaluridae family collected at the EFL included channel catfish, brown bullhead, yellow bullhead and white catfish (Appendix B-23). Channel catfish were routinely in the top ten species collected and ranged in CPUE from a high of fish per lift in 2009 to a low of less than 1 fish per lift in 2001 (Appendix B-23). Species of the Percidae family including walleye, yellow perch, tessellated darter, logperch and shield darter were also regularly collected (Appendix B-23). Walleye were the ninth most
frequently collected species and catches fluctuated from 4 fish per lift in 2008 to a low of less than 0.1 fish per lift in 1993 (Table 5.3.2-2). Centrarchids, including smallmouth bass, redbreast sunfish, largemouth bass, rock bass, pumpkinseed, green sunfish, white crappie and black crappie were also regularly collected throughout 1991 to 2009 (Appendix B-23).

Several catadromous and estuarine species were collected at the EFL. A total of 566 American eel was collected 1991 to 2009 (Appendix B-23). The highest collection rate occurred in 1992 at only less than 1 fish per lift; no American eels were collected in 2000, 2002 to 2004 or 2007 to 2009 (Appendix B-23). Sea lamprey were common at the EFL (1,245 fish, Appendix B-23).

### 5.4 Discussion

As documented by WFL data, a fish assemblage consisting of white perch, blueback herring, channel catfish, gizzard shad and other species in the 1970's has become increasingly dominated by gizzard shad (Figure 5.3.1-1). Throughout many aquatic systems within their range, gizzard shad populations increased dramatically in the mid $20^{\text {th }}$ century as broad based ecological changes on the landscape provided a potential increase in suitable habitats (Miller 1960). These same changes, while beneficial to gizzard shad, may have provided unfavorable habitat conditions for other fish species.

Gizzard shad thrive in warm, shallow bodies of water that have a soft mud bottom, high turbidity, and relatively few predators. Gizzard shad are known to be near-exclusively herbivorous throughout much of their life, feeding heavily on microscopic plant life, phytoplankton, and algae. However, in very early life stages consume zooplankton, often to the detriment of other young fishes, including species popularly targeted by anglers (NAI, 1994). They have a high reproductive capacity and grow rapidly, potentially aiding in the avoidance of predation (Miller 1960).

Gizzard shad were inadvertently introduced into Conowingo Pond in 1972 and the population has increased exponentially. In 1997 the introduction of volitional passage at the EFL into Conowingo Pond exacerbated the population growth.

As gizzard shad have trended upward in population, many other species have declined. White crappie catches at the WFL have declined substantially since the mid 1970's (Figure 5.3.1-1). It has been noted that one of the primary mechanisms of low recruitment of white crappie is the competition for zooplankton with juvenile gizzard shad (NAI, 1994).

TABLE 5.3.1-1: TOTAL ANNUAL CATCH OF FISHES PER YEAR AT THE CONOWINGO DAM WEST FISH LIFT 1972-2009.

| Year | $\mathbf{1 9 7 2}$ | $\mathbf{1 9 7 3}$ | $\mathbf{1 9 7 4}$ | $\mathbf{1 9 7 5}$ | $\mathbf{1 9 7 6}$ | $\mathbf{1 9 7 7}$ | $\mathbf{1 9 7 8}$ | $\mathbf{1 9 7 9}$ | Totals <br> $\mathbf{1 9 7 2 - 1 9 7 9}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 54 | 62 | 58 | 55 | 63 | 61 | 35 | 29 | 417 |
| Lifts | 817 | 1,527 | 819 | 514 | 684 | 707 | 358 | 301 | 5,727 |
| Est. Oper. <br> Time (HR) | 608 | 996 | 500 | 500 | 307 | 375 | 413 | 187 | 3,886 |
| Fishing <br> Time (HR) | 313 | 623 | 222 | 222 | 189 | 252 | 136 | 123 | 2,080 |
| \#Taxa | 40 | 43 | 42 | 41 | 38 | 40 | 44 | 37 | 55 |
| Total | 241,419 | $1,300,345$ | $1,617,887$ | 917,043 | $1,175,616$ | $1,169,061$ | 276,045 | 197,769 | $6,895,185$ |
| Fish per <br> Lift | 296 | 852 | 1,975 | 1,784 | 1,719 | 1,654 | 771 | 657 | 1,204 |


| Year | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | $\begin{gathered} \text { Totals } \\ \mathbf{1 9 8 0 - 1 9 8 9} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 30 | 37 | 44 | 29 | 34 | 55 | 59 | 60 | 63 | 51 | 462 |
| Lifts | 403 | 490 | 725 | 648 | 519 | 1,118 | 831 | 1,414 | 1,330 | 1,117 | 8,595 |
| Est. Oper. Time (HR) | 221 | 275 | 502 | 299 | 251 | 542 | 546 | 639 | 637 | 539 | 4,451 |
| Fishing Time (HR) | 117 | 178 | 336 | 224 | 192 | 421 | 449 | 532 | 513 | 457 | 3,419 |
| \#Taxa | 42 | 48 | 46 | 40 | 35 | 41 | 43 | 46 | 49 | 45 | 72 |
| Total | 372,379 | 1,353,310 | 1,403,176 | 1,028,092 | 957,821 | 2,317,797 | 1,830,569 | 2,593,445 | 1,592,938 | 1,035,121 | 14,484,648 |
| Fish per Lift | 924 | 2,762 | 1,935 | 1,587 | 1,846 | 2,073 | 2,203 | 1,834 | 1,198 | 927 | 1,685 |

Table 5.3.1-1: Cont.

| Year |  |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | $\begin{gathered} \text { Totals } 2000- \\ 2009 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Totals } \\ \mathbf{1 9 7 2 - 2 0 0 9} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 34 | 41 | 31 | 31 | 14 | 30 | 37 | 29 | 34 | 28 | 309 | 1,695 |
| Lifts | 424 | 425 | 417 | 367 | 151 | 295 | 349 | 288 | 481 | 282 | 3,479 | 27,481 |
| Est. <br> Oper. <br> Time <br> (HR) | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| Fishing Time (HR) | 206 | 195 | 147 | 171 | 74 | 166 | 215 | 135 | 174 | 144 | 1,627 | 11,436 |
| \#Taxa | 37 | 38 | 35 | 30 | 30 | 36 | 38 | 35 | 37 | 39 | 53 | 72 |
| Total | 458,349 | 309,804 | 419,103 | 147,388 | 37,589 | 94,767 | 163,131 | 159,389 | 733,553 | 225,794 | 2,748,867 | 31,533,545 |
| Fish per lift | 1,081 | 729 | 1,005 | 402 | 249 | 321 | 467 | 553 | 1,525 | 801 | 790 | 1,148 |

TABLE 5.3.1-2: TOP TEN TOTALS OF ANNUAL CATCH OF FISHES BY YEAR, CPUE (FISH PER LIFT) AND PROPORTIONAL ABUNDANCE AT THE CONOWINGO DAM EAST AND WEST FISH LIFTS:YEARS 1972-2009.

| 1972 |  |  |  | 1973 |  |  |  | 1974 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Channel Catfish | 61,042 | 75 | 25\% | White Perch | 647,493 | 424 | 50\% | White Perch | 897,113 | 1,095 | 55\% |
| Blueback Herring | 58,198 | 71 | 24\% | Blueback Herring | 330,341 | 216 | 25\% | Blueback Herring | 340,084 | 415 | 21\% |
| White Perch | 50,991 | 62 | 21\% | Alewife | 144,727 | 95 | 11\% | Gizzard Shad | 119,672 | 146 | 7\% |
| Gizzard Shad | 24,849 | 30 | 10\% | Channel Catfish | 55,084 | 36 | 4\% | American Eel | 91,937 | 112 | 6\% |
| Alewife | 10,345 | 13 | 4\% | Gizzard Shad | 45,668 | 30 | 4\% | Channel Catfish | 75,663 | 92 | 5\% |
| Quillback | 7,119 | 9 | 3\% | Quillback | 27,780 | 18 | 2\% | Common Carp | 34,383 | 42 | 2\% |
| Yellow Perch | 5,955 | 7 | 2\% | Common Carp | 16,362 | 11 | 1\% | Alewife | 16,675 | 20 | 1\% |
| White Crappie | 4,457 | 5 | 2\% | White Catfish | 6,394 | 4 | 0.5\% | Quillback | 14,565 | 18 | 1\% |
| Common Carp | 4,370 | 5 | 2\% | Brown Bullhead | 5,328 | 3 | 0.4\% | White Crappie | 4,371 | 5 | 0.3\% |
| Striped Bass | 3,142 | 4 | 1\% | Shorthead Redhorse | 4,420 | 3 | 0.3\% | Comely Shiner | 3,870 | 5 | 0.2\% |


| 1975 |  |  |  | 1976 |  |  |  | 1977 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| White Perch | 511,699 | 996 | 56\% | White Perch | 568,018 | 830 | 48\% | Gizzard Shad | 742,056 | 1,050 | 63\% |
| Gizzard Shad | 139,222 | 271 | 15\% | Gizzard Shad | 382,275 | 559 | 33\% | White Perch | 224,843 | 318 | 19\% |
| Channel Catfish | 74,042 | 144 | 8\% | American Eel | 60,409 | 88 | 5\% | Channel Catfish | 90,442 | 128 | 8\% |
| Blueback Herring | 69,916 | 136 | 8\% | Spotfin Shiner | 45,879 | 67 | 4\% | Blueback Herring | 24,395 | 35 | 2\% |
| American Eel | 64,375 | 125 | 7\% | Channel Catfish | 41,508 | 61 | 4\% | Common Carp | 16,256 | 23 | 1\% |
| Common Carp | 15,114 | 29 | 2\% | Blueback Herring | 35,519 | 52 | 3\% | American Eel | 14,601 | 21 | 1\% |
| White Crappie | 9,290 | 18 | 1\% | Quillback | 9,882 | 14 | 1\% | Redbreast Sunfish | 8,277 | 12 | 1\% |
| Quillback | 8,388 | 16 | 1\% | Common Carp | 6,755 | 10 | 1\% | Spottail Shiner | 8,107 | 11 | 1\% |
| White Catfish | 6,178 | 12 | 1\% | Redbreast Sunfish | 3,772 | 6 | 0.3\% | Spotfin Shiner | 7,960 | 11 | 1\% |
| Alewife | 4,311 | 8 | 1\% | White Crappie | 2,987 | 4 | 0.3\% | Quillback | 6,734 | 10 | 1\% |

Table 5.3.1-2: Cont

| 1978 |  |  |  | 1979 |  |  |  | 1972-1979 Total |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| White Perch | 113,164 | 316 | 41\% | Gizzard Shad | 75,553 | 251 | 38\% | White Perch | 3,056,424 | 534 | 44\% |
| Gizzard Shad | 55,104 | 154 | 20\% | White Perch | 43,103 | 143 | 22\% | Gizzard Shad | 1,584,399 | 277 | 23\% |
| Channel Catfish | 48,575 | 136 | 18\% | Channel Catfish | 38,251 | 127 | 19\% | Blueback Herring | 873,833 | 153 | 13\% |
| Blueback Herring | 13,098 | 37 | 5\% | Common Carp | 14,946 | 50 | 8\% | Channel Catfish | 484,607 | 85 | 7\% |
| Common Carp | 11,842 | 33 | 4\% | Quillback | 5,134 | 17 | 3\% | American Eel | 241,657 | 42 | 4\% |
| Spottail Shiner | 8,506 | 24 | 3\% | Redbreast Sunfish | 3,466 | 12 | 2\% | Alewife | 176,495 | 31 | 3\% |
| American Eel | 5,878 | 16 | 2\% | Walleye | 2,491 | 8 | 1\% | Common Carp | 120,028 | 21 | 2\% |
| Redbreast Sunfish | 4,187 | 12 | 2\% | Blueback Herring | 2,282 | 8 | 1\% | Quillback | 81,963 | 14 | 1\% |
| Spotfin Shiner | 3,751 | 10 | 1\% | Shorthead Redhorse | 2,163 | 7 | 1\% | Spotfin Shiner | 62,016 | 11 | 1\% |
| Quillback | 2,361 | 7 | 1\% | Comely Shiner | 1,707 | 6 | 1\% | Redbreast Sunfish | 26,903 | 5 | 0.4\% |


| 1980 |  |  |  | 1981 |  |  |  | 1982 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 275,736 | 684 | 74\% | Gizzard Shad | 1,156,662 | 2,361 | 85\% | Gizzard Shad | 1,226,374 | 1,692 | 61\% |
| Channel Catfish | 38,929 | 97 | 10\% | White Perch | 83,363 | 170 | 6\% | White Perch | 53,527 | 74 | 3\% |
| White Perch | 26,971 | 67 | 7\% | Channel Catfish | 55,528 | 113 | 4\% | Channel Catfish | 40,941 | 56 | 2\% |
| Common Carp | 8,879 | 22 | 2\% | Common Carp | 18,313 | 37 | 1\% | Blueback Herring | 25,249 | 35 | 1\% |
| Walleye | 4,153 | 10 | 1\% | American Eel | 11,329 | 23 | 1\% | Common Carp | 15,362 | 21 | 1\% |
| Quillback | 2,929 | 7 | 1\% | Shorthead Redhorse | 6,533 | 13 | 1\% | Comely Shiner | 14,214 | 20 | 1\% |
| Striped Bass x White Bass | 2,674 | 7 | 1\% | Quillback | 3,622 | 7 | 0.3\% | Shorthead Redhorse | 6,974 | 10 | 0.3\% |
| Redbreast Sunfish | 1,524 | 4 | 0.4\% | Striped Bass | 3,277 | 7 | 0.2\% | American Eel | 3,961 | 5 | 0.2\% |
| Shorthead Redhorse | 1,394 | 4 | 0.4\% | Walleye | 2,645 | 5 | 0.2\% | Alewife | 3,433 | 5 | 0.2\% |
| White Sucker | 1,145 | 3 | 0.3\% | White Catfish | 2,199 | 4 | 0.2\% | American Shad | 2,039 | 3 | 0.1\% |

Table 5.3.1-2: Cont

| 1983 |  |  |  | 1984 |  |  |  | 1985 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 950,252 | 1,466 | 92.4\% | Gizzard Shad | 912,666 | 1,759 | 95\% | Gizzard Shad | 2,182,888 | 1,952 | 94\% |
| White Perch | 23,151 | 36 | 2.3\% | Channel Catfish | 20,479 | 39 | 2\% | White Perch | 68,344 | 61 | 3\% |
| Common Carp | 16,273 | 25 | 1.6\% | Common Carp | 8,012 | 15 | 1\% | Channel Catfish | 15,200 | 14 | 1\% |
| Channel Catfish | 12,559 | 19 | 1.2\% | White Perch | 6,402 | 12 | 1\% | Blueback Herring | 6,763 | 6 | 0.3\% |
| Shorthead Redhorse | 7,558 | 12 | 0.7\% | Shorthead Redhorse | 3,467 | 7 | 0.4\% | Common Carp | 6,729 | 6 | 0.3\% |
| Quillback | 4,679 | 7 | 0.5\% | Quillback | 1,942 | 4 | 0.2\% | Bluegill | 6,048 | 5 | 0.3\% |
| Comely Shiner | 3,176 | 5 | 0.3\% | Comely Shiner | 871 | 2 | 0.1\% | Comely Shiner | 5,141 | 5 | 0.2\% |
| Spottail Shiner | 2,132 | 3 | 0.2\% | Smallmouth Bass | 608 | 1 | 0.1\% | Spottail Shiner | 3,525 | 3 | 0.2\% |
| American Eel | 1,080 | 2 | 0.1\% | Yellow Perch | 487 | 1 | 0.1\% | Redbreast Sunfish | 3,366 | 3 | 0.1\% |
| Smallmouth Bass | 1,003 | 2 | 0.1\% | Redbreast Sunfish | 465 | 1 | 0.0\% | Shorthead Redhorse | 3,362 | 3 | 0.1\% |


| 1986 |  |  |  | 1987 |  |  |  | 1988 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 1,714,441 | 2,063 | 94\% | Gizzard Shad | 2,488,618 | 1,760 | 96\% | Gizzard Shad | 1,402,565 | 1,055 | 88\% |
| White Perch | 56,977 | 69 | 3\% | White Perch | 29,995 | 21 | 1\% | White Perch | 90,651 | 68 | 6\% |
| Channel Catfish | 18,898 | 23 | 1\% | Comely Shiner | 21,199 | 15 | 1\% | Channel Catfish | 36,212 | 27 | 2\% |
| Blueback Herring | 6,327 | 8 | 0.3\% | Channel Catfish | 11,699 | 8 | 0.5\% | Blueback Herring | 14,570 | 11 | 1\% |
| Spottail Shiner | 6,247 | 8 | 0.3\% | American Shad | 7,667 | 5 | 0.3\% | Comely Shiner | 11,734 | 9 | 1\% |
| American Shad | 5,195 | 6 | 0.3\% | Striped Bass x White Bass | 5,895 | 4 | 0.2\% | Common Carp | 8,535 | 6 | 1\% |
| Common Carp | 2,930 | 4 | 0.2\% | Blueback Herring | 5,861 | 4 | 0.2\% | $\begin{aligned} & \hline \text { Striped Bass x White } \\ & \text { Bass } \\ & \hline \end{aligned}$ | 6,203 | 5 | 0.4\% |
| Alewife | 2,822 | 3 | 0.2\% | Common Carp | 4,607 | 3 | 0.2\% | American Shad | 5,146 | 4 | 0.3\% |
| Quillback | 2,327 | 3 | 0.1\% | Shorthead Redhorse | 3,583 | 3 | 0.1\% | Shorthead Redhorse | 4,782 | 4 | 0.3\% |
| Yellow Perch | 2,267 | 3 | 0.1\% | Bluegill | 2,436 | 2 | 0.1\% | White Catfish | 3,849 | 3 | 0.2\% |

Table 5.3.1-2: Cont

| $\mathbf{1 9 8 9}$ |  |  |  | 1980-1989 Total |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 926,213 | 829 | $89 \%$ | Gizzard Shad | $13,236,415$ | 1,540 | $91 \%$ |
| Comely Shiner | 35,239 | 32 | $3 \%$ | White Perch | 455,094 | 53 | $3 \%$ |
| Channel Catfish | 21,692 | 19 | $2 \%$ | Channel Catfish | 272,137 | 32 | $2 \%$ |
| White Perch | 15,713 | 14 | $2 \%$ | Comely Shiner | 93,198 | 11 | $1 \%$ |
| American Shad | 8,218 | 7 | $1 \%$ | Common Carp | 90,515 | 11 | $1 \%$ |
| Spotfin Shiner | 5,381 | 5 | $1 \%$ | Blueback Herring | 64,316 | 7 | $0.4 \%$ |
| Striped Bass x White <br> Bass | 5,243 | 5 | $1 \%$ | Shorthead Redhorse | 42,445 | 5 | $0.3 \%$ |
| Blueback Herring | 3,598 | 3 | $0.3 \%$ | American Shad | 30,858 | 4 | $0.2 \%$ |
| Shorthead Redhorse | 2,735 | 2 | $0.3 \%$ | Striped Bass x White <br> Bass | 23,941 | 3 | $0.2 \%$ |
| Alewife | 1,902 | 2 | $0.2 \%$ | Quillback | 21,702 | 3 | $0.1 \%$ |


| 1990 |  |  |  | 1991 |  |  |  | 1992 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 1,084,073 | 795 | 93\% | Gizzard Shad | 433,108 | 345 | 81\% | Gizzard Shad | 1,450,299 | 930 | 93\% |
| White Perch | 24,581 | 18 | 2\% | Comely Shiner | 18,356 | 15 | 3\% | White Perch | 37,521 | 24 | 2\% |
| American Shad | 15,719 | 12 | 1\% | Blueback Herring | 15,616 | 12 | 3\% | Blueback Herring | 27,533 | 18 | 2\% |
| Blueback Herring | 9,658 | 7 | 1\% | White Perch | 14,996 | 12 | 3\% | American Shad | 10,335 | 7 | 1\% |
| Channel Catfish | 8,689 | 6 | 1\% | American Shad | 13,330 | 11 | 3\% | Comely Shiner | 8,974 | 6 | 1\% |
| Comely Shiner | 5,798 | 4 | 1\% | Channel Catfish | 10,252 | 8 | 2\% | Channel Catfish | 7,070 | 5 | 1\% |
| Shorthead Redhorse | 4,228 | 3 | 0.4\% | Common Carp | 8,257 | 7 | 2\% | Common Carp | 4,105 | 3 | 0.3\% |
| Common Carp | 2,761 | 2 | 0.2\% | Quillback | 2,990 | 2 | 1\% | Alewife | 3,344 | 2 | 0.2\% |
| Quillback | 1,270 | 1 | 0.1\% | Shorthead Redhorse | 2,871 | 2 | 1\% | American Eel | 2,622 | 2 | 0.2\% |
| Striped Bass x White Bass | 1,172 | 1 | 0.1\% | Alewife | 2,649 | 2 | 1\% | Striped Bass | 2,094 | 1 | 0.1\% |

Table 5.3.1-2: Cont

| 1993 |  |  |  | 1994 |  |  |  | 1995 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 666,010 | 645 | 93\% | Gizzard Shad | 511,139 | 530 | 91\% | Gizzard Shad | 799,694 | 642 | 80\% |
| Channel Catfish | 10,841 | 11 | 2\% | Comely Shiner | 13,973 | 15 | 2\% | Blueback Herring | 93,859 | 75 | 9\% |
| Common Carp | 8,488 | 8 | 1\% | White Perch | 9,537 | 10 | 2\% | White Perch | 55,719 | 45 | 6\% |
| Comely Shiner | 7,358 | 7 | 1\% | Common Carp | 7,403 | 8 | 1\% | American Shad | 15,588 | 13 | 2\% |
| American Shad | 5,343 | 5 | 1\% | American Shad | 5,615 | 6 | 1\% | Common Carp | 6,209 | 5 | 1\% |
| Blueback Herring | 4,052 | 4 | 1\% | Striped Bass | 4,261 | 4 | 1\% | Striped Bass | 5,467 | 4 | 1\% |
| White Perch | 3,892 | 4 | 1\% | Channel Catfish | 3,551 | 4 | 1\% | Alewife | 5,405 | 4 | 1\% |
| Striped Bass | 1,595 | 2 | 0\% | Blueback Herring | 2,603 | 3 | 0.5\% | Channel Catfish | 2,432 | 2 | 0.2\% |
| American Eel | 1,487 | 1 | 0\% | Shorthead Redhorse | 1,994 | 2 | 0.4\% | Shorthead Redhorse | 2,098 | 2 | 0.2\% |
| Shorthead Redhorse | 858 | 1 | 0\% | Quillback | 1,576 | 2 | 0.3\% | Comely Shiner | 1,746 | 1 | 0.2\% |


| 1996 |  |  |  | 1997 |  |  |  | 1998 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 196,019 | 422 | 84\% | Blueback Herring | 133,257 | 218 | 39\% | Gizzard Shad | 497,375 | 1,045 | 86\% |
| American Shad | 11,473 | 25 | 5\% | Gizzard Shad | 126,570 | 207 | 37\% | White Perch | 32,891 | 69 | 6\% |
| Common Carp | 5,726 | 12 | 2\% | White Perch | 58,685 | 96 | 17\% | Channel Catfish | 17,250 | 36 | 3\% |
| Channel Catfish | 5,487 | 12 | 2\% | American Shad | 12,974 | 21 | 4\% | Common Carp | 8,206 | 17 | 1\% |
| White Perch | 4,583 | 10 | 2\% | Shorthead Redhorse | 3,134 | 5 | 1\% | American Shad | 6,577 | 14 | 1\% |
| Comely Shiner | 2,180 | 5 | 1\% | Striped Bass | 2,665 | 4 | 1\% | Blueback Herring | 5,511 | 12 | 1\% |
| Striped Bass | 1,845 | 4 | 1\% | Common Carp | 2,281 | 4 | 1\% | Striped Bass | 2,570 | 5 | 0.4\% |
| Walleye | 964 | 2 | 0.4\% | Walleye | 1,063 | 2 | 0.3\% | Walleye | 827 | 2 | 0.1\% |
| Blueback Herring | 871 | 2 | 0.4\% | Spottail Shiner | 1,041 | 2 | 0.3\% | Smallmouth Bass | 812 | 2 | 0.1\% |
| Shorthead Redhorse | 754 | 2 | 0.3\% | Channel Catfish | 977 | 2 | 0.3\% | Comely Shiner | 570 | 1 | 0.1\% |

Table 5.3.1-2: Cont

| $\mathbf{1 9 9 9}$ |  |  |  | 1990-1999 Total |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 652,770 | 921 | $90 \%$ | Gizzard Shad | $6,417,057$ | 663 | $87 \%$ |
| White Perch | 35,357 | 50 | $5 \%$ | Blueback Herring | 301,506 | 31 | $4 \%$ |
| American Shad | 9,658 | 14 | $1 \%$ | White Perch | 277,762 | 29 | $4 \%$ |
| Blueback Herring | 8,546 | 12 | $1 \%$ | American Shad | 106,612 | 11 | $1 \%$ |
| Common Carp | 5,124 | 7 | $1 \%$ | Channel Catfish | 69,113 | 7 | $1 \%$ |
| Channel Catfish | 2,564 | 4 | $0.4 \%$ | Comely Shiner | 60,046 | 6 | $1 \%$ |
| Alewife | 1,795 | 3 | $0.2 \%$ | Common Carp | 58,560 | 6 | $1 \%$ |
| Shorthead Redhorse | 1,485 | 2 | $0.2 \%$ | Striped Bass | 24,248 | 3 | $0.3 \%$ |
| Smallmouth Bass | 1,306 | 2 | $0.2 \%$ | Shorthead Redhorse | 19,592 | 2 | $0.3 \%$ |
| Striped Bass | 1,001 | 1 | $0.1 \%$ | Alewife | 14,303 | 1 | $0.2 \%$ |


| 2000 |  |  |  | 2001 |  |  |  | 2002 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 366,099 | 863 | 80\% | Gizzard Shad | 218,124 | 513 | 70\% | Gizzard Shad | 339,292 | 814 | 81\% |
| White Perch | 40,318 | 95 | 9\% | White Perch | 44,364 | 104 | 14\% | White Perch | 65,031 | 156 | 16\% |
| Blueback Herring | 14,326 | 34 | 3\% | Blueback Herring | 16,320 | 38 | 5\% | American Shad | 9,347 | 22 | 2\% |
| American Shad | 9,785 | 23 | 2\% | American Shad | 10,940 | 26 | 4\% | Striped Bass | 2,086 | 5 | 1\% |
| Alewife | 9,189 | 22 | 2\% | Alewife | 7,824 | 18 | 3\% | Channel Catfish | 844 | 2 | 0.2\% |
| Channel Catfish | 8,394 | 20 | 2\% | Spottail Shiner | 5,833 | 14 | 2\% | Blueback Herring | 428 | 1 | 0.1\% |
| Common Carp | 3,236 | 8 | 1\% | Comely Shiner | 1,228 | 3 | 0.4\% | Smallmouth Bass | 390 | 1 | 0.1\% |
| Striped Bass | 2,453 | 6 | 1\% | Common Carp | 994 | 2 | 0.3\% | Shorthead Redhorse | 317 | 1 | 0.1\% |
| Shorthead Redhorse | 1,317 | 3 | 0.3\% | Redbreast Sunfish | 783 | 2 | 0.3\% | Common Carp | 225 | 1 | 0.1\% |
| Smallmouth Bass | 764 | 2 | 0.2\% | Striped Bass | 710 | 2 | 0.2\% | Redbreast Sunfish | 179 | 0.4 | 0.04\% |

Table 5.3.1-2: Cont.

| 2003 |  |  |  | 2004 |  |  |  | 2005 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 118,852 | 324 | 81\% | Gizzard Shad | 22,899 | 152 | 61\% | Gizzard Shad | 82,412 | 279 | 87\% |
| White Perch | 14,476 | 39 | 10\% | Channel Catfish | 4,839 | 32 | 13\% | American Shad | 3,896 | 13 | 4\% |
| American Shad | 9,802 | 27 | 7\% | American Shad | 3,426 | 23 | 9\% | Channel Catfish | 1,692 | 6 | 2\% |
| Common Carp | 1,110 | 3 | 1\% | Common Carp | 2,702 | 18 | 7\% | Common Carp | 1,179 | 4 | 1\% |
| Shorthead Redhorse | 749 | 2 | 1\% | Brown Bullhead | 1,599 | 11 | 4\% | White Perch | 1,102 | 4 | 1\% |
| Striped Bass | 703 | 2 | 0.5\% | White Perch | 976 | 6 | 3\% | Shorthead Redhorse | 863 | 3 | 1\% |
| Channel Catfish | 626 | 2 | 0.4\% | Striped Bass | 458 | 3 | 1\% | Quillback | 848 | 3 | 1\% |
| Smallmouth Bass | 232 | 1 | 0.2\% | White Catfish | 271 | 2 | 1\% | Brown Bullhead | 713 | 2 | 1\% |
| Blueback Herring | 183 | 0.5 | 0.1\% | Redbreast Sunfish | 70 | 0.5 | 0.2\% | Smallmouth Bass | 560 | 2 | 1\% |
| Brown Bullhead | 104 | 0.3 | 0.1\% | Comely Shiner | 67 | 0.4 | 0.2\% | Striped Bass | 489 | 2 | 1\% |


| 2006 |  |  |  | 2007 |  |  |  | 2008 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 149,250 | 428 | 91\% | Gizzard Shad | 146,821 | 510 | 92\% | Gizzard Shad | 724,737 | 1,507 | 99\% |
| American Shad | 3,970 | 11 | 2\% | American Shad | 4,272 | 15 | 3\% | American Shad | 2,627 | 6 | 0.4\% |
| Channel Catfish | 2,880 | 8 | 2\% | White Perch | 2,276 | 8 | 1\% | White Perch | 2,036 | 4 | 0.3\% |
| Walleye | 1,962 | 6 | 1\% | Walleye | 1,776 | 6 | 1\% | Walleye | 1,971 | 4 | 0.3\% |
| Brown Bullhead | 1,060 | 3 | 1\% | Channel Catfish | 1,480 | 5 | 1\% | Channel Catfish | 781 | 2 | 0.1\% |
| White Perch | 1,001 | 3 | 1\% | Spottail Shiner | 986 | 3 | 1\% | Common Carp | 400 | 1 | 0.1\% |
| Common Carp | 716 | 2 | 0.4\% | Common Carp | 372 | 1 | 0.2\% | Shorthead Redhorse | 325 | 1 | 0.04\% |
| Comely Shiner | 548 | 2 | 0.3\% | Striped Bass | 263 | 1 | 0.2\% | Smallmouth Bass | 95 | 0.2 | 0.01\% |
| Striped Bass | 383 | 1 | 0.2\% | Brown Bullhead | 237 | 1 | 0.1\% | Spotfin Shiner | 83 | 0.2 | 0.01\% |
| Smallmouth Bass | 306 | 1 | 0.2\% | Blueback Herring | 153 | 1 | 0.1\% | Spottail Shiner | 76 | 0.2 | 0.01\% |

Table 5.3.1-2: Cont.

| $2009$ |  |  |  | 2000-2009 Total |  |  |  | 1972-2009 Total |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | CPUE | \% | Species | No. | CPUE | \% | Species | No. | CPUE | \% |
| Gizzard Shad | 210,633 | 747 | 93\% | Gizzard Shad | 2,379,119 | 684 | 87\% | Gizzard Shad | 23,616,990 | 859 | 75\% |
| American Shad | 6,534 | 23 | 3\% | White Perch | 174,675 | 50 | 6\% | White Perch | 3,963,955 | 144 | 13\% |
| White Perch | 3,095 | 11 | 1\% | American Shad | 64,599 | 19 | 2\% | Blueback Herring | 1,271,244 | 46 | 4\% |
| Channel Catfish | 2,393 | 8 | $1 \%$ | Blueback Herring | 31,589 | 9 | $1 \%$ | Channel Catfish | 850,014 | 31 | 3\% |
| Walleye | 977 | 3 | $0.4 \%$ | Channel Catfish | 24,157 | 7 | $1 \%$ | Common Carp | 280,436 | 10 | 1\% |
| Common Carp | 399 | 1 | $0.2 \%$ | Alewife | 17,201 | 5 | 1\% | American Eel | 268,870 | 10 | 1\% |
| Bluegill | 313 | 1 | 0.1\% | Common Carp | 11,333 | 3 | 0.4\% | Alewife | 217,753 | 8 | 1\% |
| Brown Bullhead | 198 | 1 | 0.1\% | Striped Bass | 7,766 | 2 | 0.3\% | American Shad | 202,875 | 7 | 1\% |
| Flathead Catfish | 196 | 1 | 0.1\% | Walleye | 7,558 | 2 | 0.3\% | Comely Shiner | 165,985 | 6 | 1\% |
| Striped Bass | 179 | 1 | 0.1\% | Spottail Shiner | 6,920 | 2 | 0.3\% | Quillback | 115,476 | 4 | 0.4\% |

TABLE 5.3.2-1: TOTALS OF ANNUAL CATCH OF FISHES AT THE CONOWINGO DAM, EAST FISH LIFT 1991-2009.

| Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| No. Days | 60 | 49 | 42 | 55 | 68 | 49 | 64 |
| Lifts | 1,168 | 599 | 848 | 955 | 986 | 599 | 652 |
| Est. Oper. <br> Time(HR) | 647.2 | 454.1 | 463.5 | 574.8 | 706.2 | 454.1 | 640 |
| \#Taxa | 42 | 45 | 29 | 36 | 36 | 35 | 36 |
| Total | 650,940 | $2,394,583$ | 529,594 | $1,062,634$ | $1,796,460$ | 492,384 | 719,297 |
| Fish per Lift | 557 | 3,998 | 624 | 1,113 | 1,822 | 822 | 1,103 |


| Year | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| No. Days | 50 | 53 | 45 | 43 | 51 | 44 | 44 |
| Lifts | 652 | 610 | 570 | 559 | 560 | 645 | 590 |
| Est. Oper. <br> Time(HR) | 640 | 467 | 367.8 | 359.8 | 440.7 | 416.6 | 390.3 |
| \#Taxa | 33 | 31 | 31 | 30 | 31 | 25 | 31 |
| Total | 712,993 | $1,184,101$ | 493,953 | 921,916 | 656,894 | 589,177 | 715,664 |
| Fish per Lift | 1,094 | 1,941 | 86 | 1,649 | 1,173 | 914 | 1,213 |


| Year |  |  |  |  | Total |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| No. Days | 2005 | 2006 | 2007 | 2008 | 2009 | $1991-2009$ |
| Lifts | 52 | 641 | 619 | 39 | 51 | 57 |
| Est. Oper. Time(HR) | 434.3 | 429.8 | 379 | 483 | 618 | 12,733 |
| \#Taxa | 28 | 32 | 312 | 409 | 495.6 | 9126 |
| Total | 377,762 | 714,918 | 539,203 | 943,838 | 915,417 | $16,411,728$ |
| Fish per Lift | 698 | 1,155 | 1,126 | 1,954 | 1,481 | 1,289 |

TABLE 5.3.2-2: TOP TEN TOTALS OF ANNUAL CATCH OF FISHES BY YEAR, CPUE (FISH PER LIFT) AND PROPORTIONAL ABUNDANCE AT THE CONOWINGO DAM EAST FISH LIFT:YEARS 1991-2009.

| 1991 |  |  |  | 1992 |  |  |  | 1993 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 575,505 | 493 | 88\% | Gizzard Shad | 2,351,351 | 3,925 | 98\% | Gizzard Shad | 504,116 | 594 | 95\% |
| Common Carp | 23,833 | 20 | 4\% | American Shad | 15,386 | 26 | 1\% | American Shad | 8,203 | 10 | 2\% |
| American Shad | 13,897 | 12 | 2\% | White Perch | 8,725 | 15 | 0.4\% | Common Carp | 6,649 | 8 | 1\% |
| Blueback Herring | 13,149 | 11 | 2\% | Blueback Herring | 7,347 | 12 | 0.3\% | Blueback Herring | 4,574 | 5 | 1\% |
| Comely Shiner | 11,847 | 10 | 2\% | Common Carp | 6,072 | 10 | 0.3\% | Comely Shiner | 3,563 | 4 | 1\% |
| Quillback | 3,220 | 3 | 0.5\% | Channel Catfish | 1,124 | 2 | 0.05\% | Quillback | 540 | 1 | 0.1\% |
| Spotfin Shiner | 2,647 | 2 | 0.4\% | Comely Shiner | 650 | 1 | 0.03\% | Channel Catfish | 534 | 1 | 0.1\% |
| White Perch | 2,610 | 2 | 0.4\% | Carps and Minnows | 554 | 1 | 0.02\% | Striped Bass | 327 | 0.4 | 0.1\% |
| Striped Bass x White Bass | 827 | 1 | 0.1\% | Smallmouth Bass | 494 | 1 | 0.02\% | White Perch | 215 | 0.3 | 0.04\% |
| Smallmouth Bass | 671 | 1 | 0.1\% | Quillback | 483 | 1 | 0.02\% | Smallmouth Bass | 185 | 0.2 | 0.03\% |


| 1994 |  |  |  | 1995 |  |  |  | 1996 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 1,025,418 | 1,074 | 96\% | Gizzard Shad | 1,737,685 | 1,762 | 97\% | Gizzard Shad | 455,317 | 760 | 92\% |
| American Shad | 26,715 | 28 | 3\% | American Shad | 46,062 | 47 | 3\% | American Shad | 26,040 | 43 | 5\% |
| Common Carp | 5,042 | 5 | 0.5\% | Blueback Herring | 4,004 | 4 | 0.2\% | Common Carp | 4,139 | 7 | 1\% |
| Quillback | 2,507 | 3 | 0.2\% | Common Carp | 3,262 | 3 | 0.2\% | Quillback | 3,773 | 6 | 1\% |
| Channel Catfish | 544 | 1 | 0.1\% | Quillback | 2,910 | 3 | 0.2\% | Channel Catfish | 1,037 | 2 | 0.2\% |
| Striped Bass | 506 | 1 | 0.05\% | White Perch | 528 | 1 | 0.03\% | Smallmouth Bass | 531 | 1 | 0.1\% |
| Comely Shiner | 433 | 0.5 | 0.04\% | Striped Bass | 505 | 1 | 0.03\% | Walleye | 351 | 1 | 0.1\% |
| Walleye | 255 | 0.3 | 0.02\% | Walleye | 271 | 0.3 | 0.02\% | Striped Bass | 276 | 1 | 0.1\% |
| Blueback Herring | 248 | 0.3 | 0.02\% | Redbreast Sunfish | 185 | 0.2 | 0.01\% | Blueback Herring | 261 | 0.4 | 0.1\% |
| Shorthead Redhorse | 242 | 0.3 | 0.02\% | Alewife | 170 | 0.2 | 0.01\% | Shorthead Redhorse | 228 | 0.4 | 0.05\% |

Table 5.3.2-2: Cont.

| 1997 |  |  |  | 1998 |  |  |  | 1999 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 344,332 | 528 | 48\% | Gizzard Shad | 654,575 | 1,004 | 92\% | Gizzard Shad | 950,500 | 1,558 | 80\% |
| Blueback Herring | 242,815 | 372 | 34\% | American Shad | 39,904 | 61 | 6\% | Blueback Herring | 130,625 | 214 | 11\% |
| American Shad | 90,971 | 140 | 13\% | Common Carp | 6,205 | 10 | 1\% | American Shad | 697,12 | 114 | 6\% |
| White Perch | 27,312 | 42 | 4\% | Channel Catfish | 4,135 | 6 | 1\% | White Perch | 27,133 | 44 | 2\% |
| Common Carp | 3,256 | 5 | 0.5\% | White Perch | 2,731 | 4 | 0.4\% | Common Carp | 2,430 | 4 | 0.2\% |
| Quillback | 2,488 | 4 | 0.3\% | Striped Bass | 1,467 | 2 | 0.2\% | Striped Bass | 1,231 | 2 | 0.1\% |
| Walleye | 2,334 | 4 | 0.3\% | Shorthead Redhorse | 885 | 1 | 0.1\% | Smallmouth Bass | 797 | 1 | 0.1\% |
| Shorthead Redhorse | 1,475 | 2 | 0.2\% | Blueback Herring | 700 | 1 | 0.1\% | Walleye | 421 | 1 | 0.04\% |
| Channel Catfish | 1,178 | 2 | 0.2\% | Walleye | 685 | 1 | 0.1\% | Channel Catfish | 266 | 0.4 | 0.02\% |
| Striped Bass | 1,015 | 2 | 0.1\% | Smallmouth Bass | 508 | 1 | 0.1\% | Shorthead Redhorse | 245 | 0.4 | 0.02\% |


| 2000 |  |  |  | 2001 |  |  |  | 2002 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 317,753 | 557 | 64\% | Gizzard Shad | 429,461 | 768 | 47\% | Gizzard Shad | 513,794 | 917 | 78\% |
| American Shad | 153,546 | 269 | 31\% | Blueback Herring | 284,921 | 510 | 31\% | American Shad | 108,001 | 193 | 16\% |
| Blueback Herring | 14,963 | 26 | 3\% | American Shad | 193,574 | 346 | 21\% | White Perch | 29,404 | 53 | 4\% |
| White Perch | 4,387 | 8 | 1\% | Alewife | 7,458 | 13 | 1\% | Blueback Herring | 2,037 | 4 | 0.3\% |
| Striped Bass | 802 | 1 | 0.2\% | White Perch | 2,659 | 5 | 0.3\% | Striped Bass | 913 | 2 | 0.1\% |
| Channel Catfish | 677 | 1 | 0.1\% | Common Carp | 1,267 | 2 | 0.1\% | Smallmouth Bass | 597 | 1 | 0.1\% |
| Smallmouth Bass | 427 | 1 | 0.1\% | Striped Bass | 543 | 1 | 0.1\% | Quillback | 400 | 1 | 0.1\% |
| Quillback | 408 | 1 | 0.1\% | Smallmouth Bass | 404 | 1 | 0.04\% | Sea Lamprey | 316 | 1 | 0.05\% |
| Common Carp | 388 | 1 | 0.1\% | Shorthead Redhorse | 382 | 1 | 0.04\% | Shorthead Redhorse | 292 | 1 | 0.04\% |
| Walleye | 177 | 0.3 | 0.04\% | Spottail Shiner | 318 | 1 | 0.03\% | Yellow Perch | 258 | 0.5 | 0.04\% |

Table 5.3.2-2: Cont.

| 2003 |  |  |  | 2004 |  |  |  | 2005 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 459,634 | 713 | 78\% | Gizzard Shad | 602,677 | 1,021 | 84\% | Gizzard Shad | 305,378 | 564 | 80.8\% |
| American Shad | 125,135 | 194 | 21\% | American Shad | 109,360 | 185 | 15\% | American Shad | 68,926 | 127 | 18.2\% |
| White Perch | 1,572 | 2 | 0.3\% | Channel Catfish | 928 | 2 | 0.1\% | Quillback | 2,145 | 4 | 0.6\% |
| Common Carp | 561 | 1 | 0.1\% | White Perch | 512 | 1 | 0.1\% | Common Carp | 540 | 1 | 0.1\% |
| Quillback | 548 | 1 | 0.1\% | Striped Bass | 391 | 1 | 0.1\% | Smallmouth Bass | 256 | 0.5 | 0.1\% |
| Blueback Herring | 530 | 1 | 0.1\% | Quillback | 308 | 1 | 0.04\% | Shorthead Redhorse | 131 | 0.2 | 0.03\% |
| Shorthead Redhorse | 304 | 0.5 | 0.1\% | Comely Shiner | 291 | 0.5 | 0.04\% | Striped Bass | 89 | 0.2 | 0.02\% |
| Striped Bass | 272 | 0.4 | 0.05\% | Common Carp | 257 | 0.4 | 0.04\% | Channel Catfish | 83 | 0.2 | 0.02\% |
| Smallmouth Bass | 247 | 0.4 | 0.04\% | Smallmouth Bass | 172 | 0.3 | 0.02\% | Walleye | 47 | 0.1 | 0.01\% |
| Sea Lamprey | 68 | 0.1 | 0.01\% | Brown Bullhead | 161 | 0.3 | 0.02\% | Sea Lamprey | 35 | 0.1 | 0.01\% |


| 2006 |  |  |  | 2007 |  |  |  | 2008 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gizzard Shad | 655,990 | 1,060 | 91\% | Gizzard Shad | 508,627 | 1,062 | 94\% | Gizzard Shad | 919,975 | 1,905 | 97\% |
| American Shad | 56,899 | 92 | 8\% | American Shad | 25,464 | 53 | 5\% | American Shad | 19,914 | 41 | 2\% |
| Walleye | 641 | 1 | 0.1\% | White Perch | 1,434 | 3 | 0.3\% | Walleye | 2,088 | 4 | 0.2\% |
| Quillback | 407 | 1 | 0.1\% | Quillback | 1,236 | 3 | 0.2\% | Channel Catfish | 496 | 1 | 0.1\% |
| White Perch | 277 | 0.4 | 0.04\% | Walleye | 695 | 1 | 0.1\% | Quillback | 400 | 1 | 0.04\% |
| Smallmouth Bass | 165 | 0.3 | 0.02\% | Blueback Herring | 460 | 1 | 0.1\% | White Perch | 388 | 1 | 0.04\% |
| Sea Lamprey | 128 | 0.2 | 0.02\% | Alewife | 429 | 1 | 0.1\% | Common Carp | 199 | 0.4 | 0.02\% |
| Common Carp | 108 | 0.2 | 0.02\% | Shorthead Redhorse | 173 | 0.4 | 0.03\% | Smallmouth Bass | 96 | 0.2 | 0.01\% |
| Channel Catfish | 75 | 0.1 | 0.01\% | Striped Bass | 127 | 0.3 | 0.02\% | Shorthead <br> Redhorse | 66 | 0.1 | 0.01\% |
| Striped Bass | 73 | 0.1 | 0.01\% | Smallmouth Bass | 123 | 0.3 | 0.02\% | Bluegill | 65 | 0.1 | 0.01\% |

Table 5.3.2-2: Cont.

| 2009 |  |  | $\mathbf{1 9 9 1 - 2 0 0 9}$ |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Gizzard Shad | 876,412 | 1,418 | $96 \%$ | Gizzard Shad | $14,188,500$ | 1,114 | $86 \%$ |
| American Shad | 29,272 | 47 | $3 \%$ | American Shad | $1,226,981$ | 96 | $7 \%$ |
| Channel Catfish | 4,201 | 7 | $0.5 \%$ | Blueback Herring | 706,811 | 56 | $4 \%$ |
| Walleye | 1,832 | 3 | $0.2 \%$ | White Perch | 110,923 | 9 | $1 \%$ |
| Quillback | 899 | 1 | $0.1 \%$ | Common Carp | 65,373 | 5 | $0.4 \%$ |
| Common Carp | 886 | 1 | $0.1 \%$ | Quillback | 23,275 | 2 | $0.1 \%$ |
| White Perch | 839 | 1 | $0.1 \%$ | Comely Shiner | 17,372 | 1 | $0.1 \%$ |
| Sea Lamprey | 190 | 0.3 | $0.02 \%$ | Channel Catfish | 16,082 | 1 | $0.1 \%$ |
| Alewife | 160 | 0.3 | $0.02 \%$ | Walleye | 10,747 | 1 | $0.1 \%$ |
| Brown Bullhead | 153 | 0.2 | $0.02 \%$ | Striped Bass | 9,420 | 1 | $0.1 \%$ |

FIGURE 5.1-1: WEST FISH LIFT SCHEMATIC DRAWING


## FIGURE 5.1-2: EAST FISH LIFT SCHEMATIC DRAWING

 NEW LIFT DESIGN AND CONSTRUCTION

FIGURE 5.3.1-1: WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES (FISH PER LIFT) 1972-2009.


FIGURE 5.3.2-1: EAST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES (FISH PER LIFT) $1991-2009$.


### 6.0 ICHTHYOPLANKTON, GILL NETS, AND ELECTROFISHING

Pursuant to a new FERC license issued to Philadelphia Electric Company (PECO) and Susquehanna Electric Company in 1980 for continued operation of the Conowingo Project (FERC No. 405), biologicalbased studies (by FERC-approved Study Plan) were planned for two successive 5-year periods (1982 to 1986; 1987 to 1991) to address minimum flow issues. The various elements of study continued through 1987. In 1988, PECO (now Exelon) reached a settlement agreement with the resource agencies and the studies were halted. Because the studies were terminated following the settlement agreement, most collected data were only tabulated, processed electronically and stored on PECO's mainframe system; these data were not analyzed or formally presented in any reports. The biological data stored on electronic media were subsequently lost. As a result, data presented here are drawn from three progress reports and other available hard-copy data stored by Normandeau.

As part of the biological based studies and in addition to the fisheries information obtained during annual operations (mostly in spring) at the Conowingo WFL, the seasonal occurrence, distribution, and relative abundance of young-of-the-year (YOY), juvenile and adult fishes below Conowingo Dam were characterized from extensive electrofishing and gill net collections at three primary locations: the Conowingo tailrace, the non-tidal riverine area (Lee's Ferry and The Pool) and the tidal zone (Figure 6.11). These collections also provided specimens for concurrent studies to determine fish food habits and age/growth analysis. Reproduction of fishes below Conowingo Dam was investigated by spring ichthyoplankton collections.

The combination of various gears analyzed in conjunction with the fish lift catches were complimentary for characterizing species composition and abundance, size and trophic structures and reproductive success in the fish assemblage in thoroughly examining the habiats below Conowingo. Each collection method and associated gear type has a finite efficacy in sampling targeted stratas of habitat and corresponding assembladges of fish. The success of the gear is affected by the working physical conditions found in the habitats targeted during sampling. Physical conditions are variable due to dynamic factors (e.g., temperature, flow, turbidity, dissolved oxygen, conductivity and are influenced by seasonality, climate conditions, hydro operations and other anthrogenic and natural factors.

### 6.1 Icthyoplankton Sample Methods

From 1982 to 1984 ichthyoplankton sampling was generally performed from late March to late June to characterize the use of the lower Susquehanna River by resident and anadromous fishes as a spawning and nursery area. In 1982 eight regular sampling transects or locations were established for the study;
three in the tidal area below Spencer Island and five in the riverine area above Robert Island (Figure 6.12). In addition to samples at these river stations, surface tows were made in Deer and Octoraro Creeks near their confluences with the Susquehanna.

The sampling was conducted weekly at all transect stations, but during the suspected peak of spawning, samples were generally collected twice per week at each station. Samples were collected with 0.5 m plankton net ( 0.5 mm mesh). Nets were towed for five minutes heading upstream at speeds sufficient to maintain the net no more than one foot below the surface. Bottom samples were collected at the midchannel stations on the three lowermost transects in 1982 and 1983 and at the two lowermost transects in 1984. The surface and bottom tows collected a sample of pelagic eggs and larvae suspended in the water column; demersal and adhesive eggs were collected by the sampling gear type less effectively.

In 1983, 33 additional samples were collected near Spencer Island targeting American shad, supplemented by 21 additional collections taken at other areas in tidal water where telemetered American shad frequented or were observed spawning.

Water volume filtered was measured with a General Oceanics (model 2030) digital flowmeter mounted in the center of each net mouth. The volume $\left(\mathrm{m}^{3}\right)$ of water filtered in each tow was calculated by the formula:

$$
\mathrm{V}=\mathrm{n}(\mathrm{f} / 100) \mathrm{A}
$$

Where "V" was the volume of water filtered, " $n$ " was the number of flowmeter revolutions, " f " was the factor for converting counts $/ \mathrm{sec}$ to velocity ( $\mathrm{cm} / \mathrm{sec}$ ); divided by 100 to convert $\mathrm{cm} / \mathrm{sec}$ to $\mathrm{m} / \mathrm{sec}$, and "A" was the cross sectional area of the net mouth $\left(0.196349 \mathrm{~m}^{2}\right)$.

Field samples were immediately preserved in 20 to $25 \%$ formalin, rinsed in the laboratory, stained with rose bengal (to facilitate sorting), sorted, and stored in vials in 5\% buffered formalin until identification. Specimens were examined under a stereoscopic dissecting microscope, identified to the lowest taxon, and enumerated.

Specimens were classified as eggs, larvae (prolarvae or postlarvae), or young. Damaged specimens too distorted to identify were tabulated as unidentifiable. Larvae of the genus Alosa, particularly Alosa aestivalis (blueback herring) and Alosa pseudoharengus (alewife) were indistinguishable to species and were termed (for the purposes of reporting) Alosa spp. The larval stage was defined as the early development after hatching during which the yolk sac and larval finfold were absorbed, and the fin rays were formed. The larval stage was subdivided into prolarvae and postlarvae. Prolarvae were those
specimens that had not completely absorbed their yolk sac. Postlarvae had absorbed the yolk sac but not completely differentiated to their adult form. Young were fully transformed larvae, characterized by complete absorption of the larval finfold and attainment of the adult compliment of rays and spines in all fins.

### 6.2 Electrofishing Sample Methods

Electrofishing surveys were conducted at least monthly (meteorological and hydrological conditions permitting) from 1981 to 1987 at multiple stations in three general areas below Conowingo Dam: the Conowingo tailrace, the riverine area (Lee's Ferry and The Pool) and the tidal zone (Figure 6.1-1). Flow conditions sampled included periods of interim minimum flow (i.e., summer) and zero generation (other months) at Conowingo Dam. These sampling areas all shared the characteristic of retaining deep water (at least 6.5 ft in some locations) even after 8 hours or more of reduced generation. Stations in these areas were selected so that sampling at night at reduced generation ( $<5,000 \mathrm{cfs}$ ) was feasible. In 1987 electrofishing was conducted below Conowingo Dam only in the tailrace and tidal zone. Fish were sampled at night with a boat-mounted electrofisher, oriented upstream, with pulsed DC current at 5-6 amps. Water temperature and dissolved oxygen (DO) were recorded at each location sampled.

Some of the factors commonly associated with parameterizing the limitations of electrofishing catch include: range of efficacy (distance at which fish are affected), water depth (difficulty capturing immobilized fish at depths beyond $0.9-1.2 \mathrm{~m}$ ) due to visibility and length of dipnet handle, size of investigator, water quality (best response depends on the species and water temperature, conductivity, etc.), and selectivity for species and size related swimming ability (Klemm et al. 1993).

Substrate and cover varied somewhat among the sampling locations. Tailrace stations had a rubble substrate punctuated with various size boulders, whereas tidal zone stations had primarily rubble substrate with frequent boulders and occasional stretches of sand or silt. Patches of dense submerged aquatic vegetation (primarily Myriophyllum) were present along most tidal zone stations. In the tailrace and tidal zone, all stations were bordered by an island or main channel shoreline. Stations in Lee's Ferry and The Pool were characterized by rubble and boulder substrate, a lack of vegetation, and interspersed areas of shallow runs, riffles, and pockets behind boulders. These stations were located along the "boundaries" of each pooled location.

Each station was sampled for approximately 30 minutes in a continuous run. Generally, the same four or five stations in each area were sampled each night. During a run, most stunned fishes were retrieved and placed in a flow-through live well. The numbers of carp and eels were visually estimated. At the
completion of a run, all captured fishes were identified, counted, measured, and released. A representative length frequency ( 10 mm FL groups) was obtained for those species most abundant in an area. Selected specimens were utilized for various life history studies, which included age and growth, food habits and movement (conventional anchor tags and radio telemetry). The raw data were adjusted for fishing effort to facilitate comparison or relative abundance among areas and months. CPUE data are expressed as number of specimens captured (or observed, in the case of carp and eels) per hour shocked.

### 6.3 Gill Net Sample Methods

Gill netting was conducted monthly from July through November in 1981 to 1984 and in the same general areas sampled by electrofishing. The experimental monofilament gill nets used were 125 ft long, 6 ft deep, and consisted of five 25 ft panels increasing in mesh size from 0.5 inch to 2.5 inch at 0.5 inch intervals. Generally, the size of the mesh determines the species and size of the fish to be caught in the gill net. Varying mesh sizes are used to increase the potential of obtaining samples of several year classes of a single species while also providing a greater potential to increase the number of species caught (Klemm et al. 1993). Conditions permitting, two to three nets were fished perpendicular to shore, usually from sunset to sunrise (average fishing time was 12 to 13 hours per set). Water temperature and DO were recorded at each location at the time nets were set and retrieved.

Gill nets were set in three general areas below Conowingo Dam: the Conowingo tailrace, the riverine area (Lee's Ferry and The Pool) and the tidal zone (Figure 6.1-1), to determine spatial and temporal differences in relative abundance of fishes. Substrate at all sampling sites consisted of gravel, rocks, and boulders. Water depth at net sites never exceeded eight $\mathrm{ft}(2.44 \mathrm{~m})$ during a fishing period. Captured fishes were identified, enumerated and measured, and selected fishes were processed for age/growth and food habitats analysis.

Relative abundance of fishes was determined by the number of each species, the percent composition and the catch per unit effort (no. fish/net-night, CPUE = no. fish captured/no. nets set). Abundance indices were considered by month and location. Mean lengths of the five most abundant species were analyzed for each location by combined monthly collections. Length ranges for each location were determined for combined monthly collections only.

### 6.4 Icthyoplankton Results

From 1982 to 1984275,710 eggs, prolarvae or postlarvae from 27 taxa were collected from 1,322 icthyoplankton samples (Table 6.4-1). The most frequently collected species was white perch, constituting $72 \%$ of all icthyoplankton collected. River herring (alewife and blueback herring) comprised
an additional $24 \%$ of the icthyoplankton collected (Table 6.4-1). Gizzard shad comprised $3 \%$ of the ichthyoplankton collected (Table 6.4-1). Other less commonly collected species included American shad and carp; these along with all other species accounted for less than $1 \%$ of the specimens collected (Table 6.4-1).

During each year eggs were the most abundant life stage collected, comprising $81 \%, 84 \%$ and $90 \%$ of the total, respectively (Tables 6.4-2, 6.4-3, and 6.4-4). Prolarvae and postlarvae were the second and third most frequently collected life stages while older icthyoplankters were relatively scarce. In 1982 prolarvae and postlarvae were collected in near equal amounts while in 1983 and 1984 more prolarvae was collected than postlarvae.

The numerical peak of ichthyoplankton collections generally occurred in mid May through early June (Figure 6.4-1). The overall number of icthyoplankton were strongly related to the large numbers of white perch eggs and river herring eggs collected during this period. For example, in 1984 eggs represented $91 \%$ of the 61,589 white perch specimens collected (Table 6.4-5) and $95 \%$ of the 26,677 river herring specimens collected (Table 6.4-6).

### 6.4.1 Alosa spp.

From 1982 to 1984, highest egg densities of river herrings occurred from late April to early June when river temperatures ranged from 45 to $80^{\circ} \mathrm{F}$ (RMC 1985b), indicating earlier spawning compared to other species. Egg densities were highest at sites near Conowingo Dam (Table 6.4-6). The increased abundance of prolarvae at stations further downstream suggested that recently hatched Alosa spp. were quickly transported downstream and/or Alosa spp. eggs hatch in the tidal area. The near absence of postlarvae and young suggests either most development takes place downstream of the study area in the Susquehanna Flats or upper Chesapeake Bay, or the rate of larval mortality is extremely high.

In 1983 Alosa spp. were collected throughout the study area from 22 April through 26 June (Appendix C1) at water temperatures of 45 to $80^{\circ} \mathrm{F}$ (RMC 1985b). The peak density occurred on 25 May at $66^{\circ} \mathrm{F}$ (Appendix C-1, RMC 1985b). Mean density by station was highest in the tailrace near Conowingo Dam with a secondary peak at the old bridge piers near the head of tide (Table 6.4-6).

In 1984 Alosa spp. were collected from 17 April through 28 June. Eggs were first collected on 17 April (Appendix C-1) at $47^{\circ} \mathrm{F}$ (RMC 1985c). Peak abundance occurred on 15 May (Appendix C-2). Alosa spp. were most abundant in the upper riverine, non-tidal reach, particularly in the tailrace (Table 6.4-6). As in 1982 and 1983, stations within the tidal area in 1984 had higher numbers of prolarvae and postlarvae than those stations located further upstream (Table 6.4-6).

### 6.4.2 American shad

In 1982 only 7 American shad eggs were collected (Table 6.4-2) and the highest density occurred at the end of May (RMC 1985a). In 1983 a total of 138 American shad eggs were collected at regular sampling stations (Table 6.4-3); nearly all were collected below the tide line. A single prolarvae was collected at the mouth of Deer Creek on 5 May at $60^{\circ} \mathrm{F}$.

Thirty three additional samples were collected near Spencer Island in 1983 yielding 145 additional eggs from 15 May through 20 June (Appendix C-3). RMC 1985b reported that densities ranged from 0 to 0.3 eggs $/ \mathrm{m}^{3}$ and were taken from 25 to 27 May and 19 to 20 June. Generally, egg densities were lower in May and increased in early June, remaining above $0.1 \mathrm{eggs} / \mathrm{m}^{3}$ through 12 to 13 June. The highest densities $\left(0.3 \mathrm{eggs} / \mathrm{m}^{3} \cdot\right)$ occurred on 5 to 6 June. Water temperature ranged from 59 to $73^{\circ} \mathrm{F}$ when eggs were collected and was near $66^{\circ} \mathrm{F}$ at the time of peak egg density. Samples were taken at station depths ranging from 2 to 6 ft depending upon tidal stage and/or river flow. Although American shad were observed spawning in the immediate vicinity of Spencer Island, no telemetered shad were located at the time samples were collected.

Also in spring 1983, a total of 21 collections was taken at other areas where telemetered American shad frequented or were observed spawning (Appendix C-4). RMC 1985b reported that eggs were taken from 20 May through 5 June. A total of 533 eggs was collected and densities ranged from 0 to $2.8 \mathrm{eggs} / \mathrm{m}^{3}$. Higher egg densities $\left(>0.5 \mathrm{eggs} / \mathrm{m}^{3}\right)$ were estimated on 20 to 21 May, 1 to 2 June and 5 to 6 June. Water temperature ranged from 61 to $66^{\circ} \mathrm{F}$. Eggs were taken at station water depths of 3 to 13 ft ; the highest egg densities $\left(>0.5 \mathrm{eggs} / \mathrm{m}^{3}\right.$ ) were observed where water depths ranged from 3 to 4 ft . Surface water velocities in areas of intense spawning activity ranged from 0.7 to $2.4 \mathrm{ft} / \mathrm{s}$. Bottom substrate was primarily gravel and cobble. High egg densities and observed spawning activity indicated that a primary shad spawning area was along the east and west shores of Spencer Island.

In 1984, 48 night collections on 15, 17, and 22 May and 5 June yielded 179 eggs (Table 6.4-4). RMC 1985c reported that a majority of eggs were collected on 5 June ( 170 eggs, Appendix C-2) and water temperature averaged $52^{\circ} \mathrm{F}$ on this date (RMC 1985c). Samples were taken at station depths of 6 to 12 ft depending upon tidal stage and/or river flow. Most eggs (91 eggs) were collected at Station 7009 (along the west shore of the Susquehanna River near Lapidum boat launch) (RMC 1985c).

### 6.4.3 Gizzard shad

From 1982 to 1984, the greater abundance of gizzard shad postlarvae relative to prolarvae and comparatively few eggs (Tables 6.4-2, 6.4-3, 6.4-4) suggested most successful spawning occurred upstream of the study area.

In 1982 there were two periods of higher gizzard shad egg density. The first occurred in late May when the river temperature was $70^{\circ} \mathrm{F}$. A second peak occurred late in June at river temperatures of 70 to $77^{\circ} \mathrm{F}$ (RMC 1985a). The total density of gizzard shad in 1982 was 0.19 eggs, prolarvae and postlarvae $/ \mathrm{m}^{3}$ (Table 6.4-2). The greater density of postlarvae compared to prolarvae ( 0.15 postlarvae $/ \mathrm{m}^{3}$ to 0.01 prolarvae $/ \mathrm{m}^{3}$, Table 6.4-2) suggested that most spawning occurred upstream of the dam.

In 1983 gizzard shad eggs were absent (Table 6.4-3), indicating little spawning occurred below Conowingo Dam. Gizzard shad prolarvae and postlarvae were first taken on 17 May when water temperature averaged $64.2^{\circ} \mathrm{F}$ (RMC 1985b). Highest density of prolarvae occurred on 30 May at $66^{\circ} \mathrm{F}$ (RMC 1985b). All prolarval gizzard shad collected in 1983 had already attained the one to three day old stage of development (Jones et al. 1978), indicating their source was outside the study area. Peak density of postlarvae occurred on 10 June at $71.6^{\circ} \mathrm{F}$ (RMC 1985b). As in 1982, the overall abundance of postlarvae relative to prolarvae ( 0.13 postlarvae $/ \mathrm{m}^{3}$ to 0.04 prolarvae $/ \mathrm{m}^{3}$, Table 6.4-3) and the decline in larval density with distance downstream from Conowingo Dam suggested most successful spawning occurred upstream of the study area.

In 1984, gizzard shad eggs were again nearly absent (Table 6.4-4). Gizzard shad prolarvae were first taken on 17 April; the water temperature averaged $50.9^{\circ} \mathrm{F}$. The highest combined egg, prolarvae and postlarvae density occurred on 20 June at $74.4^{\circ} \mathrm{F}$ (RMC 1985c). All prolarval gizzard shad taken in 1984 had attained the one to three day old stage of development (Jones et al. 1978), indicating their source was outside the study area. Peak density of postlarvae occurred on 10 June at $71.6^{\circ} \mathrm{F}$ ( ${ }^{\text {RMC 1985c }}$ ).

### 6.4.4 White Perch

White perch was the most frequently caught species in each sample year. In 1982 spawning activity for white perch was extensive, as indicated by relatively high egg density ( $1.0 \mathrm{eggs} / \mathrm{m}^{3}$, Table 6.4-2). High egg and prolarval densities were recorded from late April through the end of May when the river temperatures ranged from 54 to $79^{\circ} \mathrm{F}$ (RMC 1985a).

In 1983, white perch was the most abundant ichthyoplankton taxa and accounted for $80 \%$ of the eggs and $75 \%$ of the prolarvae collected (Table 6.4-3). Eggs were first collected on 5 April at $47^{\circ} \mathrm{F}$ (RMC 1985b).

Peak spawning activity and egg density occurred on 21 May ( 168.7 eggs and prolarvae $/ \mathrm{m}^{3}$ ) at $63^{\circ} \mathrm{F}$ ( RMC 1985b). Highest daily abundance of prolarvae occurred on 24 May ( 3.9 prolarvae $/ \mathrm{m}^{3}$ ). White perch eggs are highly demersal and adhesive and, therefore, were much more abundant in bottom collections. Egg density at the three tidal Stations combined was $0.4 \mathrm{eggs} / \mathrm{m}^{3}$ on the surface compared to $2.2 \mathrm{eggs} / \mathrm{m}^{3}$ on the bottom (RMC 1985b). Greatest spawning activity occurred near the tide line. Egg densities were greatest near Wood Island, followed by stations immediately downstream near Lapidum. Some spawning occurred throughout the non-tidal riverine area, but densities of eggs and prolarvae at the three Stations in and just below the Conowingo tailrace were lower (RMC 1985b).

In 1984, white perch accounted for $68 \%$ of the eggs and $90 \%$ of the prolarvae collected (Table 6.4-4). Eggs were first collected on 12 April at $47^{\circ} \mathrm{F}$ (RMC 1985c). Peak spawning activity and egg density occurred on 22 May at $62.9^{\circ} \mathrm{F}$ (RMC 1985c). The greatest spawning activity occurred in the upper tidal area as indicated by the large amount of eggs collected (Table 6.4-5). Egg densities were greatest near Wood Island followed by stations just above Deer Creek off of Lapidum. Similar to 1983, less spawning occurred further upstream throughout the non-tidal, riverine area as evidenced by the lower numbers of eggs and prolarvae.

### 6.4.5 Other Fishes

Eggs, prolarvae and postlarvae of other fishes generally accounted for $1 \%$ or less of the total catch. Collections of the early lifestages of carp, yellow perch, walleye, bluegill, pumpkinseed, white crappie, smallmouth bass and largemouth bass were relatively low (Appendix $\mathrm{C}-1 \& \mathrm{C}-2$ ). All of these species have demersal, adhesive or both adhesive and demersal egg types that are not as effectively collected by the gear. No redbreast sunfish were collected, a species which commonly occurs as a resident. No striped bass were collected despite a nomadic presence in the sample area. It's unlikely that any striped bass spawning occurs downstream of Conowingo Dam to the tidal zone. Striped bass eggs are buoyant, pelagic and remain suspended in the water column increasing the likelihood of being recruited to the gear if present. Very few hickory shad (Alosa mediocris) were collected in 1984, mostly in early May.

### 6.5 Electrofishing Results

Over the course of the 1982 to 1987 electrofishing series, 247 hours (Table 6.5-1) of effort were spent sampling the Conowingo tailrace, Lee's Ferry, The Pool and tidal zone areas. A collection was completed in every month of the year except February. More effort was allocated to the tidal zone ( 93 hr ) and Conowingo tailrace ( 92 hr ) than The Pool (31 hr) and Lee's Ferry (30.5 hr, Table 6.5-1). Overall, the
most effort was expended during the months of July (43 hr), August (49 hr) and September (40 hr, Table 6.5-2).

A total of 235,458 fish of 66 taxa (Table 6.5-3) were collected in the lower Susquehanna River electrofishing program from May 1982 to October 1987. Overall, the rate of collection (CPUE) from 1982 to 1987 was 953 fish/hr (Table 6.5-3). Yearly CPUE ranged from 487 fish/hr in 1986 to 2,295 fish/hr in 1985 (Table 6.5-4). Correspondingly, the most fish were collected in 1985 (89,633 fish) and the least in 1986 (24,653 fish, Table 6.5-5).

The Conowingo tailrace had the highest overall CPUEs of the four locations from 1982 to 1984 and in 1987; the tidal zone had the highest CPUEs in 1985 and 1986 (Table 6.5-4). From 1982 to 1987 the tidal zone had the highest combined overall CPUE at 1,170 fish $/ \mathrm{hr}$, followed by the Conowingo tailrace $(1,042$ fish/hr), Lee's Ferry ( 533 fish/hr) and The Pool (454 fish/hr, Table 6.5-5).

The highest monthly rate of collection occurred in November 1985 in the tidal zone, at a CPUE of 32,868 fish/hr (Table 6.5-6); this is the equivalent of collecting nine fish per second. Masses of juvenile gizzard shad were responsible for the high CPUE. Overall, May and November had the highest average CPUEs across all locations and years at 1,432 fish $/ \mathrm{hr}$ and 4,418 fish $/ \mathrm{hr}$, respectively (Table 6.5-6).

In the Conowingo tailrace electrofishing samples occurred from May through December 1982, January and May through November in 1983, January, March and June through December in 1984, June through December in 1985, March through December in 1986, and July through October in 1987 (Appendix D-1). In the Conowingo tailrace, the September 1987 rate of collection for gizzard shad was the highest of any species ( 9,425 fish/hr, Appendix D-2) 1982 to 1987. Gizzard shad was the most frequently caught species in the Conowingo tailrace every year except 1982 when carp was the most frequently collected species (222 fish/hr, Table 6.5-7). The other frequently collected species in the Conowingo tailrace 1982 to 1987 included American eel, white perch, bluegill, channel catfish, yellow perch, pumpkinseed, smallmouth bass, striped bass, redbreast sunfish and green sunfish (Table 6.5-7).

Species assemblage and proportional abundance based on rates of collection in the Conowingo tailrace varied seasonally. Thirteen different species accounted for $68 \%$ to $92 \%$ of the overall CPUE in each month from 1982 to 1987 (Table 6.5-8). Gizzard shad dominated the catch from September through December and were relatively high in proportion in April through June. Comely shiners were relatively abundant in both December and March collections. Carp were most prevalent in spring, and accounted for more than $50 \%$ of the May CPUE across all years (Table 6.5-8); a significant collection of carp in May 1982 yielded 3,291 fish/hr (Appendix D-2). A relatively high rate of collection for American eel was
maintained throughout all months of sampling. American eel had the highest rates of collection in January, March, July and August across all years. White perch were also relatively abundant May through October.

At Lee's Ferry electrofishing samples occurred from July to September 1982 to 1985 and from June to September in 1986 (Appendix D-1). White perch was the most frequently collected fish in all years ranging from an overall catch rate of 43 fish $/ \mathrm{hr}$ in 1983 to 358 fish $/ \mathrm{hr}$ in 1984 (Table 6.5-9). The highest monthly catch rate for white perch among all years (430 fish/hr) occurred in July 1984 (Appendix D-2). Other species frequently collected 1982 to 1987 included American eel, channel catfish, carp, gizzard shad, shorthead redhorse and yellow perch (Table 6.5-9).

Proportional abundance based on rates of collection for electrofishing samples in Lee's Ferry during the summer months had very little month to month variability. In every month June through September across all years, white perch was the dominant species, comprising nearly $50 \%$ of the catch for each month (Table 6.5-10). American eel was the second most frequently collected fish June through September ranging from $28 \%$ in June to $23 \%$ of the overall CPUE in August (Table 6.5-10). Channel catfish were third in proportional abundance, followed by carp throughout the summer months (Table 6.510).

At The Pool, electrofishing samples occurred from July to September 1982 to 1985 and June to September in 1986 (Appendix D-1). White perch was the most frequently collected fish in all years ranging from 176 fish $/ \mathrm{hr}$ in 1982 to 256 fish $/ \mathrm{hr}$ in 1983 (Table 6.5-11). The highest monthly catch rate in The Pool of any species across all years occurred in July 1983 (white perch; 345 fish/hr, Appendix D-2); 1983 also had the highest catch rate for all species amongst all years at The Pool ( $535 \mathrm{fish} / \mathrm{hr}$, Table 6.54). Other species frequently collected at The Pool included American eel, channel catfish, carp and shorthead redhorse (Table 6.5-11).

Proportional abundance based on rates of collection for electrofishing samples remained consistent throughout the summer months in The Pool. White perch was the dominant species each month over all years, ranging from $42 \%$ of the monthly overall CPUE in September to $51 \%$ of the monthly overall CPUE in June (Table 6.5-12). American eel was the second most proportionally abundant species, ranging from $25 \%$ of collections in June to $37 \%$ in September (Table 6.5-12). Channel catfish was third in proportional abundance for each month followed by carp and shorthead redhorse (Table 6.5-12).

At the tidal zone, electrofishing samples occurred May through December in 1982 and 1986, January and May through November in 1983, January and March and June through December in 1984, June through

December in 1985 and July, August and October 1987 (Appendix D-1). The most frequently collected species varied more from year to year than the other survey areas. White perch was the most frequently collected species in 1982 ( $82 \mathrm{fish} / \mathrm{hr}$, Table 6.5-13). Gizzard shad were the most frequently collected species in 1983 ( 84 fish/hr) and 1985 (4,706 fish/hr, Table 6.5-13). In November 1985, an unusually large number of gizzard shad were observed ( 32,500 fish $/ \mathrm{hr}$ ), the highest amount of fish at any station in any month across all years (Appendix D-2). Yellow perch was the most frequently collected species in 1984 (71 fish/hr). Redbreast sunfish were the most frequently collected species in 1986 ( $84 \mathrm{fish} / \mathrm{hr}$ ) and 1987 ( 90 fish $/ \mathrm{hr}$, Table 6.5-13). Other species common in the tidal zone included American eel, carp, pumpkinseed, bluegill, channel catfish, smallmouth bass, white sucker, spottail shiner and comely shiner (Table 6.5-13).

Proportional abundance based on rates of collection varied seasonally in the tidal zone. Catches in January 1983 and 1984 and March 1984 were dominated by American eel, gizzard shad, comely shiner, green sunfish, yellow perch and pumpkinseed (Table 6.5-14). These species accounted for $88 \%$ of the combined CPUE during both these winter months. Comely shiner was collected more frequently in the winter months and represented $69 \%$ of the total CPUE in March of 1984 (Table 6.5-14). December was dominated by many of the same species as in January, though with fewer American eel and more spottail shiner (Table 6.5-14). The late spring through early fall months (May through October) were dominated by catches of white perch, yellow perch, redbreast sunfish, and American eel. These species combined constituted at least $55 \%$ of the CPUE in these months (Table 6.5-14). The late fall catch was dominated by gizzard shad, which constituted $95 \%$ of the collections in November (Table 6.5-14).

### 6.6 Gill Net Results

A total of 4,054 fish of 28 taxa was collected from 118 gill net sets in the Conowingo tailrace, Lee's Ferry, The Pool and the tidal zone from 1982 to 1984 (Table 6.6-1). Effort in 1982 was 48 net-nights and yielded 1,298 fish of 21 different taxa (Table 6.6-1). A total effort of 45 net-nights occurred in 1983; 1,563 fish of at least eight different taxa were collected (Table 6.6-1). A total of 25 net-nights in 1984 yielded 1,193 fish of 25 different taxa (Table 6.6-1). Because of high river flows in 1984 only one collection was made in The Pool, thus 1984 data from this location was excluded from the analysis.

Channel catfish was the most frequently collected species in gill nets during 1982 to 1984. A total of 1,691 channel catfish were collected overall at a rate of 14 fish/net-night (Table 6.6-1). White perch (8 fish/net-night) and gizzard shad ( 8 fish/net-night) were more commonly collected than other species in all areas across all years (Table 6.6-1).

The overall total catch from 1982 to 1984 in the Conowingo tailrace was 2,592 fish, collected at a rate of 96 fish/net-night (Table 6.6-2). Both the tailrace total catch and collection rate were the highest of the four locations. The largest catch occurred in 1983 ( 934 fish) whereas 1982 yielded the highest catch rate (105 fish/net-night, Table 6.6-2). Channel catfish were the most frequently caught species in the tailrace during 1982 to 1984. The overall CPUE was 43 fish/net-night and channel catfish comprised $44 \%$ of the fish collected (Table 6.6-2). Species composition in the tailrace was very similar from 1982 to 1984; besides channel catfish, gizzard shad (23 fish/net-night), white perch (20 fish/net-night), striped bass (3 fish/net-night), hybrid striped bass ( 3 fish/net-night) and carp ( 2 fish/net-night) were caught most frequently (Table 6.6-2).

The overall catch from 1982 to 1984 at Lee's Ferry totaled 617 fish, collected at a rate of 22 fish/net-night (Table 6.6-3). The most fish were caught in 1983 (253 fish) whereas 1984 yielded the highest rate of collection ( 25 fish/net-night, Table 6.6-3). Overall, from 1982 to 1984 channel catfish was the most commonly collected species at Lee's Ferry. Channel catfish were collected at a rate of 9 fish/net-night and comprised $39 \%$ of the fish assemblage collected across all years (Table 6.6-3). White perch was the most frequently collected species in 1983 and 1984. CPUE was 8 fish/net-night and 10 fish/net-night comprising $38 \%$ and $40 \%$ of the fish collected in those years, respectively (Table 6.6-3). Shorthead redhorse ( 2 fish/net-night), gizzard shad ( 2 fish/net-night), and carp ( 1 fish/net-night) were also frequently collected species at Lee's Ferry (Table 6.6-3).

Sampling at The Pool was limited to only 1982 and 1983 due to higher than average natural river flows in 1984. Overall, 258 fish were collected at Lee's Ferry during the two years at a CPUE of 10 fish/net-night (Table 6.6-4). The highest number of fish was collected in 1982 ( 186 fish) at an overall rate of 14 fish/net-night (Table 6.6-4). Channel catfish was the most frequently collected species in both 1982 and 1983. They comprised $48 \%$ of the fish assemblage and were collected at a rate of 5 fish/net-night across both years (Table 6.6-4). White perch (3 fish/net-night), shorthead redhorse ( 1 fish/net-night) and gizzard shad (1 fish/net-night) were also frequently collected species in The Pool. (Table 6.6-4)

The overall total catch from 1982 to 1984 at tidal zone stations was 587 fish, collected at a CPUE of 15 fish/net-night (Table 6.6-5). The largest collection occurred in 1983 ( 304 fish) at a rate of 25 fish/netnight (Table 6.6-5). Overall, gizzard shad were the most frequently collected species ( 5 fish/net-night) from 1982 to 1984 at the tidal zone and comprised $33 \%$ of the fish assemblage (Table 6.6-5). In 1984, channel catfish was the most frequently collected species (11 fish/net-night) and comprised $45 \%$ of the fish assemblage for that year (Table 6.6-5). Only 32 fish were collected in the tidal zone in 1982; white
perch ( $<1$ fish/net-night) and spottail shiner ( $<1$ fish/net-night) were the most frequently collected species (Table 6.6-5).

Seasonal variability existed among several of the common species caught by gill nets. White perch exhibited several seasonal trends based on CPUEs from 1982 to 1984. At the Conowingo tailrace white perch catches were higher in July and August than in other months, averaging 45 fish/net-night and 27 fish/net-night respectively (Table 6.6-6). At Lee's Ferry white perch CPUE's were significantly higher in September, averaging 13 fish/net-night (Table 6.6-6). There was no discernable monthly trend in white perch CPUE in the tidal zone amongst the years sampled (Table 6.6-6).

The monthly CPUE for gizzard shad was variable, but the highest CPUEs consistently occurred in September and October. CPUEs from the tailrace in 1982 and 1984 averaged 42 fish/net-night in October (Table 6.6-7). At Lee's Ferry the average catch was higher in September ( 5 fish/net-night) than in other months (Table 6.6-7), but overall was lower than in the other locations. The highest CPUE for gizzard shad in the tidal zone occurred in October ( 22 fish/net-night,Table 6.6-7). Catches of gizzard shad consistently increased in the late summer and early fall at each location downstream of the Conowingo Dam.

In contrast, the monthly CPUE of channel catfish was variable at the different locations, with no discernable trend in catch from month to month at the tailrace, Lee's Ferry, or the tidal zone (Table 6.68). Similarly, the monthly CPUE of striped bass was low and variable at each of the sampling areas. There was no discernable monthly trend in catch at the tailrace, Lee's Ferry or the tidal zone (Table 6.6-9)

TABLE 6.4-1: TOTAL NUMBER ICHTHYOPLANKTON COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUEHANNA RIVER 1982 TO 1984

| Species | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ |  |  |
| :--- | :---: | :---: | :---: | ---: | ---: |
| No. Taxa | $\mathbf{1 8}$ | $\mathbf{2 0}$ | $\mathbf{2 2}$ | Total | \% of <br> Total |
| No. Samples | $\mathbf{4 0 5}$ | $\mathbf{4 4 6}$ | $\mathbf{4 7 1}$ |  |  |
| American shad | 7 | 138 | 179 | 324 | $<1 \%$ |
| River herrings | 11,772 | 26,827 | 26,677 | 65,276 | $24 \%$ |
| White perch | 23,270 | 112,249 | 61,589 | 197,108 | $72 \%$ |
| Carp | 371 | 307 | 118 | 796 | $<1 \%$ |
| Gizzard shad | 3,911 | 3,464 | 1,886 | 9,261 | $3 \%$ |
| Other | 1,040 | 1,099 | 806 | 2,945 | $1 \%$ |
| TOTAL | 40,371 | 144,084 | 91,255 | 275,710 |  |

TABLE 6.4-2: SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M3) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, APRIL THROUGH JUNE 1982.

| Species | Eggs | Pro <br> larvae | Post <br> larvae | Older | Total <br> Density | Total <br> Number |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| American shad | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 7 |
| Blueback herring | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 197 |
| River herrings | 0.52 | 0.04 | 0.02 | 0.00 | 0.58 | 11,772 |
| White perch | 1.03 | 0.11 | 0.00 | 0.00 | 1.14 | 23,270 |
| Carp | 0.01 | 0.00 | 0.00 | 0.00 | 0.02 | 371 |
| Gizzard shad | 0.03 | 0.01 | 0.15 | 0.00 | 0.19 | 3,911 |
| Other | 0.01 | 0.02 | 0.01 | 0.00 | 0.04 | 843 |
| TOTAL | 1.60 | 0.19 | 0.18 | 0.00 | 1.97 | 40,371 |
| \% Composition | $81 . \%$ | $10 \%$ | $9 \%$ | $<1 \%$ |  |  |

TABLE 6.4-3:SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M3) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, 30 MARCH THROUGH 28 JUNE 1983.

| Species | Eggs | Pro <br> larvae | Post <br> larvae | Older | Total <br> Density | Total <br> Number |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| American shad | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 138 |
| River herrings | 1.13 | 0.17 | 0.00 | 0.00 | 1.30 | 26,827 |
| Gizzard shad | 0.00 | 0.04 | 0.13 | 0.00 | 0.17 | 3,464 |
| Carp | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 307 |
| White perch | 4.65 | 0.69 | 0.00 | 0.00 | 5.34 | 112,249 |
| Other | 0.02 | 0.02 | 0.01 | 0.00 | 0.05 | 1,099 |
| TOTAL | 5.82 | 0.92 | 0.14 | 0.00 | 6.88 | 144,084 |
| \% Composition | $84 \%$ | $13 \%$ | $2 \%$ | $0 \%$ |  |  |

TABLE 6.4-4: SUMMARY OF ICHTHYOPLANKTON DENSITIES (N/M3) COLLECTED BY 0.5M PLANKTON NETS LOWER SUSQUHANNA RIVER, 3 APRIL THROUGH 28 JUNE 1984.

| Species | Eggs | Pro <br> larvae | Post <br> larvae | Older | Total <br> Density | Total <br> Number |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| American shad | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 179 |
| River herrings | 0.85 | 0.04 | 0.01 | 0.00 | 0.90 | 26,677 |
| White perch | 1.89 | 0.18 | 0.00 | 0.00 | 2.07 | 61,589 |
| Carp | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 118 |
| Gizzard shad | 0.00 | 0.01 | 0.06 | 0.00 | 0.06 | 1,886 |
| Other | 0.01 | 0.01 | 0.00 | 0.00 | 0.03 | 806 |
| TOTAL | 2.76 | 0.24 | 0.07 | 0.00 | 3.074 | 91,255 |
| \% Composition | $90 \%$ | $8 \%$ | $2 \%$ | $0 \%$ |  |  |

TABLE 6.4-5: MONTHLY SUMMARY OF WHITE PERCH (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984.

|  | April |  |  | May |  |  | June |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Egg | Pro <br> larvae | Post larvae | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | Post larvae | Egg | Pro <br> larvae | Post larvae | Egg | Pro larvae | $\begin{gathered} \text { Post } \\ \text { larvae } \end{gathered}$ |
| Upper river | 64 |  |  | 346 | 2 |  | 892 | 4 |  | 1,302 | 6 |  |
| Lower river | 286 |  |  | 4,240 | 54 |  | 3,024 | 29 | 1 | 7,550 | 83 | 1 |
| Upper tidal | 5,657 | 448 |  | 24,650 | 2,486 |  | 14,260 | 1,029 | 4 | 44,567 | 3,963 | 4 |
| Lower tidal | 59 | 284 |  | 467 | 537 |  | 49 | 426 | 1 | 775 | 1,247 | 1 |
| Creeks | 157 |  |  | 1,732 | 14 |  | 166 | 21 |  | 2,055 | 5 |  |
| TOTAL | 6,223 | 731 |  | 31,435 | 3,093 |  | 18,591 | 1,509 | 6 | 56,249 | 5,334 | 6 |

TABLE 6.4-6: MONTHLY SUMMARY OF ALOSA SPP. (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984.

|  | April |  |  | May |  |  | June |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Egg | Pro <br> larvae | Post larvae | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | Post larvae | Egg | Pro larvae | Post larvae | Egg | Pro larvae | Post larvae |
| Upper river | 383 |  |  | 9,443 | 46 |  | 1,743 | 9 |  | 11,569 | 55 |  |
| Lower river | 584 | 1 |  | 5,913 | 37 | 4 | 296 | 5 |  | 6,793 | 43 | 4 |
| Upper tidal | 344 | 4 |  | 5,884 | 443 |  | 388 | 290 | 4 | 6,616 | 737 | 4 |
| Lower tidal |  | 1 |  | 22 | 281 |  | 3 | 39 | 229 | 25 | 321 | 229 |
| Creeks | 16 | 16 |  | 158 | 44 |  | 40 | 7 |  | 214 | 67 |  |
| TOTAL | 1,327 | 22 |  | 21,420 | 851 | 4 | 2,470 | 350 | 233 | 25,217 | 1,223 | 237 |

TABLE 6.4.3-1: MONTHLY SUMMARY OF GIZZARD SHAD (EGGS, PROLARVAE AND POSTLARVAE) BY AREA APRIL THROUGH JUNE 1984.

|  | April |  |  | May |  |  | June |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | Post larvae | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | $\begin{gathered} \text { Post } \\ \text { larvae } \end{gathered}$ | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | $\begin{gathered} \text { Post } \\ \text { larvae } \end{gathered}$ | Egg | $\begin{gathered} \text { Pro } \\ \text { larvae } \end{gathered}$ | $\begin{gathered} \text { Post } \\ \text { larvae } \end{gathered}$ |
| Upper river |  | 1 |  |  | 20 | 2 |  | 15 | 476 |  | 36 | 478 |
| Lower river |  | 2 |  |  | 33 | 1 |  | 8 | 319 |  | 43 | 320 |
| Upper tidal |  | 2 |  |  | 33 | 1 |  | 18 | 440 |  | 53 | 441 |
| Lower tidal |  | 1 |  |  | 10 |  |  | 5 | 498 |  | 16 | 498 |
| Creeks |  |  |  |  |  |  |  |  | 1 |  |  | 1 |
| TOTAL |  | 6 |  |  | 96 | 4 |  | 46 | 1,734 |  | 148 | 1,738 |

TABLE 6.5-1:TOTAL NUMBER HOURS OF ELECTROFISHING LOWER SUSQUEHANNA RIVER: ALL LOCATIONS COMBINED 1982-1987

| Conowingo Tailrace | Lees Ferry | The Pool | Tidal Zone | Total Hours Shocked |
| ---: | ---: | ---: | ---: | ---: |
| 92.0 | 30.5 | 31.3 | 93.2 | 247.0 |

TABLE 6.5-2: MONTHLY TOTAL HOURS OF ELECTROFISHING LOWER SUSQUEHANNA RIVER: ALL LOCATIONS 1982 - 1987

|  | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| January |  | 3.6 | 3.0 |  |  |  | 6.5 |
| March |  |  | 3.2 |  | 1.7 |  | 4.9 |
| April |  |  |  |  | 1.9 |  | 1.9 |
| May | 2.1 | 2.8 |  |  | 3.9 |  | 8.8 |
| June | 3.5 | 4.0 | 3.9 | 8.0 | 8.3 |  | 27.7 |
| July | 7.8 | 8.3 | 8.0 | 7.9 | 8.0 | 4.0 | 43.9 |
| August | 12.3 | 8.5 | 8.3 | 7.9 | 8.3 | 4.0 | 49.2 |
| September | 9.6 | 8.5 | 8.4 | 3.5 | 8.3 | 2.2 | 40.4 |
| October | 8.7 | 4.5 | 4.5 | 4.0 | 4.0 | 3.8 | 29.5 |
| November | 3.9 | 4.0 | 4.4 | 4.0 | 3.1 |  | 19.4 |
| December | 3.9 |  | 3.9 | 3.8 | 3.3 |  | 14.9 |
| Total | 51.8 | 44.1 | 47.5 | 39.1 | 50.6 | 13.9 | 247.0 |

TABLE 6.5-3: ELECTROFISHING LOWER SUSQUEHANNA RIVER ANNUAL L HOURS OF EFFORT, TOTAL FISH AND CPUE (FISH/HR) 1982-1987

|  | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Total Hours |  |  |  |  |  |  |  |
| Shocked | 51.8 | 44.1 | 47.5 | 39.1 | 50.6 | 13.94 | 247.0 |
| Total Fish | 31,599 | 33,204 | 29,091 | 89,633 | 24,653 | 27,278 | 235,458 |
| Total CPUE | 611 | 753 | 612 | 2,295 | 487 | 1,957 | 953 |
| \# Taxa | 52 | 42 | 51 | 42 | 54 | 35 | 66 |

TABLE 6.5-4: ELECTROFISHING LOWER SUSQUEHANNA RIVER OVERALL ANNUAL CPUE (FISH/HR) BY LOCATION 1982-1987

|  |  | Cono <br> Tailrace | Lees <br> Ferry | The <br> Pool | Tidal <br> Zone | All Areas |
| ---: | :--- | ---: | ---: | ---: | ---: | ---: |
| 1982 | Total | 822 | 488 | 425 | 496 | 610 |
| 1983 | Total | 1,175 | 96 | 535 | 528 | 753 |
| 1984 | Total | 874 | 638 | 449 | 417 | 612 |
| 1985 | Total | 907 | 578 | 433 | 5,102 | 2,295 |
| 1986 | Total | 559 | 396 | 434 | 701 | 487 |
| 1987 | Total | 3,053 |  |  | 432 | 1,957 |

TABLE 6.5-5: ELECTROFISHING LOWER SUSQUEHANNA RIVER ANNUAL TOTAL FISH AND COMBINED OVERALL CPUE (FISH/HR) BY LOCATION 1982-1987

|  |  | Cono <br> Tailrace | Lees <br> Ferry | The <br> Pool | Tidal <br> Zone |
| ---: | :--- | ---: | ---: | ---: | ---: |
| 1982 | Total | 16,147 | 2,563 | 2,547 | 10,342 |
| 1983 | Total | 17,018 | 3,578 | 3,212 | 9,396 |
| 1984 | Total | 14,841 | 3,805 | 2,694 | 7,751 |
| 1985 | Total | 12,549 | 3,167 | 2,487 | 71,430 |
| 1986 | Total | 10,597 | 3,164 | 3,273 | 7,619 |
| 1987 | Total | 24,761 |  |  | 2,517 |
|  | Total Catch | 95,913 | 16,277 | 14,213 | 109,055 |
|  | Total Hours | 92.0 | 30.5 | 31.3 | 93.2 |
|  | Combined | 1,042 | 533 | 454 | 1,170 |

TABLE 6.5-6: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY CPUE BY AREA 1982-1987

|  | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Cono <br> Tailrace | Cono <br> Tailrace | Cono <br> Tailrace | Cono <br> Tailrace | Cono <br> Tailrace | Cono <br> Tailrace | Lees <br> Ferry | Lees <br> Ferry |
| January |  | 199 | 187 |  |  |  |  |  |
| March |  |  | 184 |  | 102 |  |  |  |
| April |  |  |  |  | 803 |  |  |  |
| May | 4,003 | 1,366 |  |  | 1,190 |  |  |  |
| June | 650 | 1,054 | 1,490 | 1,075 | 524 |  |  |  |
| July | 592 | 902 | 777 | 808 | 948 | 465 | 457 | 770 |
| August | 538 | 499 | 606 | 578 | 536 | 633 | 445 | 528 |
| September | 569 | 797 | 726 | 821 | 572 | 10,076 | 561 | 561 |
| October | 498 | 740 | 1,298 | 1,456 | 659 | 460 |  |  |
| November | 571 | 3,806 | 2,187 | 1,313 | 399 |  |  |  |
| December | 1,335 |  | 179 | 230 | 243 |  |  |  |
| Total | 822 | 1,174 | 874 | 907 | 559 | 12,443 | 488 | 96 |


|  | 1984 | 1985 | 1986 | 1987 | 1982 | 1983 | 1984 | 1985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lees <br> Ferry | Lees <br> Ferry | Lees <br> Ferry | Lees <br> Ferry | The <br> Pool | The <br> Pool | The <br> Pool | The <br> Pool |
| January |  |  |  |  |  |  |  |  |
| March |  |  |  |  |  |  |  |  |
| April |  |  |  |  |  |  |  |  |
| May |  |  |  |  |  |  |  |  |
| June |  | 536 | 315 |  |  |  |  | 516 |
| July | 716 | 569 | 338 |  | 371 | 717 | 421 | 328 |
| August | 545 | 647 | 450 |  | 460 | 418 | 352 | 448 |
| September | 653 |  | 480 |  | 443 | 472 | 575 |  |
| October |  |  |  |  |  |  |  |  |
| November |  |  |  |  |  |  |  |  |
| December |  |  |  |  |  |  |  |  |
| Total | 638 | 578 | 396 |  | 425 | 535 | 449 | 433 |

Table 6.5-6: Cont.

|  | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | The <br> Pool |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | The <br> Pool | Tidal <br> Zone | Tidal <br> Zone | Tidal <br> Zone | Tidal <br> Zone | Tidal <br> Zone | Tidal <br> Zone | Average CPUEs <br> All Areas |  |
| January |  |  |  | 74 | 163 |  |  |  | 156 |
| March |  |  |  |  | 137 |  |  |  | 141 |
| April |  |  |  |  |  |  |  |  | 803 |
| May |  |  | 662 | 803 |  |  | 570 |  | 1,432 |
| June | 422 |  | 438 | 502 | 784 | 565 | 467 | 667 |  |
| July | 285 |  | 323 | 524 | 490 | 354 | 449 | 495 | 550 |
| August | 459 |  | 454 | 411 | 474 | 476 | 429 | 416 | 491 |
| September | 578 |  | 438 | 423 | 448 | 469 | 502 |  | 1,061 |
| October |  |  | 386 | 454 | 513 | 544 | 368 | 381 | 646 |
| November |  |  | 618 | 1,072 | 455 | 32,868 | 895 |  | 4,418 |
| December |  |  | 955 |  | 87 | 439 | 141 |  | 451 |
| Total | 434 |  | 496 | 528 | 417 | 5,102 | 701 | 432 | 1,315 |

TABLE 6.5-7: ELECTROFISHING CONOWINGO TAILRACE LOWER SUSQUEHANNA ANNUAL TOP TEN SPECIES, CPUES (FISH/HR)

| 1982 |  |  | 1983 |  |  | 1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | Carp | 222 | 1 | Gizzard Shad | 473 | 1 | Gizzard Shad | 400 |
| 2 | Gizzard Shad | 148 | 2 | American Eel | 200 | 2 | American Eel | 143 |
| 3 | American Eel | 119 | 3 | White Perch | 89 | 3 | White Perch | 70 |
| 4 | White Perch | 59 | 4 | Yellow Perch | 78 | 4 | Bluegill | 41 |
| 5 | Channel Catfish | 41 | 5 | Pumpkinseed | 57 | 5 | Yellow Perch | 39 |
| 6 | Pumpkinseed | 40 | 6 | Bluegill | 57 | 6 | Pumpkinseed | 33 |
| 7 | Bluegill | 33 | 7 | Channel Catfish | 44 | 7 | Carp | 33 |
| 8 | Smallmouth Bass | 15 | 8 | Striped Bass | 31 | 8 | Channel Catfish | 32 |
| 9 | Green Sunfish | 14 | 9 | Green Sunfish | 27 | 9 | Green Sunfish | 14 |
| 10 | Striped Bass | 11 | 10 | Carp | 26 | 10 | Smallmouth Bass | 12 |
| 1985 |  |  | 1986 |  |  | 1987 |  |  |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | Gizzard Shad | 323 | 1 | Gizzard Shad | 106 | 1 | Gizzard Shad | 10,382 |
| 2 | White Perch | 142 | 2 | American Eel | 100 | 2 | American Eel | 734 |
| 3 | American Eel | 118 | 3 | Yellow Perch | 72 | 3 | White Perch | 345 |
| 4 | Yellow Perch | 98 | 4 | White Perch | 61 | 4 | Yellow Perch | 185 |
| 5 | Channel Catfish | 45 | 5 | Channel Catfish | 27 | 5 | Bluegill | 172 |
| 6 | Bluegill | 41 | 6 | Carp | 27 | 6 | Striped Bass | 131 |
| 7 | Pumpkinseed | 37 | 7 | Redbreast Sunfish | 26 | 7 | Smallmouth Bass | 88 |
| 8 | Carp | 27 | 8 | Bluegill | 26 | 8 | Carp | 84 |
| 9 | Redbreast Sunfish | 16 | 9 | Smallmouth Bass | 17 | 9 | Channel Catfish | 84 |
| 10 | Striped Bass | 12 | 10 | Pumpkinseed | 17 | 10 | Redbreast Sunfish | 66 |

TABLE 6.5-8: ELECTROFISHING LOWER SUSQUEHANNA MONTHLY PROPORTIONAL ABUNDANCE BASED ON CPUE
(FISH/HR) 1982-1987 AT THE CONOWINGO TAILRACE

|  | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alewife |  |  |  | $6 \%$ |  |  |  |  |  |  |  |  |
| American Eel | $30 \%$ |  | $26 \%$ | $15 \%$ | $11 \%$ | $13 \%$ | $26 \%$ | $22 \%$ | $9 \%$ | $15 \%$ | $10 \%$ | $11 \%$ |
| Bluegill |  |  |  |  |  |  | $9 \%$ | $10 \%$ | $3 \%$ |  |  |  |
| Carp | $9 \%$ |  |  | $12 \%$ | $52 \%$ |  |  |  |  |  |  |  |
| Channel Catfish |  |  | $21 \%$ |  |  |  |  |  |  | $6 \%$ | $2 \%$ | $9 \%$ |
| Comely Shiner |  |  | $14 \%$ |  |  |  |  |  |  |  |  | $4 \%$ |
| Gizzard Shad | $22 \%$ |  |  | $28 \%$ | $9 \%$ | $25 \%$ |  | $10 \%$ | $72 \%$ | $38 \%$ | $72 \%$ | $51 \%$ |
| Pumpkinseed | $6 \%$ |  | $9 \%$ |  |  | $7 \%$ | $7 \%$ |  |  |  | $2 \%$ |  |
| Shad sp |  |  |  | $7 \%$ |  |  |  |  |  |  |  |  |
| Spottail Shiner | $9 \%$ |  |  |  |  |  |  |  |  |  |  |  |
| White Perch |  |  |  |  | $8 \%$ | $17 \%$ | $21 \%$ | $20 \%$ | $5 \%$ | $5 \%$ |  |  |
| White Sucker |  |  | $5 \%$ |  |  |  |  |  |  |  |  |  |
| Yellow Perch |  |  |  |  | $6 \%$ | $9 \%$ | $1 \%$ | $8 \%$ | $2 \%$ | $10 \%$ | $6 \%$ | $7 \%$ |

TABLE 6.5-9: ELECTROFISHING LEES FERRY LOWER SUSQUEHANNA ANNUAL TOP FIVE SPECIES, CPUES (FISH/HR)

| 1982 |  |  | 1983 |  |  | 1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | White Perch | 241 | 1 | White Perch | 43 | 1 | White Perch | 358 |
| 2 | American Eel | 139 | 2 | American Eel | 26 | 2 | American Eel | 133 |
| 3 | Channel Catfish | 80 | 3 | Channel Catfish | 24 | 3 | Channel Catfish | 92 |
| 4 | Yellow Perch | 11 | 4 | Gizzard Shad | 1 | 4 | Gizzard Shad | 17 |
| 5 | Carp | 9 | 5 | Carp | 0.3 | 5 | Carp | 13 |
| 1985 |  |  | 1986 |  |  |  |  |  |
|  | Species | CPUE |  | Species | CPUE |  |  |  |
| 1 | White Perch | 309 | 1 | White Perch | 177 |  |  |  |
| 2 | American Eel | 166 | 2 | American Eel | 95 |  |  |  |
| 3 | Channel Catfish | 68 | 3 | Channel Catfish | 81 |  |  |  |
| 4 | Shorthead Redhorse | 9 | 4 | Carp | 14 |  |  |  |
| 5 | Carp | 8 | 5 | Shorthead Redhorse | 8 |  |  |  |

TABLE 6.5-10: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY RELATIVE ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT LEES FERRY

|  | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| American Eel |  |  |  |  |  | $28 \%$ | $24 \%$ | $23 \%$ | $28 \%$ |  |  |  |
| Carp |  |  |  |  |  | $3 \%$ | $2 \%$ | $2 \%$ | $2 \%$ |  |  |  |
| Channel Catfish |  |  |  |  |  | $11 \%$ | $16 \%$ | $18 \%$ | $17 \%$ |  |  |  |
| Gizzard Shad |  |  |  |  |  |  |  | $2 \%$ | $2 \%$ |  |  |  |
| Shorthead <br> Redhorse |  |  |  |  |  | $2 \%$ | $2 \%$ |  |  |  |  |  |
| White Perch |  |  |  |  |  | $52 \%$ | $53 \%$ | $52 \%$ | $48 \%$ |  |  |  |

TABLE 6.5-11 ELECTROFISHING THE POOL LOWER SUSQUEHANNA ANNUAL TOP FIVE SPECIES, CPUES

| 1982 |  |  | 1983 |  |  | 1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | White Perch | 176 | 1 | White Perch | 256 | 1 | White Perch | 177 |
| 2 | American Eel | 135 | 2 | American Eel | 158 | 2 | American Eel | 138 |
| 3 | Channel Catfish | 90 | 3 | Channel Catfish | 87 | 3 | Channel Catfish | 77 |
| 4 | Carp | 11 | 4 | Shorthead Redhorse | 9 | 4 | Carp | 25 |
| 5 | Shorthead Redhorse | 5 | 5 | Carp | 9 | 5 | Shorthead Redhorse | 8 |
| 1985 |  |  | 1986 |  |  |  |  |  |
|  | Species | CPUE |  | Species | CPUE |  |  |  |
| 1 | White Perch | 201 | 1 | White Perch | 192 |  |  |  |
| 2 | American Eel | 159 | 2 | American Eel | 128 |  |  |  |
| 3 | Channel Catfish | 43 | 3 | Channel Catfish | 68 |  |  |  |
| 4 | Shorthead Redhorse | 8 | 4 | Carp | 16 |  |  |  |
| 5 | Carp | 5 | 5 | Shorthead Redhorse | 9 |  |  |  |

TABLE 6.5-12: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY RELATIVE ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT THE POOL

|  | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| American Eel |  |  |  |  |  | $26 \%$ | $28 \%$ | $31 \%$ | $38 \%$ |  |  |  |
| Carp |  |  |  |  |  | $2 \%$ | $3 \%$ | $3 \%$ | $4 \%$ |  |  |  |
| Channel Catfish |  |  |  |  |  | $15 \%$ | $19 \%$ | $14 \%$ | $12 \%$ |  |  |  |
| Gizzard Shad |  |  |  |  |  |  |  | $2 \%$ |  |  |  |  |
| Shorthead <br> Redhorse |  |  |  |  |  |  | $2 \%$ | $2 \%$ | $1 \%$ | $2 \%$ |  |  |
| White Perch |  |  |  |  |  | $51 \%$ | $45 \%$ | $46 \%$ | $42 \%$ |  |  |  |

TABLE 6.5-13 ELECTROFISHING TIDAL ZONE LOWER SUSQUEHANNA ANNUAL TOP TEN SPECIES, CPUES

| 1982 |  |  | 1983 |  |  | 1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | White Perch | 82 | 1 | Gizzard Shad | 84 | 1 | Yellow Perch | 71 |
| 2 | Yellow Perch | 85 | 2 | Yellow Perch | 80 | 2 | Redbreast Sunfish | 68 |
| 3 | Redbreast Sunfish | 67 | 3 | American Eel | 77 | 3 | American Eel | 58 |
| 4 | American Eel | 66 | 4 | White Perch | 66 | 4 | White Perch | 55 |
| 5 | Gizzard Shad | 62 | 5 | Redbreast Sunfish | 65 | 5 | Pumpkinseed | 26 |
| 6 | Carp | 38 | 6 | Pumpkinseed | 35 | 6 | Gizzard Shad | 25 |
| 7 | Pumpkinseed | 21 | 7 | Carp | 24 | 7 | Carp | 20 |
| 8 | Channel Catfish | 18 | 8 | Bluegill | 22 | 8 | Channel Catfish | 19 |
| 9 | White Sucker | 15 | 9 | Channel Catfish | 17 | 9 | Bluegill | 18 |
| 10 | Smallmouth Bass | 8 | 10 | Smallmouth Bass | 14 | 10 | Smallmouth Bass | 17 |
| 1985 |  |  | 1986 |  |  | 1987 |  |  |
|  | Species | CPUE |  | Species | CPUE |  | Species | CPUE |
| 1 | Gizzard Shad | 4,706 | 1 | Redbreast Sunfish | 84 | 1 | Redbreast Sunfish | 90 |
| 2 | Yellow Perch | 93 | 2 | Yellow Perch | 71 | 2 | White Perch | 85 |
| 3 | White Perch | 91 | 3 | Gizzard Shad | 66 | 3 | Yellow Perch | 68 |
| 4 | Redbreast Sunfish | 56 | 4 | White Perch | 66 | 4 | American Eel | 64 |
| 5 | American Eel | 45 | 5 | American Eel | 51 | 5 | Pumpkinseed | 31 |
| 6 | Carp | 24 | 6 | Pumpkinseed | 34 | 6 | Brown Trout | 19 |
| 7 | Channel Catfish | 17 | 7 | Channel Catfish | 19 | 7 | Bluegill | 17 |
| 8 | Pumpkinseed | 13 | 8 | Bluegill | 14 | 8 | Channel Catfish | 14 |
| 9 | Spottail Shiner | 12 | 9 | Smallmouth Bass | 12 | 9 | Gizzard Shad | 8 |
| 10 | Bluegill | 12 | 10 | Carp | 11 | 10 | Smallmouth Bass | 8 |

TABLE 6.5-14: ELECTROFISHING LOWER SUSQUEHANNA RIVER MONTHLY PROPORTIONAL ABUNDANCE BASED ON CPUE (FISH/HR) 1982-1987 AT THE TIDAL ZONE

|  |  | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| American Eel |  | $17 \%$ |  |  |  | $8 \%$ | $13 \%$ | $18 \%$ | $16 \%$ | $18 \%$ | $13 \%$ | $0.4 \%$ |  |
| Carp |  |  |  |  |  |  |  |  |  |  | $8 \%$ |  |  |
| Comely Shiner | $15 \%$ |  | $69 \%$ |  |  |  |  |  |  |  |  |  |  |
| Gizzard Shad |  | $37 \%$ |  |  |  |  |  |  |  |  |  |  |  |
| Green Sunfish |  |  |  | $4 \%$ |  |  |  | $6 \%$ |  |  |  |  |  |
| Pumpkinseed |  | $11 \%$ |  | $3 \%$ |  |  |  | $7 \%$ | $7 \%$ | $7 \%$ |  |  |  |
| Redbreast <br> Sunfish |  |  |  |  |  | $14 \%$ | $23 \%$ | $20 \%$ | $15 \%$ | $15 \%$ | $16 \%$ | $1 \%$ |  |
| Smallmouth <br> Bass |  |  |  |  |  | $4 \%$ | $6 \%$ |  |  |  |  |  |  |
| Spottail Shiner |  |  |  | $5 \%$ |  |  |  |  |  |  |  |  |  |
| White Perch |  |  |  |  |  | $28 \%$ | $15 \%$ | $15 \%$ | $25 \%$ | $23 \%$ | $10 \%$ | $0.3 \%$ |  |
| Yellow Perch |  | $9 \%$ |  | $7 \%$ |  | $23 \%$ | $16 \%$ | $18 \%$ | $11 \%$ | $12 \%$ | $16 \%$ | $2 \%$ | $17 \%$ |

TABLE 6.6-1: GILL NETS LOWER SUSQUEHANNA RIVER ANNUAL TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984

|  | 1982 Total |  |  | 1983 Total |  |  | 1984 Total ${ }^{3}$ |  |  | 1982-1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. of Species | 21 |  |  | 8 (minimum) ${ }^{1}$ |  |  | 25 |  |  | 28 |  |  |
| No. Nets Set | 48 |  |  | 45 |  |  | 25 |  |  | 118 |  |  |
|  | No. | CPUE ${ }^{2}$ | \% | No. | CPUE | \% | \# | CPUE | \% | \# | CPUE | \% |
| Channel Catfish | 714 | 15 | 55\% | 540 | 12 | 35\% | 437 | 17 | 37\% | 1,691 | 14 | 42\% |
| White Perch | 316 | 7 | 24\% | 361 | 8 | 23\% | 246 | 10 | 21\% | 923 | 8 | 23\% |
| Gizzard shad | 151 | 3 | 12\% | 452 | 10 | 29\% | 292 | 12 | 25\% | 895 | 8 | 22\% |
| Shorthead Redhorse | 42 | 1 | 3\% | 46 | 1 | 3\% | 20 | 1 | 5\% | 108 | 1 | 3\% |
| Striped Bass | 8 | 0.2 | 1\% | 71 | 2 | 5\% | 28 | 1 | 2\% | 107 | 1 | 3\% |
| Carp | 24 | 1 | 2\% | 25 | 1 | 2\% | 52 | 2 | 4\% | 101 | 1 | 2\% |
| Hybrid Striped Bass | 8 | 0.2 | 1\% | 12 | 0.3 | 1\% | 67 | 3 | 6\% | 87 | 1 | 2\% |
| Atlantic Menhaden | 6 | 0.1 | 0.5\% | 31 | 1 | 2\% | 9 | 0.4 | 4\% | 46 | 0.4 | 1\% |
| Other Fishes |  |  |  | 25 | 1 | 2\% |  |  |  | 25 | 0.2 | 1\% |
| Spot | 2 | 0.04 | 0.2\% |  |  |  | 9 | 0.4 | 4\% | 11 | 0.1 | 0.3\% |
| White Crappie | 3 | 0.1 | 0.2\% |  |  |  | 8 | 0.3 | 1\% | 11 | 0.1 | 0.3\% |
| Spottail shiner | 8 | 0.2 | 0.6\% |  |  |  |  |  |  | 8 | 0.1 | 0.2\% |
| White catfish | 5 | 0.1 | 0.4\% |  |  |  | 3 | 0.1 | 1\% | 8 | 0.1 | 0.2\% |

1 Other species collected were combined under the category 'Other Fishes' in 1983; the species denoted represent the known minimum species richness

2 The CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes zero catches as part of the effort. In 1982 if a net yielded zero fish the effort was not included in determining the species specific CPUE;
3 Because of high river flows in 1984 only one collection was made in the Pool, thus data from this location has been excluded from the analysis.

Table 6.6-1:Cont.

|  | 1982 Total |  |  | 1983 Total |  |  | 1984 Total $^{3}$ |  |  | 1982-1984 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. of Species | 21 |  |  | 8 (minimum) ${ }^{1}$ |  |  | 25 |  |  | 28 |  |  |
| No. Nets Set | 48 |  |  | 45 |  |  | 25 |  |  | 118 |  |  |
|  | No. | CPUE ${ }^{\text {2 }}$ | \% | No. | CPUE | \% | \# | CPUE | \% | \# | CPUE | \% |
| Walleye | 1 | 0.02 | 0.1\% |  |  |  | 4 | 0.2 | 0.4 | 5 | 0.04 | 0.1\% |
| Yellow perch | 2 | 0.04 | 0.2\% |  |  |  | 2 | 0.1 | 0.2 | 4 | 0.03 | 0.1\% |
| American shad | 1 | 0.02 | 0.1\% |  |  |  | 2 | 0.1 | 0.2 | 3 | 0.03 | 0.1\% |
| Blueback herring | 3 | 0.06 | 0.2\% |  |  |  |  |  |  | 3 | 0.03 | 0.1\% |
| Quillback |  |  |  |  |  |  | 3 | 0.1 | 0.3 | 3 | 0.03 | 0.1\% |
| White sucker | 2 | 0.04 | 0.2\% |  |  |  | 1 | 0.04 | 0.1 | 3 | 0.03 | 0.1\% |
| Largemouth Bass | 1 | 0.02 | 0.1\% |  |  |  | 1 | 0.04 | 0.1 | 2 | 0.02 | 0.0\% |
| Shad |  |  |  |  |  |  | 2 | 0.1 | 0.2 | 2 | 0.02 | 0.0\% |
| Yellow bullhead | 1 | 0.02 | 0.1\% |  |  |  | 1 | 0.04 | 0.1 | 2 | 0.02 | 0.0\% |
| Alewife |  |  |  |  |  |  | 1 | 0.04 | 0.4 | 1 | 0.01 | 0.0\% |
| Black Crappie |  |  |  |  |  |  | 1 | 0.04 | 0.1 | 1 | 0.01 | 0.0\% |
| Hickory shad |  |  |  |  |  |  | 1 | 0.04 | 0.4 | 1 | 0.01 | 0.0\% |
| Redbreast Sunfish |  |  |  |  |  |  | 1 | 0.04 | 0.4 | 1 | 0.01 | 0.0\% |
| Smallmouth Bass |  |  |  |  |  |  | 1 | 0.04 | 0.1 | 1 | 0.01 | 0.0\% |
| Tiger Musky |  |  |  |  |  |  | 1 | 0.04 | 0.1 | 1 | 0.01 | 0.0\% |
| Totals | 1,298 | 27 |  | 1,563 | 35 |  | 1,193 | 48 |  | 4,054 | 34 |  |

1 Other species collected were combined under the category 'Other Fishes' in 1983; the species denoted represent the known minimum species richness
2 The CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes zero catches as part of the effort. In 1982 if a net yielded zero fish the effort was not included in determining the species specific CPUE;
3 Because of high river flows in 1984 only one collection was made in the Pool, thus data from this location has been excluded from the analysis.

TABLE 6.6-2: GILL NETS TAILRACE TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984

| Year | 1982 |  |  | 1983 |  |  | 1984 |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Tailrace |  |  | Tailrace |  |  | Tailrace |  |  | Tailrace |  |  |
| No. Nets Set | 8 |  |  | 9 |  |  | 10 |  |  | 27 |  |  |
|  | No. | $\text { CPUE }^{1}$ | \% | No. | CPUE | \% | No. | CPUE | \% | \# | CPUE | \% |
| Channel catfish | 491 | 61 | 58\% | 381 | 42 | 41\% | 279 | 28 | 34\% | 1,151 | 43 | 44\% |
| Gizzard shad | 125 | 16 | 15\% | 281 | 31 | 30\% | 220 | 22 | 27\% | 626 | 23 | 24\% |
| White perch | 191 | 24 | 23\% | 168 | 19 | 18\% | 174 | 17 | 21\% | 533 | 20 | 21\% |
| Striped bass | 3 | 0.4 | 1\% | 65 | 7 | 7\% | 19 | 2 | 2\% | 87 | 3 | 3\% |
| Hybrid Striped | 7 | 1 | 1\% | 11 | 1 | 1\% | 64 | 6 | 8\% | 82 | 3 | 3\% |
| Carp | 11 | 1 | 2\% | 10 | 1 | 1\% | 39 | 4 | 5\% | 60 | 2 | 2\% |
| Other Fishes |  |  |  | 14 | 2 | 2\% |  |  |  | 14 | 1 | 1\% |
| White crappie | 3 | 0.4 | 1\% |  |  |  | 8 | 1 | 1\% | 11 | 0.4 | 0.4\% |
| Atlantic menhaden | 5 | 1 | 1\% | 4 | 0.4 | 0.4\% |  |  |  | 9 | 0.3 | 0.3\% |
| Walleye | 1 | 0.1 | 0\% |  |  |  | 3 | 0.3 | 0.4\% | 4 | 0.1 | 0.2\% |

In certain instances the CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes 'zero catches' as part of the effort. In 1982 if a gill net yielded zero fish of a particular species the effort was not included in determining the species specific CPUE. All data presented here as CPUE is number of fish / number of nets set. The 1983 and 1984 data were derived from "Annual Report (Article 34; Objective 5): 1983 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" and "Annual Report (Article 34; Objective 5): 1984 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" respectively.

Table 6.6-2: Cont.

| Year | 1982 |  |  | 1983 |  |  | 1984 |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Tailrace |  |  | Tailrace |  |  | Tailrace |  |  | Tailrace |  |  |
| No. Nets Set | 8 |  |  | 9 |  |  | 10 |  |  | 27 |  |  |
|  | No. | CPUE ${ }^{1}$ | \% | No. | CPUE | \% | No. | CPUE | \% | \# | CPUE | \% |
| Yellow perch | 2 | 0.3 | 1 |  |  |  | 1 | 0.1 | 0.1 | 3 | 0.11 | 0.1\% |
| Quillback |  |  |  |  |  |  | 2 | 0.2 | 0.2 | 2 | 0.07 | 0.1\% |
| Yellow bullhead | 1 | 0.1 | 0.4 |  |  |  | 1 | 0.1 | 0.1 | 2 | 0.07 | 0.1\% |
| American shad |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| Black Crappie |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| Blueback herring | 1 | 0.1 | 0.3 |  |  |  |  |  |  | 1 | 0.04 | 0.04\% |
| Largemouth bass |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| Smallmouth Bass |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| Tiger Musky |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| White catfish | 1 | 0.1 | 0.4 |  |  |  |  |  |  | 1 | 0.04 | 0.04\% |
| White sucker |  |  |  |  |  |  | 1 | 0.1 | 0.1 | 1 | 0.04 | 0.04\% |
| Totals | 842 | 105 |  | 934 | 104 |  | 816 | 82 |  | 2,592 | 96 |  |

In certain instances the CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes 'zero catches' as part of the effort. In 1982 if a gill net yielded zero fish of a particular species the effort was not included in determining the species specific CPUE. All data presented here as CPUE is number of fish / number of nets set. The 1983 and 1984 data were derived from "Annual Report (Article 34; Objective 5): 1983 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" and "Annual Report (Article 34; Objective 5): 1984 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" respectively.

TABLE 6.6-3: GILL NETS LEES FERRY TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982 1984

| Year | 1982 |  |  | 1983 |  |  | 1984 |  |  | Total |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Lee's Ferry |  |  | Lee's Ferry |  |  | Lee's Ferry |  |  | Lee's Ferry |  |  |  |
| No. Nets Set | $\mathbf{1 1}$ |  |  | No. | CPUE $^{1}$ | \% | No. | CPUE | \% | $\#$ | CPUE | \% | \# |
|  | CPUE | \% |  |  |  |  |  |  |  |  |  |  |  |
| Channel catfish | 139 | 13 | $58 \%$ | 60 | 5 | $24 \%$ | 44 | 9 | $35 \%$ | 243 | 9 | $39 \%$ |  |
| White perch | 60 | 5 | $25 \%$ | 95 | 8 | $38 \%$ | 50 | 10 | $40 \%$ | 205 | 7 | $33 \%$ |  |
| Shorthead <br> redhorse | 17 | 2 | $7 \%$ | 32 | 3 | $13 \%$ | 14 | 3 | $11 \%$ | 63 | 2 | $10 \%$ |  |
| Gizzard shad | 5 | 0 | $2 \%$ | 49 | 4 | $19 \%$ | 6 | 1 | $5 \%$ | 60 | 2 | $10 \%$ |  |
| Carp | 8 | 1 | $3 \%$ | 10 | 1 | $4 \%$ | 8 | 2 | $6 \%$ | 26 | 1 | $4 \%$ |  |
| Striped bass | 4 | 0.4 | $2 \%$ | 2 | 0.2 | $1 \%$ | 1 | 0.2 | $1 \%$ | 7 | 0.3 | $1 \%$ |  |
| Other Fishes |  |  |  | 5 | 0.4 | $2 \%$ |  |  |  | 5 | 0.2 | $1 \%$ |  |
| White catfish | 4 | 0.4 | $2 \%$ |  |  |  |  |  |  | 4 | 0.1 | $1 \%$ |  |
| Hybrid Striped |  |  |  |  |  |  | 2 | 0.4 | $2 \%$ | 2 | 0.1 | $0.3 \%$ |  |
| Atlantic <br> menhaden | 1 | 0.1 | $0.4 \%$ |  |  |  |  |  |  |  |  | 0.0 |  |
| Yellow perch |  |  |  |  |  |  | 1 | 0.2 | $1 \%$ | 1 | 0.04 | $0.2 \%$ |  |
| Totals | $\mathbf{2 3 8}$ | $\mathbf{2 2}$ |  | $\mathbf{2 5 3}$ | $\mathbf{2 1}$ |  | $\mathbf{1 2 6}$ | $\mathbf{2 5}$ |  | $\mathbf{6 1 7}$ | $\mathbf{2 2}$ |  |  |

In certain instances the CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes 'zero catches' as part of the effort. In 1982 if a gill net yielded zero fish of a particular species the effort was not included in determining the species specific CPUE. All data presented here as CPUE is number of fish / number of nets set. The 1983 and 1984 data were derived from "Annual Report (Article 34; Objective 5): 1983 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" and "Annual Report (Article 34; Objective 5): 1984 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" respectively.

TABLE 6.6-4: GILL NETS THE POOL TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982-1984

| Year | 1982 |  |  | 1983 |  |  | Total $^{3}$ |  |  |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | The Pool |  |  | The Pool |  |  | The Pool |  |  |
| No. Nets Set | 13 |  |  | No. | CPUE | \% | No. | CPUE | \% |
|  | No. | CPUE ${ }^{\mathbf{2}}$ | \% | No |  |  |  |  |  |
| Channel catfish | 81 | 6 | $44 \%$ | 43 | 4 | $60 \%$ | 124 | 5 | $48 \%$ |
| White perch | 57 | 4 | $32 \%$ | 18 | 2 | $25 \%$ | 75 | 3 | $29 \%$ |
| Shorthead <br> redhorse | 22 | 2 | $15 \%$ | 4 | 0.3 | $6 \%$ | 26 | 1 | $10 \%$ |
| Gizzard shad | 15 | 1 | $10 \%$ | 2 | 0.2 | $3 \%$ | 17 | 1 | $7 \%$ |
| Carp | 5 | 0.4 | $3 \%$ |  |  |  | 5 | 0.2 | $2 \%$ |
| Striped bass | 1 | 0.1 | $1 \%$ | 2 | 0.2 | $3 \%$ | 3 | 0.1 | $1 \%$ |
| Blueback herring | 2 | 0.2 | $4 \%$ |  |  |  | 2 | 0.1 | $1 \%$ |
| Hybrid Striped | 1 | 0.1 | $1 \%$ | 1 | 0.1 | $1 \%$ | 2 | 0.1 | $1 \%$ |
| Other Fishes ${ }^{1}$ |  |  |  | 2 | 0.2 | $3 \%$ | 2 | 0.1 | $1 \%$ |
| American shad | 1 | 0.1 | $2 \%$ |  |  |  | 1 | 0.0 | $0.4 \%$ |
| White sucker | 1 | 0.1 | $1 \%$ |  |  |  | 1 | 0.0 | $0.4 \%$ |
| Totals | $\mathbf{1 8 6}$ | $\mathbf{1 4}$ |  | $\mathbf{7 2}$ | $\mathbf{6}$ |  | $\mathbf{2 5 8}$ | $\mathbf{1 0}$ |  |

(1) Other species collected were combined under the category 'Other Fishes' in 1983; the species denoted represent the known minimum species richness
(2) The CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes zero catches as part of the effort. In 1982 if a net yielded zero fish the effort was not included in determining the species specific CPUE;
(3) Because of high river flows in 1984 only one collection was made in the Pool, thus data from this location has been excluded from the analysis.

TABLE 6.6-5: GILL NETS TIDAL ZONE TOTAL FISH, CPUE (FISH/NET-NIGHT) AND PROPORTIONAL ABUNDANCE 1982 1984

| Year | 1982 |  |  | 1983 |  |  | 1984 |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | Tidal Zone |  |  | Tidal Zone |  |  | Tidal Zone |  |  | Tidal Zone |  |  |
| No. Nets Set | 16 |  |  | 12 |  |  | 10 |  |  | 38 |  |  |
|  | No. | CPUE ${ }^{1}$ | \% | No. | CPUE | \% | \# | CPUE | \% | \# | CPUE | \% |
| Gizzard shad | 6 | 0.4 | 19\% | 120 | 10 | 39\% | 66 | 7 | 26 | 192 | 5 | 33\% |
| Channel catfish | 3 | 0.2 | 9\% | 56 | 5 | 18\% | 114 | 11 | 45 | 173 | 5 | 29\% |
| White perch | 8 | 0.5 | 25\% | 80 | 7 | 26\% | 22 | 2 | 9 | 110 | 3 | 19\% |
| Atlantic menhaden |  |  |  | 27 | 2 | 9\% | 9 | 1 | 4 | 36 | 1 | 6\% |
| Shorthead redhorse | 3 | 0.2 | 9\% | 10 | 1 | 3\% | 6 | 1 | 2 | 19 | 1 | 3\% |
| Spot | 2 | 0.1 | 6\% |  |  |  | 9 | 1 | 4 | 11 | 0.3 | 2\% |
| Carp |  |  |  | 5 | 0.4 | 2\% | 5 | 1 | 2 | 10 | 0.3 | 2\% |
| Striped bass |  |  |  | 2 | 0.2 | 1\% | 8 | 1 | 3 | 10 | 0.3 | 2\% |
| Spottail shiner | 8 | 1 | 25\% |  |  |  |  |  |  | 8 | 0.2 | 1\% |
| Other Fishes |  |  |  | 4 | 0.3 | 1\% |  |  |  | 4 | 0.1 | 1\% |
| White catfish |  |  |  |  |  |  | 3 | 0.3 | 1 | 3 | 0.1 | 1\% |
| Shad |  |  |  |  |  |  | 2 | 0.2 | 1 | 2 | 0.1 | 0.3\% |
| Alewife |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| American shad |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| Hickory shad |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| Hybrid Striped |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| Largemouth bass | 1 | 0.1 | 3\% |  |  |  |  |  |  | 1 | 0.03 | 0.2\% |
| Quillback |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| Redbreast Sunfish |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| Walleye |  |  |  |  |  |  | 1 | 0.1 | 0.4 | 1 | 0.03 | 0.2\% |
| White sucker | 1 | 0.1 | 3\% |  |  |  |  |  |  | 1 | 0.03 | 0.2\% |
| Totals | 32 | 2 |  | 304 | 25 |  | 251 | 25 |  | 587 | 15 |  |

TABLE 6.6-6: COMPARISON OF THE MONTHLY CPUE (FISH/NET-NIGHT) OF WHITE PERCH COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Zone | Year | Jul | Aug | Sep | Oct | Nov | $\begin{gathered} \text { Annual } \\ \text { CPUE } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Conowingo Tailrace | 1982 | - | 43 | 17 | 7 | - | 24 |
|  | 1983 | 41 | 19 | 12 | - | 3 | 19 |
|  | 1984 | 50 | 14 | 7 | 11 | - | 17 |
| * | $\begin{gathered} 1982- \\ 1984 \end{gathered}$ | 45 | 27 | 11 | 9 | 3 | 20 |
| Lee's Ferry | 1982 | 6 | 4 | 6 | - | - | 6 |
|  | 1983 | 1 | 1 | 15 | - | - | 8 |
|  | 1984 | 2 | - | 15 | - | - | 10 |
| * | $\begin{gathered} 1982- \\ 1984 \end{gathered}$ | 4 | 2 | 13 | - | - | 7 |
| Tidal Zone | 1982 | - | - | - | 4 | - | 4 |
|  | 1983 | 8 | 14 | 1 | 4 | - | 7 |
|  | 1984 | 4 | 0 | 3 | 3 | - | 2 |
| * | $\begin{gathered} \hline 1982- \\ 1984 \end{gathered}$ | 6 | 7 | 2 | 4 | - | 5 |

*Weighted mean, all years

- No sample taken

TABLE 6.6-7: COMPARISON OF MONTHLY CPUE (FISH/NET-NIGHT) OF GIZZARD SHAD COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM 1982-1984

| Zone | Year | Jul | Aug | Sep | Oct | Nov | Annual <br> CPUE |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Conowingo <br> Tailrace | 1982 | - | 8 | 7 | 41 | - | 16 |
|  | 1983 | 11 | 11 | 20 | - | 160 | 31 |
|  | 1984 | 29 | 7 | 16 | 43 | - | 22 |
| $*$ | $1982-$ <br> 1984 | 20 | 8 | 15 | 42 | 160 | 23 |
| Lee's Ferry | 1982 | 0 | 1 | 1 | - | - | 1 |
|  | 1983 | 0 | 0 | 8 | - | - | 4 |
|  | 1984 | 1 | - | 1 | - | - | 1 |
| $*$ | $1982-$ | 0 | 1 | 5 | - | - | 2 |
| Tidal Zone | 1984 | 1982 | - | - | - | 28 | - |
|  | 1983 | 6 | 6 | 5 | 22 | - | 10 |
|  | 1984 | 2 | 3 | 7 | 17 | - | 7 |
| $*$ | $1982-$ | 4 | 5 | 6 | 22 | - | 10 |

*Weighted mean, all years

- No sample taken

TABLE 6.6-8: COMPARISON COMBINED MONTHLY CPUE (FISH/NET-NIGHT) OF CHANNEL CATFISH COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Zone | Year | Jul | Aug | Sep | Oct | Nov | Annual <br> CPUE |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Conowingo <br> Tailrace | 1982 | - | 34 | 67 | 94 | - | 61 |
|  | 1983 | 49 | 65 | 37 | - | 7 | 42 |
|  | 1984 | 38 | 24 | 31 | 17 | - | 28 |
| $*$ | $1982-$ <br> 1984 | 43 | 40 | 43 | 55 | - | 43 |
| Lee's Ferry | 1982 | 7 | 5 | 29 | - | - | 13 |
|  | 1983 | 4 | 5 | 6 | - | - | 5 |
|  | 1984 | 12 | - | 7 | - | - | 9 |
| $*$ | $1982-$ | 7 | 5 | 12 | - | - | 9 |
| Tidal Zone | 1984 | 1982 | - | - | - | 2 | - |
|  | 1983 | 8 | 3 | 5 | 3 | - | 5 |
|  | 1984 | 10 | 12 | 16 | 5 | - | 11 |
| $*$ | $1982-$ | 9 | 8 | 10 | 3 | - | 7 |

*Weighted mean, all years

- No sample taken

TABLE 6.6-9: COMPARISON OF THE MONTHLY CPUE(FISH/NET-NIGHT) OF STRIPED BASS COLLECTED BY EXPERIMENTAL GILL NET AT THREE LOCATIONS IN THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM.

| Zone | Year | Jul | Aug | Sep | Oct | Nov | Annual <br> CPUE |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Conowingo <br> Tailrace | 1982 | - | 0.3 | - | 1 | - | 0.4 |
|  | 1983 | 5 | 20 | 4 | - | - | 7 |
|  | 1984 | 5 | 4 | 1 | 0 | - | 2 |
| $*$ | $1982-$ <br> 1984 | 5 | 7 | 2 | 0.5 | - | 3 |
| Lee's Ferry | 1982 | 0.4 | - | 1 | - | - | 0.4 |
|  | 1983 | 7 | - | - | - | - | 0.2 |
|  | 1984 | 5 | - | - | - | - | 0.2 |
| $*$ | $1982-$ |  |  |  |  |  |  |
| 1984 | 1 | - | 0.2 | - | - | 0.3 |  |
| Tidal Zone | 1982 | - | - | - | - | - | 0 |
|  | 1983 | - | - | 0 | 1 | - | 0.2 |
|  | 1984 | 1 | 1 | 1 | 2 | - | 1 |
| $*$ | $1982-$ | 1984 | 0.2 | 0.3 | 0.3 | 1 | - |

*Weighted mean, all years

- No sample taken



## EXELON GENERATION COMPANY, LLC

PRE-APPLICATION DOCUMENT CONOWINGO HYDROELECTRIC PROJECT


EXELON GENERATION COMPANY, LLC
PRE-APPLICATION DOCUMENT

Figure 6.1-2:
Locations for Electrofishing, Gill Net, and Ichthyoplankton Sampling Below Conowingo Dam, 1980-1991

FIGURE 6.4-1: MONTHLY TOTAL ICHTHYOPLANKTON COLLECTED 1983 AND 1984 FOR ALOSA SPP., WHITE PERCH AND GIZZARD SHAD.


### 7.0 SUPPLEMENTAL STUDIES OF THE FISH COMMUNITY

### 7.1 Fish Food Habits Below Conowingo Dam

The benthic invertebrate studies summarized in Section 4 noted which autochthonous food items (organisms originating from the sample locations) appeared most often in stomachs of some of the resident fishes below Conowingo Dam. Detailed stomach analyses of individual white perch, channel catfish and yellow perch taken by electrofishing in the tailrace below Conowingo Dam July through December 1982 and 1983 were reported by Weisberg and Janicki (1985). Small zooplankters were abundant in white perch stomachs, but caddisfly larva (Cheumatopsyche) and chironomid larva were more important on a frequency basis, with caddis larvae most important based on percent of the biomass eaten. Chironomids were most important to channel catfish numerically and on a frequency basis. However, similar to white perch, caddis larva formed most of the diet biomass. The amphipod Gammarus was the most important food of yellow perch.

The three common resident fishes examined appeared to utilize both autochthonous (primarily benthic taxa) and allochthonous (organisms originating and transported from areas other than the sampling location, likely from above Conowingo Dam and available in the drift) food resources. Major drift taxa from Conowingo Pond, as reported by Weisberg and Janicki (1985), were the large zooplankters Leptodora and Chaoborus, similar to those found by RMC in 1982 (RMC 1985a). However, the food habit studies concluded that benthic taxa originating from below Conowingo Dam (chironomids and Cheumatopsyche) were more important to these three species.

Additional food habits investigations were performed by RMC during the same temporal period that targeted smallmouth bass and striped bass below Conowingo Dam. Smallmouth bass are year-long residents that move seasonally throughout the entire river reach below Conowingo Dam, whereas striped bass can also be residents much of the year or on nomadic incursions from Chesapeake Bay. Common dietary indices employed included frequency of occurrence, percent weight of total, and percent number of total. The Index of Relative Importance (IRI, $(\mathrm{N}+\mathrm{V}) / \mathrm{F}$ where N is percentage of a certain food organism, V is percentage of food volume and P is percentage of frequency of occurrence) was used to integrate these three indices (numerical percentage, volumetric percentage, and frequency of occurrence percentage) and reduce bias of an individual index (Pinkase et al. 1971).

Most smallmouth bass examined for food habits were $>150 \mathrm{~mm}$. These yearling and older fish were highly predacious and ate mainly gizzard shad (Table 7.1-1) and other fishes. Amphipods (Gammarus)
and mayflies (Ephemeroptera), along with gizzard shad and unidentified fishes, were important food of smallmouth bass less than 150 mm .

Young-of-the-year striped bass (approximately 60 to 150 mm based on RMC 1985b) fed mostly on Chaoborus, a drift organism likely from Conowingo Pond, and chironomids based on frequency of occurrence in 181 stomachs examined (Table 7.1-2). The IRI depicted dominant food items differently, with copepods and fish more important than Chaoborus (RMC 1985b). Copepods reflected daytime feeding whereas Chaorborus were consumed at night (RMC 1985b). Gizzard shad were the most important prey of 70 sub-adult and adult striped bass (243-735 mm FL) examined (RMC 1985c). Sizes of the gizzard shad consumed reflected those abundant seasonally, typically young-of-the-year. Striped bass clearly ate numerous gizzard shad and exhibited gorging behavior; as many as 47 young gizzard shad were counted in one stomach. Gizzard shad dominated the diets of both striped bass and smallmouth bass in late summer and fall, especially in the tailrace and upper tidal reaches.

Young-of-the-year striped bass and smallmouth bass diets in 1982 were examined with an overlap index (McKeown et al. 1984). Low to moderate overlap occurred among smaller individuals due to feeding on chironomids. Overlap increased among larger young as feeding on small gizzard shad increased. The authors noted utilization of all available food resources, and that gizzard shad served to attract striped bass to the Susquehanna River below Conowingo Dam.

When taken together, the various food habits studies depict fishes below Conowingo Dam opportunistically utilizing food resources available from lotic habitats below the dam as well as forage produced in Conowingo Pond upstream. For example, the fishes eaten by smallmouth bass appeared to reflect seasonal prey availability. Minnows (Notropis spp.), catfishes, yellow perch, darters, and white perch were eaten in spring and early summer but gizzard shad dominated later in the year. Food resources recruited from Conowingo Pond and available to downstream fishes include zooplankton, important to both resident fishes and young of larger predators, and most notably young gizzard shad which sustain many of the larger resident and migratory predators below the dam (McKeown et al. 1984).

### 7.2 Age and Growth of Smallmouth Bass

Smallmouth bass have been and remain a highly sought recreational species below Conowingo Dam (see Conowingo RSP Study 3.25: Creel Survey of Conowingo Pond and the Susquehanna River below Conowingo Dam). Smallmouth bass age and growth below Conowingo Dam were evaluated over a 4year period from 1980 to 1983 (RMC 1985a). Four gear types contributed specimens for age
determination: WFL, electrofishing, experimental gill nets, and rod/reel. A total of 2,106 scale samples were aged. The WFL provided $77 \%$ of the smallmouth bass aged, with electrofishing an additional $20 \%$.

The aged samples showed three dominant year classes in the population below Conowingo Dam during the four sample years. Apparent abundance of the 1974 year class was waning by 1980 (Age 6 and Age 7 fish in 1980 and 1981, respectively, Table 7.1-3). Bass from the 1978 and 1980 year classes dominated the population by 1983 as 3 and 5 -year old fish. The electrofishing catch in 1982 foretold the abundance of the 1980 year class as Age 2 individuals before this cohort was effectively sampled by the WFL. Reproduction of smallmouth bass is highly variable in rivers, and development of periodic dominant year classes is characteristic of smallmouth bass (and many other fishes) throughout the Susquehanna River (PFBC 2006).

Growth of smallmouth bass below Conowingo Dam during 1980 to 1983 is summarized for 1,577 fish in Table 7.1-4. Mean fork length data depict a typical growth pattern. Based on mean FL attained by Age 4 ( 366 mm ), most smallmouth bass were recruited to the harvestable population below Conowingo Dam ( $\sim 305 \mathrm{~mm} \mathrm{TL}$ ) during their 4th year of life. Growth of smallmouth bass below Conowingo Dam was similar to or greater than that reported for several waters in PA and MD (Table 7.1-5 in RMC 1985b).

TABLE 7.1-1: INDEX OF RELATIVE IMPORTANCE (IRI) BY SIZE GROUPS FOR STOMACH CONTENTS OF SMALLMOUTH BASS FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM, 1982.

| Food Item | Smallmouth bass size group (mm) |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{N}$ | $\mathbf{N}$ | $\mathbf{8 0}$ | $\mathbf{8 0 - 1 0 0}$ | $\mathbf{1 0 1 - 1 2 5}$ | $\mathbf{1 2 6 - 1 5 0}$ |
|  |  | $\mathbf{3}$ | $\mathbf{1 0}$ | $\mathbf{1 5}$ | $\mathbf{1 8 9}$ |  |
| Copepoda |  | - | - | - | - | 1 |
| Amphipoda |  | 2,530 | - | 2,386 | 1,006 | 177 |
| Crayfish |  | - | - | - | 504 | 51 |
| Trichoptera |  | 580 | 150 | 115 | 237 | 448 |
| Ephemeroptera |  | 4,945 | 4,370 | 263 | 2,188 | 404 |
| Diptera |  | 409 | - | 915 | 378 | 308 |
| Gizzard shad |  | - | - | 3,939 | - | 2,992 |
| Catfishes |  | - | - | - | - | 33 |
| Yellow perch |  | - | - | - | - | 23 |
| Darters | - | - | - | 298 | 2 |  |
| Unidentified fish |  | - | 4,294 | 232 | 1,592 | 968 |

TABLE 7.1-2: SUMMARY OF FOOD ITEMS CONSUMED BY 181 YOUNG-OF-THE-YEAR STRIPED BASS FROM THE SUSQUEHANNA RIVER BELOW CONOWINGO DAM, 1982.

| Food Item | Frequeny of <br> Occurrence <br> (\%) | Index of Relative <br> Importance |
| :--- | :---: | :---: |
| Amphipoda | $19 \%$ | 29 |
| Chaoborids | $62 \%$ | 1,409 |
| Chironomids | $56 \%$ | 383 |
| Cladocerans | $18 \%$ | 105 |
| Copepods | $31 \%$ | 2,943 |
| Fish | $27 \%$ | 1,453 |
| Isopoda | $21 \%$ | 22 |
| Trichopterans | $20 \%$ | 146 |

TABLE 7.1-3: AGE GROUPS REPRESENTED IN SCALE SAMPLES FROM SMALLMOUTH BASS COLLECTIONS BELOW CONOWINGO DAM, 1980-1983.


TABLE 7.1-4: LENGTH DATA (FL-MM) FOR AGE GROUPS OF SMALLMOUTH BASS COLLECTED BELOW CONOWINGO DAM, 1980-1983. TOTAL FISH $=\mathbf{1 , 5 7 7}$.

|  | Age Group |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Length data | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 2}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |
| No. Fish | 7 | 39 | 557 | 514 | 339 | 80 | 25 | 12 | 2 | 1 | 1 |
| Min FL | 49 | 147 | 223 | 249 | 315 | 358 | 407 | 438 | 464 | 492 | 470 |
| Max FL | 130 | 310 | 392 | 427 | 456 | 478 | 480 | 485 | 470 | 492 | 470 |
| Mean FL | 90 | 206 | 297 | 366 | 405 | 432 | 446 | 459 | 467 | 492 | 470 |

TABLE 7.1-5: COMPARISON OF MEAN TOTAL LENGTH (MM) BY AGE ATTAINED BY SMALLMOUTH BASS AT THE END OF THE GROWING SEASON COLLECTED FROM MARYLAND WATERS AND THE SUSQUEHANNA RIVER DRAINAGE IN PENNSYLVANIA.

| Location | Year Collected | Number | Age Group |  |  |  |  |  |  |  | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Maryland |  |  | I | II | III | IV | V | VI | VII | VIII |  |
| Bennet Creek |  |  | 58 | 137 | 206 | 216 | 239 |  |  |  | Sanderson (1958) |
| Catoctin Creek |  |  | 76 | 173 | 226 | 282 | 340 |  |  |  | Sanderson (958) |
| Israel Creek |  |  | 74 | 168 | 234 | 287 |  |  |  |  | Sanderson (1958) |
| Potomac (Allegheny County) |  |  | 99 | 193 | 244 | 284 | 335 | 373 |  |  | Sanderson (1958) |
| Potomac River (Frederick County) |  |  | 107 | 188 | 249 | 295 | 333 | 386 |  |  | Sanderson (1958) |
| Potomac River |  |  | 99 | 196 | 262 | 328 | 404 | 457 |  |  | MDRE (1955) |
| Deep Creek Lake |  |  | 99 | 175 | 267 | 338 | 378 | 427 |  |  | MDRE (1955) |
| Loch Raven |  |  | 99 | 201 | 305 | 386 | 445 | 475 |  |  | HDRE (1955) |
| Pennsylvania |  |  |  |  |  |  |  |  |  |  |  |
| Susquehanna River |  |  |  |  |  |  |  |  |  |  |  |
| North Branch (Falls) | 1976 | 114 | 95 | 150 | 199 | 247 | 281 | 301 |  |  | Buynak (1978)8 |
| North Branch (SSES) | 1976 | 112 | 96 | 143 | 181 | 217 | 263 | 277 |  |  | Buynak (1978)a |
| Lake Clarke | 1954 | 125 | 81 | 198 | 259 | 307 | 335 |  |  |  | Campbell (1954) |
| Conowingo Pond b | 1962-1973 | 224 | 106 | 182 | 237 | 248 | 352 | 443 |  |  | Heisey (1980)a |
| Conowingo Pond | 1974-1977 | 399 | 106 | 187 | 241 | 321 | 361 | 395 |  |  | Heisey (1980) a |
| Muddy Run Pump Storage Pond | 1967-1973 | 106 | 121 | 184 | 224 | 243 | 274 | 310 |  |  | Heisey (1980) a |
| Muddy Run Pump Storage Pond b | 1974-1977 | 190 | 129 | 242 | 317 | 370 | 389 | 409 |  |  | Heisey (1980)a |
| Below Conowingo Dam | 1980-1983 | 1588 | 96 | 219 | 315 | 388 | 429 | 458 | 473 | 487 | Present Study a |

a Fork length (FL) measurements were converted to total length: (TL) $0.26+1.06$ FL.
b After the introduction of gizzard shad as a forage fish.

### 8.0 LENGTH FREQUENCY

Length frequency data were collected and summarized for several species from electrofishing, gill nets and fish lifts 1982 to 1984 (RMC 1985, $, \underline{b}, \underline{c}$ ). The original referenced plots are in Appendix F. The length frequencies for fishes collected at Lee's Ferry and the Pool were combined (riverine reach) either due to similarity or small sample size. Length frequency plots for channel catfish, red breast sunfish, bluegill, smallmouth bass, largemouth bass, yellow perch and walleye collected at the West Fish Lift in 2010 are also included at the end of Appendix F.

### 8.1 White Perch

During 1982 to 1984 electrofishing samples, the size distribution of white perch throughout the study reach from the Conowingo tailrace to the tidal zone was similar year to year (RMC 1985a: Figure 6.1.2; RMC 1985b: Figure 5.1-2). In 1982 and 1983, over $35 \%$ of white perch in each sampling area were 141 to 160 mm . In 1984, $51 \%$ of the white perch collected were 151 to 170 mm . From 1982 to 1984, white perch 151 to 170 were most abundant in the riverine reach, but were also found in the tailrace and tidal area. Juvenile white perch (length $<120 \mathrm{~mm}$ ) were collected from both the tailrace (in 1982) and the tidal zone (in 1983). Electrofishing samples taken late in the spawning run in the spring reflect the size distribution of transient adult spawners. For example in 1983, up to $30 \%$ of white perch in April and early May exceeded 170 mm with many fish $>200 \mathrm{~mm}$. By mid-summer white perch $>170 \mathrm{~mm}$ typically comprised about only $10 \%$ or less of the catch, with few fish $>200 \mathrm{~mm}$. Summer and fall samples reflect the size of white perch that remain in the river after the spring spawning run has dispersed.

The size distribution of white perch collected by gill nets was similar to that of electrofishing, and variation in length was limited. Greater than $92 \%$ of white perch collected in 1982 and $89 \%$ in 1983 were between 141 to 200 mm in length (RMC 1985a; RMC 1985b). Greater than $75 \%$ of those captured in 1984 measured between $151-190 \mathrm{~mm}$ and greater than $50 \%$ of the white perch collected measured 151 to 171 mm (RMC 1985c: Figure 5.2-4). These lengths corresponded to age groups III and IV (St. Pierre and Davis 1972). The smaller, less effective mesh of the gill nets may have caused lower estimates of the younger fish of this and other species.

### 8.2 Channel Catfish

The size distribution of channel catfish from electrofishing collections in 1982 and 1983 was similar for fish from the riverine reach and tidal zone. Nearly $50 \%$ of channel catfish collected in these areas were 151 to 200 mm (RMC 1985a: Figure 6.1.5; RMC 1985b). A majority of channel catfish this size were collected in the summer and fall. The size distribution of channel catfish in the tailrace was varied and
more small sized fish were accounted. The modal size group (101 to $150 \mathrm{~mm}, 34 \%$ ) of channel catfish in the tailrace was smaller than at the other reaches and more small channel catfish ( 51 to 100 mm ) were collected in the tailrace than the riverine reach or tidal zone. Typically, young-of-the-year channel catfish have been almost exclusively in the tailrace (RMC 1985a).

Little variation in size distribution of channel catfish was observed in gill nets 1982 to 1984. In 1982 and $1983,59 \%$ and $44 \%$ respectively, of the channel catfish collected were 181 to 260 mm in length. Greater than $50 \%$ of the channel catfish captured while gill netting in 1984 measured from 161 to 240 mm . Significant amounts of catfish greater than 381 mm were also collected. Intermediate length groups were not well represented. Catfish length groupings were clustered around 220 mm . Previous age and growth studies (Keany 1984, Fewlass 1980) indicated an over-population of catfish below Conowingo Dam exhibiting stunted growth. This may be responsible for a lack of large catfish and prevalence of small catfish. Recruitment from the overcrowded populations in Conowingo Pond may contribute to the slow growth shown by catfish.

In 1984, the length frequency distribution of channel catfish collected at the WFL exhibited a decreasing seasonal trend (RMC 1985c: Figure 4.1-7). The catch of channel catfish in the spring was dominated by larger fishes with over $65 \%$ of those measured exceeding 271 mm . This was also in the case in the 2010 WFL collections, with $94 \%$ exceeding 271 mm (Appendix F). The summer and fall catch in 1984 was dominated by channel catfish ranging from 171 to 270 mm and 71 to 170 mm , respectively. Over $85 \%$ of the channel catfish collected during the summer and fall were less than 271 mm .

### 8.3 Gizzard shad

The prevalence of small gizzard shad throughout the study reach provides an excellent forage base in all seasons. During electrofishing collections in 1982 to 1984 most gizzard shad were collected in the fall.

In 1982, collections in the tidal zone yielded a majority of smaller ( 51 tol00 mm ) gizzard shad ( $86 \%$, RMC 1985a: Figure 6.1.4); the remaining $14 \%$ were evenly distributed amongst the size classes. The riverine reach was dominated by larger gizzard shad; $80 \%$ were 201 to 300 mm . The tailrace had representative catches from all size classes; $40 \%$ of fish collected were 51 to 100 mm and $28 \%$ were 251 to 300 mm .

Similar to 1982, during the 1983 electrofishing collections the largest percentage of gizzard shad (41\%) in the tidal zone was composed of smaller ( 51 to 100 mm ) fish. The riverine reach had a more heterogeneous distribution throughout the varying size classes; $70 \%$ of the gizzard shad collected were

101 to 200 mm . In the tailrace, the overall size distribution of gizzard shad was dominated by gizzard shad 101 to 150 mm (RMC 1985b: Figure 5.1-4)

The size distribution of gizzard shad collected by electrofishing in the tailrace and riverine reach in 1984 was dominated by individuals exceeding 200 mm (RMC 1985c), whereas the tidal zone was more typical of 1982 and 1983, when size distributions were dominated by smaller (101-150 mm) gizzard shad. In the tidal zone, most of the small gizzard shad were collected in January and were likely 1983 progeny. The lack of small gizzard shad is evidence of poor recruitment in 1984. An unusual episode of high natural river flow occurred in early July, with an associated heavy silt load, may have resulted in low survival of young gizzard shad.

In 1982, a majority of the gizzard shad collected by gill nets ( $74 \%$ ) were 281 to $340 \mathrm{~mm} ; 13 \%$ were less than 281 mm . In 1983, a higher proportion of small gizzard shad were collected as fish less than 281 mm comprised $64 \%$ of the total. In 1984 gizzard shad collected by gill nets showed a relatively even distribution among length groups, though, a higher proportion were in the 281-300 mm grouping (21\%) than the 321 mm and greater groupings (16\%), RMC 1985c: Figure 5.2-3).

In 1984 the majority of gizzard shad $(90 \%)$ at the WFL were adults greater than 200 mm . From 1 to 15 June, 16 June to 15 September and 16 September to 30 November $89 \%, 86 \%$, and $99 \%$ of the gizzard shad collected were over 200 mm respectively (RMC 1985c: Figure 4.1-6). Limited numbers of young-ofyear and juvenile gizzard shad were present in fish lift catches in any season in 1984. This was probably a result of high spring flows in June which resulted in a very weak 1984 year class of gizzard shad.

### 8.4 Yellow Perch

During electrofishing collections in 1983 a majority of yellow perch collected throughout the study reach were 141 to 180 mm (RMC 1985b: Figure 5.1-5). More juvenile yellow perch ( $<120 \mathrm{~mm}$ ) were collected in the tidal zone than in the other locations. Yellow perch from the riverine reach were nearly all $>140$ mm and $12 \%$ were over 221 mm .

During the 1984 electrofishing collections the modal length group of yellow perch was $161-180 \mathrm{~mm}$ in the tailrace and 141-160 mm in the tidal zone. A higher percentage of large perch ( $>200 \mathrm{~mm}$ ) occurred in the tailrace, while substantially more young perch ( $<120 \mathrm{~mm}$ ) occurred in the tidal zone (RMC 1985c).

### 8.5 Smallmouth Bass

Electrofishing collections in 1983 exhibited seasonal variability in length for smallmouth bass as a majority of spring and summer collections consisted of larger fish ( $>151 \mathrm{~mm}$ ) while increasing numbers
of small fish ( $<150 \mathrm{~mm}$ ) were collected in the fall (RMC 1983b: Figure 5.1-6). In the tidal zone, fish 151 to 300 mm comprised $69 \%$ of the catch; most were collected in the spring and summer; $26 \%$ of those collected were $<150 \mathrm{~mm}$ and most were collected in the fall. In the tailrace, the spring and summer catch was dominated by fish 151 to 300 mm ( $56 \%$ ). Some $17 \%$ of the early June catch in the tailrace consisted of large smallmouth bass $>351 \mathrm{~mm}$, while in subsequent months only one fish $>351 \mathrm{~mm}$ was collected. Smallmouth bass $<150 \mathrm{~mm}$ comprised $39 \%$ of the collection in the tailrace, and almost all small fish were collected in the fall. The catch in the tailrace from September through November was comprised primarily of young-of-the-year ( $72 \%$ ). In the riverine reach, a majority of fish were collected in the summer and more than half of the smallmouth bass collected were larger than $250 \mathrm{~mm}(58 \%)$.

In 1984 the catch of bass was dominated by fish $100-200 \mathrm{~mm}$ in length. This size range was mostly comprised of yearling bass, and demonstrated the relative strength of the 1983 year class. Similar to 1983, more large bass ( $>300 \mathrm{~mm}$ ) were taken in the tailrace. By late summer and through fall, few bass $>300$ mm were taken. In contrast to that observed in the tailrace and tidal zone, few bass $<151 \mathrm{~mm}$ were collected in the riverine reach. The highest proportion of large bass ( 350 mm ) occurred in the riverine reach (RMC 1985c).

The size distribution of smallmouth bass (RMC 1985c) varied seasonally during fish lift catches in 1984. The spring catch of smallmouth bass was dominated by fishes that ranged from 321 to 410 mm with over $75 \%$ of the smallmouth bass collected greater than 320 mm . In the spring of $2010,98 \%$ of the 60 smallmouth bass collected, weighed and measured were greater than 320 mm (Appendix F). The catch of smallmouth bass in the summer and fall was limited compared to that in the spring and was generally comprised of smaller fish.

### 8.6 Striped bass

Striped bass collected during electrofishing in 1982 were distinctly adult or juvenile in age class as indicated by the sizes collected. Of the 210 striped bass caught in summer and fall collections at the tailrace, $98 \%$ were $<150 \mathrm{~mm}$. In the tidal zone ( 9 fish), $22 \%$ of the fish collected were 51 to 100 mm and $67 \%$ of the fish collected were $>351 \mathrm{~mm}$. In the riverine reach ( 5 fish), all striped bass were $>301 \mathrm{~mm}$ (RMC 1985a).

The catch of striped bass in the tailrace during electrofishing in 1983 was dominated by the yearling fish of the 1982 year class (RMC 1985b, Figure 5.1-7). Of the 280 fish collected in the tailrace, $75 \%$ were 101 to 150 mm and $21 \%$ were 150 to 200 mm . The mean length of yearling striped bass captured (limited to those $<200 \mathrm{~mm}$; only $6 \%$ of 300 collected exceeded 200 mm ) increased from 129 mm in June to 168 mm
in September. In the tidal and riverine reaches no yearling striped bass were captured, indicating that the tailrace was the primary habitat utilized in the study reach for yearling striped bass.

In both the tailrace and tidal zone the electrofishing catch of striped bass in 1984 was largely made up of individuals 250 to 350 mm in length (fish of the 1982 year class). In the riverine reach nearly half of the striped bass taken were $>350 \mathrm{~mm}$ ( 15 fish) and six ranged from 501-710 mm. The largest striped bass collected to date was taken in September from the tidal zone immediately below Deer Creek. It measured 912 mm and weighed 8.75 kg (RMC 1985c).

Although the sample size for striped bass was small in gill net collections 1982 to 1984, variation in size and abundance was observed. All striped bass collected by gill net in 1982 ( 8 fish) were greater than 301 mm , whereas in 1983 striped bass greater than 301 mm made up only $24 \%$ of those collected ( 15 fish). In 1983, a majority of striped bass were between 181 to 300 mm in length.

The size distribution of striped bass collected at the fish lift in the spring and summer of 1984 was similar to the size distributions from the electrofishing and gill net efforts, with over $90 \%$ of fishes measured ranging in size from 211 to 350 mm . The length frequency distribution of striped bass collected at the fish lift does not reflect the size distribution of these fishes present in the tailrace on a seasonal basis. This is evident through examination of the angler harvest. The range of harvested striped bass was 229 to 766 mm and the mean length was 400 mm (RMC 1985c). The size distribution of striped bass collected at the fish lift indicates predominantly adult striped bass.

The length frequency distribution of striped bass x white bass hybrid at the fish lift (RMC 1985c: Figure 4.1-11) was similar for all seasons. Generally, a minimum of $70 \%$ of the hybrids collected were less than 380 mm in all seasons. The observed increase in the catch of hybrids in the summer and fall was a direct result of the collection of young-of-year (YOY, fish born within the past year) hybrids recruited from Conowingo Pond. Generally, the YOY hybrids ranged from 101 to 200 mm .

### 8.7 Walleye 1984

The 1982 and 1983 electrofishing collections of walleye were comprised of larger, adult fish. In 1982 $92 \%$ of the walleye collected in the tidal zone ( 12 fish) were $>351 \mathrm{~mm}$ and $86 \%$ of the walleye collected in the tailrace ( 42 fish) were $>251 \mathrm{~mm}$. In 1983, all walleye ( 6 fish) collected in the tidal zone were $>251$ and in the tailrace $94 \%$ were $>251 \mathrm{~mm}$. In 1984 electrofishing collections most walleye taken were large; $25 \%$ of walleye in the tailrace were $>461 \mathrm{~mm}$ (RMC 1985a: Figure 6.1.10, RMC 1985b: Figure 5.1-8).

Walleye collected at the fish lift in 1984 ranged in size from 241 to 590 mm . In 2010, $76 \%$ of the Walleye processed for length and weight athe the WFL were 381 to 520 mm (Appendix F). Generally, the catch and size distribution of walleyes exhibited a similar trend in all seasons with the majority of fish collected exceeding 311 mm (RMC 1985c).

### 9.0 LENGTH WEIGHT RELATIONSHIPS AND FISH CONDITION

### 9.1 Length Weight Relationships

Length weight relationships are utilized to estimate or convert weight from length or vice versa. Variation from an expected weight from length of individuals or groups of individuals may be used to determine relative 'fatness', well being, gonad development etc (Anderson and Neuman 1996).

Relationships between length and weight can be expressed by the equation: $\mathrm{W}=\mathrm{cL}^{\mathrm{n}}$

$$
\begin{aligned}
& W=\text { weight } \\
& L=\text { length } \\
& c \text { and } n=\text { constants }
\end{aligned}
$$

The parameters " $c$ " and " $n$ " can be estimated by linear regression of logarithmically transformed length weight data $(\log \mathrm{W}=\log \mathrm{c}+\mathrm{n} \log \mathrm{L})$. The curvilinear length weight relationship becomes linear after logarithmic transformation, allowing for estimation of c and n . An " n " of 3 indicates isometric growth, as the shape of the fish does not change as it grows (Carlander 1969). In general fish with an " n " less than 3 become less rotund as length increases; fish with an " $n$ " greater than 3 become more rotund as length increases.

Several texts have compiled data on length weight relationships for fish body types, families or species or individual populations. Carlander $(\underline{1969}, \underline{1977}, \underline{1997})$ is used here to reference accepted length weight relationship ranges for families or species of fish.

Log transformed lengths and weights for five species (smallmouth bass, largemouth bass, yellow perch, walleye, channel catfish) caught at the WFL in the spring of 2010 were plotted to derive a linear regression equation and are included at the end of Appendix F . The linear regression equation provides an estimation of the length weight relationship in 2010 which can be compared to length weight relationship equations for the same fish species weighed and measured throughout the sampling season during 1982 to 1987 at the WFL and from electrofishing and gill net samples (Table 9.1-1).

The length weight relationship as expressed by the slope of the regression equation based on data collected in 2010 indicates that lengths and weights of selected species collected at the WFL are similar to those collected 1982 to 1987 (Table 9.1-2). Walleye and channel catfish slopes from 2010 are near the median ranges estimated in the 1980's. Smallmouth bass and largemouth bass slopes from 2010 are slightly lower ( $2 \%$ ) than the lowest value from the 1980 's. Though the comparison is limited to one year from the 1980's, yellow perch from 2010 are lower (17\%) than the slope in 1983. The values for both the 1980's fish and those collected in 2010 are within the reference ranges presented. Walleye, channel
catfish, largemouth bass and smallmouth bass were near the median of the ranges presented in Carlander $(\underline{1969}, \underline{1977}, \underline{1997})$. Yellow perch was within the range of means derived in the reference populations, though the 2010 slope was near the lower end of the range (Carlander 1997).

### 9.2 Fish Condition

The relative robustness of a fish can often be described via fish condition ( $\mathrm{K}=$ =weight/(length) ${ }^{3}$ ). Fish condition may express the relative nutritional state as " K " greater or less than a usual weight at a particular length. Condition factor may also vary with stage of development, maturation and sex in some species.

Fish condition from various, distinct populations of the same species have been compiled, analyzed and reported for use as a reference for range of condition factors (Carlander 1969, 1977, 1997). Length weight data were collected for several species at the WFL of the Conowingo Dam in 2010 including channel catfish, redbreast sunfish, bluegill, smallmouth bass, largemouth bass, yellow perch and walleye. For each individual fish a condition factor was derived; the range and mean condition factor for each species is presented in Table 9.2-1. Reference condition factors for each species are also presented.

Fish conditions for species collected at the WFL in 2010 were within the normal range of means presented from various populations of the same species in Carlander 1969, 1977 and 1997.

TABLE 9.1-1: RANGE OF SLOPE OF LENGTH WEIGHT REGRESSION IN 1980'S AND 2010 AND REFERENCE RANGE

| Family | Species | 1980 's | 2010 | Carlander <br> range |
| :---: | :---: | :---: | :---: | :---: |
| Centrarchidae | smallmouth bass | $3.1-3.3$ | 3.0 | $2.5-3.5$ |
| Centrarchidae | largemouth bass | $3.3-3.5$ | 3.2 | $2.5-4.0$ |
| Percidae | yellow perch | $3.0^{(1)}$ | 2.5 | $2.4-3.5$ |
| Percidae | walleye | $3.0-3.2$ | 3.2 | $2.3-3.9$ |
| Ictaluridae | channel catfish | $3.1-3.3$ | 3.2 | $2.9-3.8$ |

(1) Only 80's yellow perch sample year from 1983

TABLE 9.1-2: LENGTH WEIGHT RELATIONSHIPS BY SPECIES AND YEAR

| Year | Species | Length weight formula | $\mathrm{R}^{2}$ | N |
| :---: | :---: | :---: | :---: | :---: |
| 1982 | smallmouth bass | LogW = 3.1636LogL - 5.1825 | 0.98 | 881 |
| 1983 | smallmouth bass | $\operatorname{LogW}=3.1710 \operatorname{LogL}-5.2010$ | 0.98 | 1,005 |
| 1984 | smallmouth bass | LogW = 3.2507LogL - 5.4077 | 0.99 | 1,008 |
| 1985 | smallmouth bass | LogW = 3.1084LogL - 5.0570 | 0.99 | 553 |
| 1986 | smallmouth bass | $\operatorname{LogW}=3.1573 \operatorname{LogL}-5.1832$ | 0.99 | 602 |
| 1987 | smallmouth bass | LogW = 3.1887LogL - 5.2581 | 0.99 | 451 |
| 2010 | smallmouth bass | LogW = 3.0329LogL - 4.9376 | 0.92 | 60 |
| 1982 | largemouth bass | $\operatorname{LogW}=3.3585 \mathrm{LogL}-5.7101$ | 0.98 | 18 |
| 1983 | largemouth bass | $\operatorname{LogW}=3.2658 \mathrm{LogL}-5.4286$ | 0.99 | 52 |
| 1984 | largemouth bass | LogW = 3.4565LogL - 5.9201 | 0.99 | 65 |
| 2010 | largemouth bass | LogW = 3.1897LogL - 5.3201 | 0.77 | 23 |
| 1983 | yellow perch | $\operatorname{LogW}=3.0337 \mathrm{LogL}-4.9211$ | 0.96 | 212 |
| 2010 | yellow perch | $\mathbf{L o g W}=\mathbf{2 . 5 2 5 2 L o g L}$ - 3.8688 | 0.85 | 17 |
| 1982 | walleye | $\operatorname{LogW}=3.0246 \mathrm{LogL}-5.0238$ | 0.93 | 406 |
| 1983 | walleye | $\operatorname{LogW}=3.1047 \mathrm{LogL}-5.2281$ | 0.96 | 695 |
| 1984 | walleye | $\operatorname{LogW}=3.1618 \mathrm{LogL}-5.3972$ | 0.95 | 532 |
| 1985 | walleye | $\operatorname{LogW}=3.1953 \operatorname{LogL}-5.4762$ | 0.98 | 413 |
| 1986 | walleye | $\operatorname{LogW}=3.2436 \mathrm{LogL}$ - 5.6159 | 0.95 | 277 |
| 1987 | walleye | LogW = 2.9944LogL - 4.9457 | 0.89 | 194 |
| 2010 | walleye | LogW = 3.185LogL - 5.5276 | 0.92 | 190 |
| 1982 | channel catfish | $\operatorname{LogW}=3.1098 \operatorname{LogL}-5.1768$ | 0.99 | 1,300 |
| 1983 | channel catfish | $\operatorname{LogW}=3.1026 \operatorname{LogL}-5.1500$ | 0.96 | 1,196 |
| 1984 | channel catfish | $\operatorname{LogW}=3.2317 \mathrm{LogL}-5.4878$ | 0.98 | 1,017 |
| 1985 | channel catfish | LogW = 3.1641 LogL - 5.3075 | 0.99 | 641 |
| 1986 | channel catfish | $\operatorname{LogW}=3.2245 \mathrm{LogL}-5.4749$ | 0.96 | 477 |
| 1987 | channel catfish | $\operatorname{LogW}=3.2936 \mathrm{LogL}$ - 5.6404 | 0.97 | 372 |
| 2010 | channel catfish | LogW = 3.1956LogL - 5.5305 | 0.94 | 351 |

TABLE 9.2-1: 2010 CONDITION FACTOR

| Species | N | Range Total <br> Length | 2010 Range of <br> K | 2010 Mean <br> K | Range of <br> Mean K in <br> Carlander <br> 1969 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Channel catfish | 315 | $134-635$ | $0.3-3.6$ | 0.96 | $0.1-1.1$ |
| Redbreast sunfish | 77 | $147-220$ | $1.2-3.2$ | 2.14 | $1.9-4.2$ |
| Bluegill | 84 | $124-214$ | $0.5-3.9$ | 1.95 | $1.1-3.3$ |
| Smallmouth bass | 60 | $286-480$ | $1.0-1.7$ | 1.41 | $1.3-1.9$ |
| Largemouth bass | 23 | $290-420$ | $1.1-3.3$ | 1.50 | $1.1-1.8$ |
| Yellow perch | 17 | $123-282$ | $0.7-1.6$ | 1.14 | $1.0-1.4$ |
| Walleye | 190 | $222-545$ | $0.4-1.7$ | 0.92 | $0.6-1.1$ |

Condition factor ( $\mathrm{K}=$ weight/(length) $3 * 100,000$ ) based on fishes collected at the WFL 2010.

### 10.0 FISH STRANDING

### 10.1 Methods

Summer 2010 fish observations in rearing habitats below Conowingo Dam were made during on-ground field searches for stranded fishes conducted for Conowingo RSP Study 3.8, Downstream Flow Ramping and Stranding Study. Fish stranding surveys consisted primarily of on-ground surveys of the spillway reach below Conowingo Dam by two 2-person teams. The spillway reach areas searched include approximately 106 acres of largely rocks and boulder substrates interspersed with pool habitats of varying size and depth. A smaller area of interconnected small, shallow pools associated with the mouth of Octoraro Creek downstream of the spillway reach was also searched (see Figure 6.1-1)

Four summer stranding surveys were scheduled. The spatial coverage achieved by the on-ground teams for the summer studies (June 11 through September 1) was generally consistent among studies. The onground teams traversed the spillway reach and observed fish in and along attainable pools of various dimensions that remained at the prevailing summer minimum flow of 5,000 cfs. Survey days began after dawn, following crew transport to the study area and with allowance for sufficient light to permit safe walking and to facilitate counting and identification of fishes in spillway pools, lasted 5-6 h, and typically followed a period of generation the previous day, although this was not always the case (i.e., peaking flow levels were not arranged but rather those dispatched by PJM, Interconnection, the regional transmission organization).

Fish observations were recorded along with position of the siting by hand-held GPS and water temperatures in many of the pools. Stranded fish in de-watered areas were identified to species. Live fish in pooled areas were identified to species, if possible, by simple observation made feasible by normally high water clarity, clear weather, and little or no wind. Numerous small fishes seen but not identifiable to species were grouped as darters, minnows, and young sunfish (Lepomis). Efforts were also made to collect specimens with small-mesh nets for identification of species classified as darters, minnows, and sunfishes as well as young of larger fishes.

### 10.2 Results

Data collected in the 2010 surveys in the spillway reach below Conowingo Dam supplement the fish lift catches for a season not typically sampled by the lifts. These data also identify rearing habitats for small fish and young of larger fishes that in summer are largely unaffected by daily flow changes.

Four surveys were conducted on June 11, July 7, August 11, and September 1, 2010 on both the east and west sides of the spillway reach. The east side had more observed fish ( 6,896 fish) than the west side
(3,422 fish, Table 10.1-1). Fewer fish were caught in the June ( 134 fish) and July ( 168 fish) sample events and the most fish were observed during the August survey (5,608 fish, Table 10.1-1). The most commonly observed fish species were similar on both the east and west sides of the spillway reach. Gizzard shad ( 4,153 fish east, 1,727 fish west), banded killifish ( 1,306 fish east, 1,045 fish west) and sunfish (Lepomis spp., 918 fish east, 225 fish west) were the three most observed species comprising $57 \%, 23 \%$ and $11 \%$ of the total observations respectively (Table 10.1-1). Largemouth bass ( 420 fish) comprised an additional $4 \%$ of the observations (Table 10.1-1). The remaining $5 \%$ of the observations consisted of American eel, carp, minnows, quillback, catfishes, white perch, smallmouth bass, walleye, and darters (Table 10.1-1). Blue crab ( 15 crab) were also observed (Table 10.1-1).

### 10.3 Discussion

Species observed in the fish stranding study were those which inhabit the area below Conowingo Dam in the summer in a variety of hydrological conditions. Generally these areas would be inaccessible and not sample-able or available to other types of gear.

The observations indicate a diverse array of species immediately below Conowingo Dam. Species which are commonly collected at the EFL and WFL were observed. Seven out of the top ten species observed at the EFL in 2009 (gizzard shad, channel catfish, walleye, quillback, common carp, white perch, and brown bullhead) were accounted for in the summer stranding collections. The other three EFL species not accounted for in the stranding surveys were American shad, sea lamprey and alewife, all seasonal migrants. The stranding study also identified species which are not routinely collected at the fish lifts. For example, large amounts of banded killifish were observed in the area below Conowingo spillway but historically, no banded killifish have been lifted during the 38 years of operations at the WFL and 19 years of operation at the EFL. The presence of many juvenile centrarchids (Lepomis, Micropterus) indicate good rearing habitat for those species as well as an array of smaller minnows and darters that provide a good forage base for larger piscivorous fish.

TABLE 10.1-1: SUMMARY OF SUMMER STRANDING STUDIES, JUNE-SEPTEMBER 2010

|  | 11-Jun |  | 7-Jul |  | 11-Aug |  | 1-Sep |  | Summer Total |  | Season |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | East | West | East | West | East | West | East | West | East | West | Total | Percent |
| American eel |  |  |  |  |  | 2 |  |  |  | 2 | 2 | $0.02 \%$ |
| Gizzard shad |  | 43 |  | 22 | 2,570 | 579 | 1,583 | 1,083 | 4,153 | 1,727 | 5,880 | $57 \%$ |
| Carp | $*$ | 8 | 13 |  | 5 |  |  |  | 18 | 8 | 26 | $0.3 \%$ |
| Minnows | 8 |  | 34 | $*$ | 28 |  |  | 8 | 70 | 8 | 78 | $1 \%$ |
| Quillback |  | 9 | $*$ | 1 | 80 | 49 | 25 | 3 | 105 | 62 | 167 | $2 \%$ |
| Catfishes |  | 10 | 1 | 3 | 1 |  | 1 | 2 | 3 | 15 | 18 | $0.2 \%$ |
| Banded killifish |  |  |  | 2 | 590 | 702 | 716 | 341 | 1,306 | 1,045 | 2,351 | $23 \%$ |
| White perch |  | 51 |  | 1 |  |  |  |  |  | 52 | 52 | $1 \%$ |
| Smallmouth bass |  |  |  | 28 | 2 | 5 |  | 7 | 2 | 40 | 42 | $0.4 \%$ |
| Largemouth bass |  | 2 | 1 | 7 | 119 | 86 | 123 | 82 | 243 | 177 | 420 | $4 \%$ |
| Micropterus spp. |  |  |  |  |  | 20 |  |  | 0 | 20 | 20 | $0.2 \%$ |
| Sunfish (Lepomis) | $*$ | 2 | 8 | $*$ | 639 | 75 | 271 | 148 | 918 | 225 | 1,143 | $11 \%$ |
| Walleye |  | 1 |  |  |  |  |  |  |  | 1 | 1 | $0.01 \%$ |
| Darters |  |  | 47 | $*$ | 21 | 25 |  |  | 68 | 25 | 93 | $1 \%$ |
| Blue crabs |  |  |  |  |  |  | 10 | 5 | 10 | 5 | 15 | $0.1 \%$ |
| Unidentified |  |  |  |  |  | 10 |  |  |  | 10 | 10 | $0.1 \%$ |
|  | Totals | $\mathbf{8}$ | $\mathbf{1 2 6}$ | $\mathbf{1 0 4}$ | $\mathbf{6 4}$ | $\mathbf{4 , 0 5 5}$ | $\mathbf{1 , 5 5 3}$ | $\mathbf{2 , 7 2 9}$ | $\mathbf{1 , 6 7 9}$ | $\mathbf{6 , 8 9 6}$ | $\mathbf{3 , 4 2 2}$ | $\mathbf{1 0 , 3 1 8}$ |

* Observed, no estimate made.


### 11.0 CONCLUSION

The literature-based study, including data collected from 1972 to 2010 , provided a broad based characterization of the fisheries and macroinvertebrate communities in the aquatic ecosystem below Conowingo Dam to the area just below Spencer Island.

Over the period 1972 to 2009, Conowingo fish lift catches depicted inter-annual variability and long term trends in fish species assemblage (of species vulnerable to collection in the fish lifts). The species assemblage of both the EFL and WFL catches, dominated by gizzard shad, channel catfish, common carp, and white perch, were similar to those observed in electrofishing, gill net, and ichthyoplankton sampling conducted below Conowingo Dam during the 1980's. The year to year similarities in catches amongst the various additional sampling gears and the fish lifts suggest fish lift collections provide a baseline indicator of the dominant species in the lower Susquehanna River.

The 1980's electrofishing and gill netting collections provided spatio-temporal habitat use detail, and together with fish lift counts and supplementary analyses, such as condition factor, macro-invertebrate sampling, and fish diet analysis provided a comprehensive description of the fish community below Conowingo Dam. Icthyoplankton samples provided further insight into reproduction and utilization of the lower Susquehanna River habitats by earlier life stages of fish. A series of quantitative benthic studies conducted in the non-tidal area of the Lower Susquehanna River below Conowingo Dam from 1980 through 1991 characterized the invertebrate community as moderately rich and moderately dense. The community was generally comprised of facultative or tolerant warm-water genera. Most abundant were: Chironomidae (Cricotopus, Dicrotendipes, and Polypedilum), Cheumatopsyche, Corbicula, Dugesia, Gammarus, Goniobasis, Hydroptila, Manayunkia, and Oligochaeta (Nais). Fish food habit investitgations indicate that diverse trophic interactions are supported in the habitats below Conowingo.

Changes to the fish species assemblage were evident over the period studied; most notably with regards to clupeids. Gizzard shad became the increasingly dominant species over time, American shad generally increased proportionally, and blueback herring decreased proportionally. Gizzard shad were inadvertently introduced into Conowingo Pond in 1972 and the population has increased exponentially. In 1997, implementation of volitional upstream passage via the EFL into Conowingo Pond appears to have accelerated the gizzard shad population growth. As gizzard shad have trended upward in abundance, some other species have declined. White crappie catches at the WFL have declined substantially since the mid 1970's, and it has been noted that one of the primary mechanisms of low recruitment of white crappie is the competition for zooplankton with juvenile gizzard shad (NAI, 1994). From 1991 to 2009, blueback herring comprised $4 \%$ of the overall CPUE at the EFL. In 1997, 1999 and 2001 significant catches of
blueback herring were made. As recently as 2001, 510 herring per lift were collected, the highest amount in any year and the second most proportionally abundant species that year after gizzard shad. Since 2002; however, very few blueback herring have been passed. This decline might reflect recent population declines coast-wide due to a number of potential causes including habitat loss, targeted or bycatch in commercial fisheries, and increased numbers of striped bass and other of predators (ASMFC, 2009). In the 1970's and 1980's the proportion of American shad in the fish lift catch was very low, but as the result of restoration measures, American shad have increased through the 1990's. Since mid 1990's American shad has been one of the top 5 most abundant fish in annual fish lift counts, and is usually second most abundant in the EFL.

Although several species have increased or declined in abundance, the fish species assemblage has remained diverse below Conowingo Dam with the same core group of species as was observed in the 1980's. The fish lift catches have ranged from 30 to 49 taxa annually at the WFL and 25 to 45 taxa annually at the EFL. The taxonomically rich environment is likely a result of Conowingo's lower longitudinal position in the watershed, proximity to the convergence of the freshwater and estuarine environments and the subsequent available regional diversity (Matthews, 1998). Data collected in the 2010 fish stranding summer surveys along with recent years fish lift catches demonstrate the current species assemblage in the Susquehanna River below Conowingo Dam. Seven of the ten most abundant species found at the EFL 1991 to 2009 were represented in the fish stranding survey and the other three, alewife, American shad, and sea lamprey are seasonal migrants that would not have been present during the summer period of the stranding survey. A core assemblage consisting of gizzard shad, white perch, common carp, quillback, comely shiner, channel catfish, walleye, smallmouth and largemouth bass along with seasonal migrants like American shad, blueback herring, alewife, sea lamprey and striped bass form the primary group of inhabitants.

Condition factor of seven species collected at the WFL in 2010 were within the normal range of means presented from various reference populations of the same species in Carlander 1969, 1977 and 1997. The length weight relationship expressed by the slope of the regression equation based on data collected in 2010 indicates that lengths and weights of selected species collected at the WFL were similar to those collected 1982 to 1987 (Table 9.1-2). Both the 1980's fish and those collected in 2010 were within the reference length weight relationship ranges presented. Condition factor and length weight relationships of representative common fish species downstream of Conowingo Dam are comparable to those from other normal, natural populations and are indicative of relatively favorable conditions and habitats in the lower Susquehanna.

### 12.0 REFERENCES

ARCADIS (BBL) and 2007. PPL Martins Creek, LLC; Phase IV Completion Report. Prepared for Martin Creek Steam Electric Station, Bangor, PA.

Atlantic States Marine Fisheries Commission. 2009. Amendment Two to the Interstate Fishery Management Plan For Shad and River Herring (River Herring Management) ASMFC Vision Statement:Approved May 2009. http://www.nmfs.noaa.gov/pr/pdfs/species/riverherring_detailed.pdf

Anderson, R. O. and R. M. Neumann. 1996. Chapter 15 Length, weight, and associated structural indices. Pages $447-481$ in B. R. Murphy and D. W. Willis, editors. Fisheries techniques, second edition. American Fisheries Society, Bethesda, Maryland.

Bode, R. W., \& New York (State). 2002. Quality assurance work plan for biological stream monitoring in New York State. Albany, NY: Stream Biomonitoring Unit, Division of Water, NYS Dept. of Environmental Conservation.

Carlander, K. D. 1969. Handbook of freshwater fishery biology. The Iowa State University Press. Ames, IO.

Carlander, K. D. 1977. Life history data on Centrarchid fishes of the United States and Canada. Ames: Iowa State University Press

Carlander, Kenneth D. 1997. Handbook of freshwater fishery biology 3 Life history data on Ichthyopercid and Percid fishes of the United States and Canada. Ames, Iowa: Iowa State Univ. Press.

Federal Energy Regulatory Commission (FERC). 1989. Order Approving Settlement Agreement for the Conowingo Hydroelectric Project, Project No. 405-009. Docket No. EL80-38-000. January 24, 1989.

Fewlass, L. 1980. Life History and Management of the Channel Catfish in the Susquehanna River. MD Wildlife Admisitration, Department of Natural Resources. Completion Report for MD Federal Aid Project. F-20R.

Janicki, A. J. and R. N. Ross. 1982. Benthic Invertebrate Communities in the Fluctuating Riverine Habitat below Conowingo Dam. Martin Marietta Corp., Environmental Center, Baltimore, MD.

Janicki A. J. and S. B. Weisberg. 1983. Interim Report: Monitoring Studies on the Effect of an Interim Flow on Biota below Conowingo Dam. Martin Marietta Corp., Environmental Center, Baltimore, MD.

Jones, P.W., Martin, F.D., and Hardy, J.D., Jr., 1978, Development of fishes of the Mid-Atlantic Bight, Vol. I. Acipenseridae through Ictaluridae: U.S. Fish and Wildlife Service, Biological Services Program, FWS/OBS-78/12.

Hynes, H. B. N. 1970. The Ecology of Running Waters. The Toronto University Press.
Keany, J. J. 1985. In Prep. Age and Growth of the Channel Catfish in the Susquehann River. Muddy Run Ecology Department

Klemm, D. J., Stober, Q. J., Lazorchak, J. M., \& Environmental Monitoring Systems Laboratory (Cincinnati, Ohio). 1993. Fish field and laboratory methods for evaluating the biological integrity of surface waters. Cincinnati, Ohio: Environmental Monitoring Systems Laboratory-Cincinnati, Office of Modeling, Monitoring Systems, and Quality Assurance, Office of Research and Development, U.S. Environmental Protection Agency.

Larinier, M., Travade, F., Porcher, J. P., Conseil sup rieur de la peche (France), \& Food and Agriculture Organization of the United Nations. 2002. Fishways Biological basis, design criteria, and monitoring. Boves, France Conseil sup rieur de la peche.

Lenat, D. 1994. Revision of the North Carolina biotic index. North Caroliona.
Maryland Department of Natural Resources (MDNR). 2005. Maryland Biological Stream Survey, New Biological Indicators to Better Assess the Condition of Maryland Streams. MD DNR, (Publication Number DNR-12-0305-0100), Annapolis, MD.

Matthews, W. J. 1998. Patterns in freshwater fish ecology. New York: Chapman \& Hall.
McKeown, P.E., T.D. Brush, J.M. Rinehart, and E.T. Euston. 1984. A comparison of the diets of young-of-year striped bass and smallmouth bass in the lower Susquehanna River, Maryland. Presented at: 114th Annual Meeting, American Fisheries Society, Ithaca, NY.

Miller, R. R., \& U.S. Fish and Wildlife Service. 1960. Systematics and biology of the gizzard shad (dorosoma cepedianum) and related fishes. Washington: Fish and Wildlife Service. Government Printing Office.

Normandeau Associates, Inc. 1997 to 2010 Lift Operation Reports.
Normandeau Associates, Inc. (NAI). 1994. Analysis of Potential Factors Affecting the White Crappie Population in Conowingo Pond Prepared for: PECO Energy Company 2301 Market Street Philadelphia, Pennsylvania 19101. NAI (Radiation Ecological Corp., Muddy Run Ecological Laboratory). Drumore, Pennsylvania.

Normandeau Associates, Inc. 1999. Summary of the operations at the Conowingo Dam East fish passage facility in spring, 1998. Prepared for Susquehanna Electric Company, Darlington, MD

Normandeau Associates, Inc. (NAI). 2003 Description of Ecological Studies Conducted on the Lower Susquehanna River, 1981-2002. NAI Drumore, PA.

Normandeau Associates, Inc. 2010. PPL Bruner Island LLC; Aqutic Community Biomonitoring in the Susquehanna River near Bruner Island. NAI Drumore, PA

Pennak, R. W. 1989. Fresh-water Invertebrates of the United States (Third Edition, Protozoa to Mollusca). John Wiley and Sons, Inc. New York, NY.

Pennak, R. W. 1978. Fresh-water Invertebrates of the United States (Second Edition). John Wiley and Sons, Inc. New York, NY.

Pennsylvania Department of Environmental Protection (PA DEP). 2009. A Benthic Index of Biotic Integrity for Wadeable Freestone Riffle-Run Streams in Pennsylvania. PA DEP, Division of Water Quality Assessment and Standards, Harrisburg, PA.

PFBC. 2006. Factors influencing smallmouth bass year-class strength and future smallmouth bass fisheries. Prepared for public meeting: Smallmouth bass 2005 mortality and related trends in the Susquehanna River basin. Harrisburg, PA. January 21, 2006.

Pinkas, L., M.S. Oliphant, and I.L.K. Iverson. 1971. Food habits of albacore, bluefin tuna, and bonito in California waters. CA Dept of Fish and Game, pp. 1-105.

RMC (Radiation Management Corporation). 1985a. Annual Report (Article 34; Objective 5). 1982 fisheries studies for determination of flow needs for protection and enhancement of fish populations below Conowingo Dam. Draft Report.

RMC (Radiation Management Corporation). 1985b. Annual Report (Article 34; Objective 5). 1983 fisheries studies for determination of flow needs for protection and enhancement of fish populations below Conowingo Dam. Draft Report.

RMC (Radiation Management Corporation). 1985c. Annual Report (Article 34; Objective 5). 1984 fisheries studies for determination of flow needs for protection and enhancement of fish populations below Conowingo Dam. Draft Report.

RMC (Radiation Management Corporation). 1991 to 1996. Fish Lift Operation Reports
RMC (Radiation Management Corporation). 1992. Summary of the operations of the Conowingo Dam fish passage facilities in spring 1991. Prepared for Susquehanna Electric Company, Darlington, MD.

Scott, L. 1991. The Effects of Winter Intermittent Flows from the Conowingo Hydroelectric Dam on Macroinvertebrate Populations of the Susquehanna River. Prepared for Maryland Department of Natural Resources by Versar, Inc., Columbia, MD.

St. Pierre, R. A. and J Davis. 1972. Age, Growth and Mortality of the White Perch. Morone Americana in the James and York Rivers, Virginia. Ches. Sci. 13(4) 1272-281.

SRAFRC (Susquehanna River Anadromous Fish Restoration Cooperative). 1982 to 2006. Annual Progress Report. Maryland Department of Natural Resources, New York Div. of Fish, Wildlife \& Marine Resources, Pennsylvania Fish and Boat Commission, Susquehanna River Basin Commission, United States Fish and Wildlife Service, National Marine Fisheries Service

Weisberg, S.B. and A. Janicki. 1985. The Effects Of An Interim Minimum Flow From The Conowingo Dam On Fish Feeding And Benthos In The Susquehanna River. Prepared By Martin Marietta Environmental Systems For Power Plant Siting Program, No. PPSP-UBLS-85-4.

## APPENDIX A - WEST FISH LIFT DATA

A-1. CONOWINGO DAM WEST FISH LIFT TOTAL CATCH AND NUMBER OF FISH LIFTS 1972 - 2009


A-2. CONOWINGO DAM WEST FISH LIFT CPUE AND NUMBER OF FISH LIFTS 1972-2009



A-4. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 1972-1979


A-5. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 1980-1989


A-6. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 1980 - 1989 GIZZARD SHAD REMOVED


A-7. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 1990 - 1999


A-8. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 1990 - 1999 GIZZARD SHAD REMOVED



A-10. WEST FISH LIFT PROPORTIONAL ABUNDANCE, BASED ON TOP TEN CPUES 2000 - 2009 GIZZARD SHAD REMOVED


A-11. GIZZARD SHAD (DOROSOMA CEPEDIANUM) WEST FISH LIFT, FISH PER LIFT


A-12. BLUEBACK HERRING (ALOSA AESTIVALIS) WEST FISH LIFT, FISH PER LIFT


A-13. AMERICAN SHAD (ALOSA SAPIDISSIMA) WEST FISH LIFT, FISH PER LIFT


A-14. ALEWIFE (ALOSA PSEUDOHARENGUS) WEST FISH LIFT, FISH PER LIFT


A-15. WHITE PERCH (MORONE AMERICANA) WEST FISH LIFT, FISH PER LIFT


A-16. STRIPED BASS (MORONE SAXATILIS) WEST FISH LIFT, FISH PER LIFT


- Striped Bass West Fish

Lift, Fish per Lift

A-17.CHANNEL CATFISH (ICTALURUS PUNCTATUS) WEST FISH LIFT, FISH PER LIFT


A-18. SHORTHEAD REDHORSE (MOXOSTOMA MACROEPIDOTUM) WEST FISH LIFT, FISH PER LIFT


A-19. WHITE CRAPPIE (POMOXIS ANNULARIS) WEST FISH LIFT, FISH PER LIFT

— White Crappie West
Fish Lift, Fish per Lift

A-20. SMALLMOUTH BASS (MICROPTERUS DOLOMIEU) WEST FISH LIFT, FISH PER LIFT

$\leadsto$ Smallmouth Bass West
Fish Lift, Fish per Lift

A-21. LARGEMOUTH BASS (MICROPTERUS SALMOIDES) WEST FISH LIFT, FISH PER LIFT


A-22. WALLEYE (SANDER VITREUS) WEST FISH LIFT, FISH PER LIFT


A-23. AMERICAN EEL (ANGUILLA ROSTRATA) WEST FISH LIFT, FISH PER LIFT


A-24. MONTHLY CATCH WEST FISH LIFT 1982 \& 1984

| Month | $\begin{gathered} \text { Apr- } \\ 82 \end{gathered}$ | $\begin{gathered} \text { May- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Jun- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Jul- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \text { Aug- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Sep- } \\ 82 \end{gathered}$ | $\begin{gathered} \text { Oct- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Nov- } \\ 82 \end{gathered}$ | $\begin{gathered} \text { Dec- } \\ 82 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Lifts | 119 | 486 | 83 | 46 | 29 | 78 | 49 | 15 | 5 |
| Min. Fished | 3104 | 14564 | 2655 | 1378 | 1150 | 2995 | 1090 | 315 | 150 |
| Hours Fished | 51.7 | 242.7 | 44.3 | 23.0 | 19.2 | 49.9 | 18.2 | 5.3 | 2.5 |
| Average Temp | 56 | 66.7 | 67.7 | 81.2 | 82 | 74.9 | 68.2 | 51.9 | 50 |
| No. of Species | 32 | 43 | 34 | 26 | 26 | 26 | 18 | 8 | 8 |
| Monthly Catch | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish <br> per <br> lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift |
| Gizzard shad | 1,169.0 | 1,964.5 | 1,543.1 | 408.4 | 216.2 | 942.8 | 2,187.2 | 3,585.0 | 1,251.0 |
| Channel catfish | 14.5 | 36.3 | 147.2 | 43.3 | 42.1 | 86.3 | 251.3 | 40.3 | - |
| Creek <br> Chubsucker | - | 0.005 | - | - | - | - | - | - | 149.0 |
| American eel | 0.1 | 6.0 | 2.1 | 69.1 | 6.8 | 4.1 | 9.5 | 32.3 | 4.0 |
| Carp | 0.8 | 24.5 | 43.8 | 23.7 | 4.0 | 3.9 | 1.7 | - | 1.0 |
| White perch | 5.7 | 40.6 | 8.8 | 1.2 | 0.5 | 0.3 | - | - | - |
| Blueback herring | 0.1 | 51.4 | 0.9 | 0.1 | - | 0.1 | - | - | - |
| Comely shiner | 0.01 | 29.0 | 0.05 | 11.2 | 0.9 | 0.8 | - | 0.3 | - |
| Bluegill | 0.03 | 1.8 | 1.2 | 5.3 | 8.8 | 16.1 | 6.2 | - | - |
| Quillback | 0.5 | 3.1 | 1.7 | 20.7 | 1.3 | - | - | - | 0.6 |
| Shorthead redhorse | 9.9 | 11.8 | 0.05 | - | 0.03 | - | - | - | - |
| Yellow perch | 0.2 | 1.2 | 0.3 | 1.6 | 5.7 | 4.6 | 1.0 | - | 5.0 |
| Redbreast sunfish | - | 1.9 | 1.3 | 4.4 | 6.5 | 3.8 | 0.8 | - | - |
| Striped bass | 0.03 | 0.03 | 0.5 | 7.0 | 8.6 | 0.1 | - | - | - |
| Alewife | 10.4 | 4.5 | 0.04 | - | - | - | - | - | - |
| Pumpkinseed | - | 1.1 | 0.9 | 3.6 | 2.0 | 3.8 | 1.9 | - | - |
| Spotfin shiner | - | 1.1 | 1.3 | 3.0 | 0.9 | 0.8 | - | - | - |
| White crappie | 0.1 | 0.5 | 0.4 | 2.5 | 1.6 | 1.1 | 0.4 | 0.2 | - |

## A-24. Cont.

| Month | $\begin{gathered} \text { Apr- } \\ 82 \\ \hline \end{gathered}$ | May82 | $\begin{gathered} \hline \hline \text { Jun- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \text { Jul- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Aug- } \\ 82 \end{gathered}$ | $\begin{gathered} \hline \hline \text { Sep- } \\ 82 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { Oct- } \\ 82 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { Nov- } \\ 82 \end{gathered}$ | Dec82 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Monthly Catch | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift |
| Walleye | 0.7 | 0.8 | 0.5 | 1.1 | 0.4 | 0.4 | 0.3 | 0.1 | 0.4 |
| American shad | 0.0 | 4.2 | 0.05 | 0.02 | 0.0 | - | - | - | - |
| $\begin{gathered} \text { Smallmouth } \\ \text { bass } \end{gathered}$ | 0.7 | 2.1 | 0.1 | 0.5 | 0.2 | 0.3 | 0.3 | - | - |
| White catfish | 0.01 | 0.6 | 2.2 | 0.8 | 0.1 | - | - | - | - |
| White sucker | 2.8 | 0.4 | 0.2 | - | - | - | 0.04 | 0.1 | 0.2 |
| Striped bass <br> X White <br> Bass | 0.1 | 0.3 | 0.4 | 0.2 | 0.4 | 1.0 | 0.8 | - | - |
| Brown bullhead | 0.5 | 0.1 | 0.8 | 0.3 | 0.2 | 0.04 | 0.02 | - | - |
| Black crappie | - | 0.1 | 0.01 | 0.2 | 0.4 | 0.3 | 0.1 | - | - |
| Brown trout | 0.1 | 0.4 | 0.3 | - | - | - | - | - | - |
| Tiger muskie | 0.03 | 0.1 | 0.01 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | - |
| Spottail shiner | - | 0.6 | - | - | - | - | - | - | - |
| Largemouth bass | 0.03 | 0.03 | 0.03 | 0.1 | 0.1 | 0.1 | 0.2 | - | - |
| Golden shiner | 0.1 | 0.2 | 0.03 | 0.2 | - | - | - | - | - |
| Rock bass | 0.03 | 0.2 | - | - | 0.1 | 0.1 | - | - | - |
| Green sunfish | - | 0.1 | 0.1 | 0.04 | 0.1 | 0.0 | - | - | - |
| Sea lamprey | 0.1 | 0.1 | - | - | - | - | - | - | - |
| Yellow bullhead | 0.1 | 0.005 | 0.05 | - | - | 0.03 | - | - | - |
| Rainbow trout | 0.01 | 0.03 | 0.1 | - | - | - | - | - | - |
| Rosyface shiner | 0.1 | - | - | - | - | - | - | - | - |
| White mullet | - | - | - | - | - | - | 0.1 | - | - |
| Hickory shad | - | 0.03 | - | - | - | - | - | $-$ | - |
| Brook trout | 0.01 | 0.01 | - | - | - | - | - | - | - |

A-24. Cont.

| Month | $\begin{gathered} \hline \hline \text { Apr- } \\ 82 \end{gathered}$ | $\begin{gathered} \text { May- } \\ \mathbf{8 2} \end{gathered}$ | $\begin{gathered} \hline \hline \text { Jun- } \\ 82 \end{gathered}$ | $\begin{gathered} \text { Jul- } \\ 82 \end{gathered}$ | Aug82 | $\begin{gathered} \hline \hline \text { Sep- } \\ 82 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { Oct- } \\ 82 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Nov- } \\ 82 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Dec- } \\ 82 \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Monthly Catch | Fish <br> per <br> lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish per lift | Fish <br> per <br> lift | Fish per lift |
| Margined madton | 0.01 | 0.01 | - | - | - | - | - | - | - |
| Goldfish | - | - | - | - | - | 0.01 | - | - | - |
| Hogchoker - | - | - | - | - | - | 0.01 | - | - | - |
| Minnows | - | - | 0.01 | - | - | - | - | - | - |
| Madtons - | - | - | 0.01 | - | - | - | - | - | - |
| Northern pike | - | 0.01 | - | - | - | - | - | - | - |
| Muskellunge | - | 0.01 | - | - | - | - | - | - | - |
| Shiners | - | 0.01 | - | - | - | - | - | - | - |
| $\begin{gathered} \text { Blacknose } \\ \text { dace } \end{gathered}$ | - | 0.005 | - | - | - | - | - | - | - |
| Total | 1,216.6 | 2,189.9 | 1,758.5 | 608.9 | 307.8 | 1,070.9 | 2,462.2 | 3,658.5 | 1,411.2 |


| Month | Apr-84 | May-84 | Jun-84 | Jul-84 | Aug-84 | Sep-84 | Oct-84 | Nov-84 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total <br> Lifts | $\mathbf{8 3}$ | $\mathbf{3 3 6}$ | $\mathbf{8 9}$ | $\mathbf{8 7}$ | $\mathbf{4 3}$ | $\mathbf{4 1}$ | $\mathbf{3 9}$ | $\mathbf{8}$ |
| Min. <br> Fished | $\mathbf{2 4 3 1}$ | $\mathbf{7 4 8 3}$ | $\mathbf{2 1 2 8}$ | $\mathbf{1 0 7 0}$ | $\mathbf{1 1 2 0}$ | $\mathbf{1 0 4 2}$ | $\mathbf{7 0 0}$ | $\mathbf{7 0}$ |
| Hr <br> Fished | $\mathbf{4 0 . 5}$ | $\mathbf{1 2 4 . 7}$ | $\mathbf{3 5 . 5}$ | $\mathbf{1 7 . 8}$ | $\mathbf{1 8 . 7}$ | $\mathbf{1 7 . 4}$ | $\mathbf{1 1 . 7}$ | $\mathbf{1 . 2}$ |
| Average <br> Temp | $\mathbf{5 1 . 4}$ | $\mathbf{6 2 . 8}$ | $\mathbf{7 1 . 4}$ | $\mathbf{7 7}$ | $\mathbf{7 9 . 8}$ | $\mathbf{7 4 . 5}$ | $\mathbf{6 6 . 3}$ | $\mathbf{5 7 . 3}$ |
| Monthly <br> Catch | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift | Fish <br> per <br> Lift |
| Gizzard <br> shad | 215.27 | $2,203.54$ | $2,188.54$ | 237.36 | 485.25 | 538.74 | $2,117.46$ | $5,750.00$ |
| Channel <br> catfish | 1.87 | 37.78 | 93.44 | 36.32 | 53.72 | 60.00 | 97.43 | 152.50 |
| Carp | 0.11 | 19.28 | 17.12 | 8.23 | 34.42 | 6.76 | 0.49 | 0.25 |
| Striped <br> bass X <br> White <br> Bass | 0.65 | 0.44 | 3.18 | 6.25 | 12.65 | 13.44 | 1.26 | 2.25 |
| American <br> eel | 0.08 | 0.08 | 0.81 | 0.63 | 9.69 | 2.61 | 0.15 | 18.75 |

A-24. Cont.
\(\left.$$
\begin{array}{|c|c|c|c|c|c|c|c|c|}\hline \hline \text { Month } & \text { Apr-84 } & \text { May-84 } & \text { Jun-84 } & \text { Jul-84 } & \text { Aug-84 } & \text { Sep-84 } & \text { Oct-84 } & \text { Nov-84 } \\
\hline \begin{array}{c}\text { Monthly } \\
\text { Catch }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} & \begin{array}{c}\text { Fish } \\
\text { per } \\
\text { Lift }\end{array} \\
\hline \begin{array}{c}\text { Bluegill }\end{array}
$$ \& 0.01 \& 0.21 \& 1.85 \& 1.12 \& 4.00 \& 11.68 \& 7.67 \& 0.08 <br>
\hline \begin{array}{c}Comely <br>

shiner\end{array} \& --2.40 \& 8.26 \& \& - \& \& - \& 2.44 \& 1.04\end{array}\right]-\)| - |
| :--- |
| White perch |
| 0.01 |

A-24. Cont.

| Month | Apr-84 | May-84 | Jun-84 | Jul-84 | Aug-84 | Sep-84 | Oct-84 | Nov-84 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Monthly Catch | Fish per Lift | Fish per Lift | Fish per Lift | Fish per Lift | Fish per Lift | Fish per Lift | Fish per Lift | Fish per Lift |
| $\begin{aligned} & \text { Largemouth } \\ & \text { bass } \end{aligned}$ | - | 0.01 | 0.03 | 0.03 | 0.05 | - | 0.05 | - |
| Sea lamprey | 0.02 | 0.12 | - | - | - | - | - | - |
| Green sunfish | - | 0.01 | 0.01 | 0.02 | 0.02 | - | - | - |
| Hickory shad | 0.06 | 0.00 | - | - | - | - | - | - |
| Yellow bullhead | - | 0.01 | 0.02 | 0.02 | - | - | - | - |
| Atlantic Menhaden | - | - | - | - | - | - | 0.03 | - |
| Muskellunge | - | - | - | - | - | - | 0.03 | - |
| Rainbow trout | - | 0.01 | - | - | - | - | - | - |
| Brook trout X Lake Trout | - | 0.01 | - | - | - | - | - | - |
| Total | 219.57 | 2,286.72 | 2,335.47 | 297.28 | 605.40 | 644.87 | 2,228.54 | 5,924.68 |

A-25. CATCH WEST FISH LIFT 1971-2009

| Year | 1972 | 1972 | 1973 | 1973 | 1974 | 1974 | 1975 | 1975 | 1976 | 1976 | 1977 | 1977 | 1978 | 1978 | 1979 | 1979 | Totals 1972-1979 | Fish Per Lift |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 54 |  | 62 |  | 58 |  | 55 |  | 63 |  | 61 |  | 35 |  | 29 |  | 417 |  |
| Lifts | 817 |  | 1527 |  | 819 |  | 514 |  | 684 |  | 707 |  | 358 |  | 301 |  | 5727 |  |
| Est. Oper. Time(HR) | 608 |  | 996 |  | 500 |  | 500 |  | 307 |  | 375 |  | 413 |  | 187 |  | 3886 |  |
| Fishing Time (HR) | 313 |  | 623 |  | 222 |  | 222 |  | 189 |  | 252 |  | 136 |  | 123 |  | 2080 |  |
| \#Species | 40 | $\begin{gathered} \hline \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 43 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 42 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 41 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 38 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 40 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 44 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 37 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ |  |  |
| Gizzard Shad | 24849 | 30.4 | 45668 | 29.9 | 119672 | 146.1 | 139222 | 270.9 | 382275 | 558.9 | 742056 | 1049.6 | 55104 | 153.9 | 75553 | 251.0 | 1584399 | 276.7 |
| White Perch | 50991 | 62.4 | 647493 | 424.0 | 897113 | 1095.4 | 511699 | 995.5 | 568018 | 830.4 | 224843 | 318.0 | 113164 | 316.1 | 43103 | 143.2 | 3056424 | 533.7 |
| Blueback Herring | 58198 | 71.2 | 330341 | 216.3 | 340084 | 415.2 | 69916 | 136.0 | 35519 | 51.9 | 24395 | 34.5 | 13098 | 36.6 | 2282 | 7.6 | 873833 | 152.6 |
| Channel Catfish | 61042 | 74.7 | 55084 | 36.1 | 75663 | 92.4 | 74042 | 144.1 | 41508 | 60.7 | 90442 | 127.9 | 48575 | 135.7 | 38251 | 127.1 | 484607 | 84.6 |
| Common Carp | 4370 | 5.3 | 16362 | 10.7 | 34383 | 42.0 | 15114 | 29.4 | 6755 | 9.9 | 16256 | 23.0 | 11842 | 33.1 | 14946 | 49.7 | 120028 | 21.0 |
| American Eel | 805 | 1.0 | 2050 | 1.3 | 91937 | 112.3 | 64375 | 125.2 | 60409 | 88.3 | 14601 | 20.7 | 5878 | 16.4 | 1602 | 5.3 | 241657 | 42.2 |
| Alewife | 10345 | 12.7 | 144727 | 94.8 | 16675 | 20.4 | 4311 | 8.4 | 235 | 0.3 | 188 | 0.3 | 5 | 0.0 | 9 | 0.0 | 176495 | 30.8 |
| American Shad | 182 | 0.2 | 65 | 0.0 | 121 | 0.1 | 87 | 0.2 | 82 | 0.1 | 165 | 0.2 | 54 | 0.2 | 50 | 0.2 | 806 | 0.1 |
| Comely Shiner | 5 | 0.0 | 252 | 0.2 | 3870 | 4.7 | 2079 | 4.0 | 740 | 1.1 | 769 | 1.1 | 1152 | 3.2 | 1707 | 5.7 | 10574 | 1.8 |
| Quillback | 7119 | 8.7 | 27780 | 18.2 | 14565 | 17.8 | 8388 | 16.3 | 9882 | 14.4 | 6734 | 9.5 | 2361 | 6.6 | 5134 | 17.1 | 81963 | 14.3 |
| Shorthead Redhorse | 1097 | 1.3 | 4420 | 2.9 | 434 | 0.5 | 445 | 0.9 | 1276 | 1.9 | 1724 | 2.4 | 697 | 1.9 | 2163 | 7.2 | 12256 | 2.1 |
| Spotfin Shiner | 103 | 0.1 | 40 | 0.0 | 3011 | 3.7 | 1231 | 2.4 | 45879 | 67.1 | 7960 | 11.3 | 3751 | 10.5 | 41 | 0.1 | 62016 | 10.8 |
| Striped Bass | 3142 | 3.8 | 495 | 0.3 | 1150 | 1.4 | 174 | 0.3 | 13 | 0.0 | 1196 | 1.7 | 934 | 2.6 | 260 | 0.9 | 7364 | 1.3 |
| Spottail Shiner | 34 | 0.0 | 137 | 0.1 | 2036 | 2.5 | 268 | 0.5 | 1743 | 2.5 | 8107 | 11.5 | 8506 | 23.8 | 1533 | 5.1 | 22364 | 3.9 |
| Redbreast Sunfish | 707 | 0.9 | 2056 | 1.3 | 1398 | 1.7 | 3040 | 5.9 | 3772 | 5.5 | 8277 | 11.7 | 4187 | 11.7 | 3466 | 11.5 | 26903 | 4.7 |
| White Catfish | 3070 | 3.8 | 6394 | 4.2 | 2200 | 2.7 | 6178 | 12.0 | 1451 | 2.1 | 3081 | 4.4 | 982 | 2.7 | 515 | 1.7 | 23871 | 4.2 |
| Walleye | 1840 | 2.3 | 2734 | 1.8 | 1613 | 2.0 | 369 | 0.7 | 2267 | 3.3 | 2140 | 3.0 | 967 | 2.7 | 2491 | 8.3 | 14421 | 2.5 |
| Bluegill | 567 | 0.7 | 1423 | 0.9 | 927 | 1.1 | 3058 | 5.9 | 2712 | 4.0 | 5442 | 7.7 | 1361 | 3.8 | 813 | 2.7 | 16303 | 2.8 |
| Striped Bass x White Bass | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 270 | 0.8 | 273 | 0.9 | 543 | 0.1 |
| White Crappie | 4457 | 5.5 | 664 | 0.4 | 4371 | 5.3 | 9290 | 18.1 | 2987 | 4.4 | 1003 | 1.4 | 673 | 1.9 | 384 | 1.3 | 23829 | 4.2 |
| Yellow Perch | 5955 | 7.3 | 1090 | 0.7 | 682 | 0.8 | 494 | 1.0 | 2904 | 4.2 | 735 | 1.0 | 526 | 1.5 | 379 | 1.3 | 12765 | 2.2 |
| Brown Bullhead | 510 | 0.6 | 5328 | 3.5 | 1612 | 2.0 | 740 | 1.4 | 451 | 0.7 | 2416 | 3.4 | 125 | 0.3 | 284 | 0.9 | 11466 | 2.0 |
| Smallmouth Bass | 182 | 0.2 | 298 | 0.2 | 119 | 0.1 | 153 | 0.3 | 327 | 0.5 | 701 | 1.0 | 262 | 0.7 | 374 | 1.2 | 2416 | 0.4 |
| Pumpkinseed | 229 | 0.3 | 2578 | 1.7 | 2579 | 3.1 | 1000 | 1.9 | 878 | 1.3 | 1687 | 2.4 | 512 | 1.4 | 323 | 1.1 | 9786 | 1.7 |
| White Sucker | 363 | 0.4 | 1034 | 0.7 | 286 | 0.3 | 152 | 0.3 | 444 | 0.6 | 282 | 0.4 | 189 | 0.5 | 906 | 3.0 | 3656 | 0.6 |
| Golden Shiner | 165 | 0.2 | 430 | 0.3 | 437 | 0.5 | 751 | 1.5 | 1622 | 2.4 | 652 | 0.9 | 221 | 0.6 | 304 | 1.0 | 4582 | 0.8 |
| Brown Trout | 172 | 0.2 | 286 | 0.2 | 483 | 0.6 | 219 | 0.4 | 427 | 0.6 | 700 | 1.0 | 261 | 0.7 | 324 | 1.1 | 2872 | 0.5 |
| Rock Bass | 66 | 0.1 | 32 | 0.0 | 31 | 0.0 | 46 | 0.1 | 227 | 0.3 | 128 | 0.2 | 50 | 0.1 | 46 | 0.2 | 626 | 0.1 |
| Hickory Shad | 429 | 0.5 | 739 | 0.5 | 219 | 0.3 | 20 | 0.0 | -- |  | 1 | 0.0 | -- |  | -- |  | 1408 | 0.2 |
| Atlantic Menhaden | -- |  | -- |  | 112 | 0.1 | -- |  | 506 | 0.7 | 1596 | 2.3 | -- |  | -- |  | 2214 | 0.4 |
| Largemouth Bass | 82 | 0.1 | 80 | 0.1 | 23 | 0.0 | 19 | 0.0 | 33 | 0.0 | 14 | 0.0 | 22 | 0.1 | 22 | 0.1 | 295 | 0.1 |
| Yellow Bullhead | 7 | 0.0 | 45 | 0.0 | 1 | 0.0 | 32 | 0.1 | 2 | 0.0 | 47 | 0.1 | 25 | 0.1 | 13 | 0.0 | 172 | 0.0 |
| Black Crappie | 8 | 0.0 | 4 | 0.0 | 25 | 0.0 | 45 | 0.1 | 86 | 0.1 | 199 | 0.3 | 103 | 0.3 | 53 | 0.2 | 523 | 0.1 |
| Rainbow Trout | 34 | 0.0 | 67 | 0.0 | 20 | 0.0 | 24 | 0.0 | 54 | 0.1 | 291 | 0.4 | 70 | 0.2 | 15 | 0.0 | 575 | 0.1 |
| Sea Lamprey | - |  | 2 | 0.0 | - |  | 2 | 0.0 | 29 | 0.0 | 11 | 0.0 | 1 | 0.0 | 3 | 0.0 | 48 | 0.0 |
| Green Sunfish | 3 | 0.0 | -- |  | 4 | 0.0 | 39 | 0.1 | 81 | 0.1 | 168 | 0.2 | 25 | 0.1 | - |  | 320 | 0.1 |
| Tiger Muskie | -- |  | -- |  | - |  | -- |  | -- |  | -- |  | 13 | 0.0 | 132 | 0.4 | 145 | 0.0 |
| Muskellunge | 20 | 0.0 | 104 | 0.1 | 9 | 0.0 | 7 | 0.0 | 12 | 0.0 | 48 | 0.1 | 14 | 0.0 | 5 | 0.0 | 219 | 0.0 |
| Flathead Catfish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shiners | 264 | 0.3 | 3 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 267 | 0.0 |
| Atlantic Needlefish | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Brook Trout | 1 | 0.0 | 3 | 0.0 | 4 | 0.0 | 1 | 0.0 | -- |  | 2 | 0.0 | 23 | 0.1 | -- |  | 34 | 0.0 |
| Bluntnose Minnow | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 4 | 0.0 | -- |  | 4 | 0.0 |


| Year | 1972 | 1972 | 1973 | 1973 | 1974 | 1974 | 1975 | 1975 | 1976 | 1976 | 1977 | 1977 | 1978 | 1978 | 1979 | 1979 | Totals 1972-1979 | $\begin{gathered} \hline \text { Fish Per } \\ \text { Lift } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 54 |  | 62 |  | 58 |  | 55 |  | 63 |  | 61 |  | 35 |  | 29 |  | 417 |  |
| Lifts | 817 |  | 1527 |  | 819 |  | 514 |  | 684 |  | 707 |  | 358 |  | 301 |  | 5727 |  |
| Est. Oper. Time(HR) | 608 |  | 996 |  | 500 |  | 500 |  | 307 |  | 375 |  | 413 |  | 187 |  | 3886 |  |
| Fishing Time (HR) | 313 |  | 623 |  | 222 |  | 222 |  | 189 |  | 252 |  | 136 |  | 123 |  | 2080 |  |
| \#Species | 40 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 43 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 42 | Fish per lift | 41 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 38 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 40 | $\begin{gathered} \hline \text { Fish per } \\ \text { lift } \end{gathered}$ | 44 | Fish per lift | 37 | $\begin{gathered} \hline \text { Fish per } \\ \text { lift } \end{gathered}$ |  |  |
| Northern Hog Sucker | -- |  | 2 | 0.0 | -- |  | 1 | 0.0 | 5 | 0.0 | -- |  | 3 | 0.0 | 6 | 0.0 | 17 | 0.0 |
| Goldfish | -- |  | 27 | 0.0 | 1 | 0.0 | 9 | 0.0 | 4 | 0.0 | 1 | 0.0 | -- |  | -- |  | 42 | 0.0 |
| Tessellated Darter | -- |  | 1 | 0.0 | 4 | 0.0 | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | 7 | 0.0 |
| Logperch | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 27 | 0.1 | -- |  | 27 | 0.0 |
| Brook Trout x Lake Trout | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Northern Pike | -- |  | 2 | 0.0 | 2 | 0.0 | -- |  | -- |  | 2 | 0.0 | 2 | 0.0 | 4 | 0.0 | 12 | 0.0 |
| Striped Bass x White Perch | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chubsucker | 3 | 0.0 | 3 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | 7 | 0.0 |
| Chain Pickerel | -- |  | 1 | 0.0 | 10 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 12 | 0.0 |
| Margined Madtom | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Banded Darter | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 |
| Rosyface Shiner | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Swallowtail Shiner | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Shield Darter | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Greenside Darter | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Longnose Dace | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 4 | 0.0 | -- |  | 4 | 0.0 |
| Tadpole Madtom | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Trouts | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Sunfishes | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Trout | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Rainbow Smelt | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Blacknose Dace | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Mummichog | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 |
| Lampreys | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Lake Herring | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Striped Mullet | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Sunfish Hybrids |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Palomino (Rainbow Trout) | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Redfin Pickerel | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Carps and Minnows | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| River Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Madtoms | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Bigmouth Buffalo | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Total | 241419 | 295.5 | 1300345 | 851.6 | 1617887 | 1975.4 | 917043 | 1784.1 | 1175616 | 1718.7 | 1169061 | 1653.6 | 276045 | 771.1 | 197769 | 657.0 | 6895185 | 1204.0 |


| Year | 1980 | 1980 | 1981 | 1981 | 1982 | 1982 | 1983 | 1983 | 1984 | 1984 | 1985 | 1985 | 1986 | 1986 | 1987 | 1987 | 1988 | 1988 | 1989 | 1989 | Totals 1980-1989 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 30 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \end{array}$ | 37 |  | 44 |  | 29 |  | 34 |  | 55 |  | 59 |  | 60 |  | 63 |  | 51 |  | 462 |  |
| Lifts | 403 |  | 490 |  | 725 |  | 648 |  | 519 |  | 1118 |  | 831 |  | 1414 |  | 1330 |  | 1117 |  | 8595 |  |
| Est. Oper. Time(HR) | 221 |  | 275 |  | 502 |  | 299 |  | 251 |  | 542 |  | 546 |  | 639 |  | 637 |  | 539 |  | 4451 |  |
| Fishing Time (HR) | 117 |  | 178 |  | 336 |  | 224 |  | 192 |  | 421 |  | 449 |  | 532 |  | 513 |  | 457 |  | 3419 |  |
| \#Species | 42 | $\begin{gathered} \begin{array}{c} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \hline \end{gathered}$ | 48 | Fish per lift | 46 | Fish per lift | 40 | $\begin{aligned} & \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ & \hline \hline \end{aligned}$ | 35 | Fish per lift | 41 | Fish per lift | 43 | Fish per lift | 46 | Fish per lift | 49 | Fish per lift | 45 | Fish per lift |  | Fish per lift |
| Gizzard Shad | 275736 | 684.2 | 1156662 | 2360.5 | 1226374 | 1691.6 | 950252 | 1466.4 | 912666 | 1758.5 | 2182888 | 1952.5 | 1714441 | 2063.1 | 2488618 | 1760.0 | 1402565 | 1054.6 | 926213 | 829.2 | 13236415 | 1540.0 |
| White Perch | 26971 | 66.9 | 83363 | 170.1 | 53527 | 73.8 | 23151 | 35.7 | 6402 | 12.3 | 68344 | 61.1 | 56977 | 68.6 | 29995 | 21.2 | 90651 | 68.2 | 15713 | 14.1 | 455094 | 52.9 |
| Blueback Herring | 502 | 1.2 | 618 | 1.3 | 25249 | 34.8 | 517 | 0.8 | 311 | 0.6 | 6763 | 6.0 | 6327 | 7.6 | 5861 | 4.1 | 14570 | 11.0 | 3598 | 3.2 | 64316 | 7.5 |
| $\begin{aligned} & \text { Channel } \\ & \text { Catfish } \\ & \hline \end{aligned}$ | 38929 | 96.6 | 55528 | 113.3 | 40941 | 56.5 | 12559 | 19.4 | 20479 | 39.5 | 15200 | 13.6 | 18898 | 22.7 | 11699 | 8.3 | 36212 | 27.2 | 21692 | 19.4 | 272137 | 31.7 |
| Common Carp | 8879 | 22.0 | 18313 | 37.4 | 15362 | 21.2 | 16273 | 25.1 | 8012 | 15.4 | 6729 | 6.0 | 2930 | 3.5 | 4607 | 3.3 | 8535 | 6.4 | 875 | 0.8 | 90515 | 10.5 |
| American Eel | 377 | 0.9 | 11329 | 23.1 | 3961 | 5.5 | 1080 | 1.7 | 155 | 0.3 | 550 | 0.5 | 364 | 0.4 | 1662 | 1.2 | 103 | 0.1 | 157 | 0.1 | 19738 | 2.3 |
| Alewife | 9 | 0.0 | 129 | 0.3 | 3433 | 4.7 | 50 | 0.1 | 26 | 0.1 | 379 | 0.3 | 2822 | 3.4 | 357 | 0.3 | 647 | 0.5 | 1902 | 1.7 | 9754 | 1.1 |
| American Shad | 139 | 0.3 | 328 | 0.7 | 2039 | 2.8 | 413 | 0.6 | 167 | 0.3 | 1546 | 1.4 | 5195 | 6.3 | 7667 | 5.4 | 5146 | 3.9 | 8218 | 7.4 | 30858 | 3.6 |
| Comely Shiner | 761 | 1.9 | 281 | 0.6 | 14214 | 19.6 | 3176 | 4.9 | 871 | 1.7 | 5141 | 4.6 | 582 | 0.7 | 21199 | 15.0 | 11734 | 8.8 | 35239 | 31.5 | 93198 | 10.8 |
| Quillback | 2929 | 7.3 | 3622 | 7.4 | 1617 | 2.2 | 4679 | 7.2 | 1942 | 3.7 | 957 | 0.9 | 2327 | 2.8 | 1881 | 1.3 | 1578 | 1.2 | 170 | 0.2 | 21702 | 2.5 |
| Shorthead Redhorse | 1394 | 3.5 | 6533 | 13.3 | 6974 | 9.6 | 7558 | 11.7 | 3467 | 6.7 | 3362 | 3.0 | 2057 | 2.5 | 3583 | 2.5 | 4782 | 3.6 | 2735 | 2.4 | 42445 | 4.9 |
| Spotfin <br> Shiner | 314 | 0.8 | 524 | 1.1 | 622 | 0.9 | 501 | 0.8 | -- |  | 2695 | 2.4 | 695 | 0.8 | 796 | 0.6 | 65 | 0.0 | 5381 | 4.8 | 11593 | 1.3 |
| Striped Bass | 904 | 2.2 | 3277 | 6.7 | 60 | 0.1 | 23 | 0.0 | 181 | 0.3 | 213 | 0.2 | 194 | 0.2 | 1337 | 0.9 | 874 | 0.7 | 357 | 0.3 | 7420 | 0.9 |
| Spottail Shiner | 849 | 2.1 | 31 | 0.1 | 315 | 0.4 | 2132 | 3.3 | -- |  | 3525 | 3.2 | 6247 | 7.5 | 155 | 0.1 | 55 | 0.0 | 282 | 0.3 | 13591 | 1.6 |
| Redbreast Sunfish | 1524 | 3.8 | 1007 | 2.1 | 1335 | 1.8 | 401 | 0.6 | 465 | 0.9 | 3366 | 3.0 | 1433 | 1.7 | 1471 | 1.0 | 730 | 0.5 | 443 | 0.4 | 12175 | 1.4 |
| White Catfish | 605 | 1.5 | 2199 | 4.5 | 565 | 0.8 | 224 | 0.3 | 77 | 0.1 | 1094 | 1.0 | 284 | 0.3 | 917 | 0.6 | 3849 | 2.9 | 1740 | 1.6 | 11554 | 1.3 |
| Walleye | 4153 | 10.3 | 2645 | 5.4 | 504 | 0.7 | 663 | 1.0 | 236 | 0.5 | 609 | 0.5 | 308 | 0.4 | 267 | 0.2 | 311 | 0.2 | 319 | 0.3 | 10015 | 1.2 |
| Bluegill | 942 | 2.3 | 1299 | 2.7 | 1184 | 1.6 | 587 | 0.9 | 284 | 0.5 | 6048 | 5.4 | 1654 | 2.0 | 2436 | 1.7 | 1107 | 0.8 | 1561 | 1.4 | 17102 | 2.0 |
| Striped Bass x White Bass | 2674 | 6.6 | 39 | 0.1 | 160 | 0.2 | 355 | 0.5 | 282 | 0.5 | 1377 | 1.2 | 1713 | 2.1 | 5895 | 4.2 | 6203 | 4.7 | 5243 | 4.7 | 23941 | 2.8 |
| White Crappie | 100 | 0.2 | 231 | 0.5 | 303 | 0.4 | 450 | 0.7 | 59 | 0.1 | 345 | 0.3 | 199 | 0.2 | 272 | 0.2 | 125 | 0.1 | 230 | 0.2 | 2314 | 0.3 |
| Yellow Perch | 373 | 0.9 | 1007 | 2.1 | 724 | 1.0 | 387 | 0.6 | 487 | 0.9 | 2145 | 1.9 | 2267 | 2.7 | 632 | 0.4 | 815 | 0.6 | 310 | 0.3 | 9147 | 1.1 |
| Brown Bullhead | 675 | 1.7 | 531 | 1.1 | 338 | 0.5 | 179 | 0.3 | 69 | 0.1 | 461 | 0.4 | 134 | 0.2 | 163 | 0.1 | 345 | 0.3 | 402 | 0.4 | 3297 | 0.4 |
| Smallmouth <br> Bass | 455 | 1.1 | 881 | 1.8 | 1095 | 1.5 | 1003 | 1.5 | 608 | 1.2 | 1081 | 1.0 | 666 | 0.8 | 536 | 0.4 | 548 | 0.4 | 491 | 0.4 | 7364 | 0.9 |
| Pumpkinseed | 446 | 1.1 | 306 | 0.6 | 848 | 1.2 | 228 | 0.4 | 104 | 0.2 | 1013 | 0.9 | 402 | 0.5 | 490 | 0.3 | 135 | 0.1 | 115 | 0.1 | 4087 | 0.5 |
| White Sucker | 1145 | 2.8 | 1394 | 2.8 | 582 | 0.8 | 412 | 0.6 | 109 | 0.2 | 776 | 0.7 | 853 | 1.0 | 263 | 0.2 | 540 | 0.4 | 410 | 0.4 | 6484 | 0.8 |
| Golden Shiner | 35 | 0.1 | 155 | 0.3 | 92 | 0.1 | 216 | 0.3 | 8 | 0.0 | 292 | 0.3 | 23 | 0.0 | 40 | 0.0 | 28 | 0.0 | 5 | 0.0 | 894 | 0.1 |


| Year | 1980 | 1980 | 1981 | 1981 | 1982 | 1982 | 1983 | 1983 | 1984 | 1984 | 1985 | 1985 | 1986 | 1986 | 1987 | 1987 | 1988 | 1988 | 1989 | 1989 | Totals 1980-1989 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 30 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \end{array}$ | 37 |  | 44 |  | 29 |  | 34 |  | 55 |  | 59 |  | 60 |  | 63 |  | 51 |  | 462 |  |
| Lifts | 403 |  | 490 |  | 725 |  | 648 |  | 519 |  | 1118 |  | 831 |  | 1414 |  | 1330 |  | 1117 |  | 8595 |  |
| Est. Oper. Time(HR) | 221 |  | 275 |  | 502 |  | 299 |  | 251 |  | 542 |  | 546 |  | 639 |  | 637 |  | 539 |  | 4451 |  |
| Fishing Time (HR) | 117 |  | 178 |  | 336 |  | 224 |  | 192 |  | 421 |  | 449 |  | 532 |  | 513 |  | 457 |  | 3419 |  |
| \#Species | 42 | Fish per lift | 48 | Fish per lift | 46 | Fish per lift | 40 | Fish per <br> lift | 35 | Fish per lift | 41 | Fish per lift | 43 | Fish per lift | 46 | Fish per lift | 49 | Fish per lift | 45 | Fish per lift |  | Fish per lift |
| Brown Trout | 258 | 0.6 | 207 | 0.4 | 219 | 0.3 | 225 | 0.3 | 141 | 0.3 | 175 | 0.2 | 65 | 0.1 | 83 | 0.1 | 85 | 0.1 | 110 | 0.1 | 1568 | 0.2 |
| Rock Bass | 88 | 0.2 | 381 | 0.8 | 138 | 0.2 | 269 | 0.4 | 158 | 0.3 | 122 | 0.1 | 200 | 0.2 | 231 | 0.2 | 110 | 0.1 | 352 | 0.3 | 2049 | 0.2 |
| Hickory Shad | 1 | 0.0 | 1 | 0.0 | 15 | 0.0 | 5 | 0.0 | 6 | 0.0 | 9 | 0.0 | 45 | 0.1 | 35 | 0.0 | 64 | 0.0 | 28 | 0.0 | 209 | 0.0 |
| Atlantic Menhaden | 16 | 0.0 | 42 | 0.1 | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 60 | 0.0 |
| $\begin{aligned} & \text { Largemouth } \\ & \text { Bass } \\ & \hline \end{aligned}$ | 41 | 0.1 | 13 | 0.0 | 20 | 0.0 | 17 | 0.0 | 8 | 0.0 | 67 | 0.1 | 75 | 0.1 | 69 | 0.0 | 117 | 0.1 | 164 | 0.1 | 591 | 0.1 |
| Yellow Bullhead | 18 | 0.0 | 36 | 0.1 | 61 | 0.1 | 10 | 0.0 | 7 | 0.0 | 21 | 0.0 | 35 | 0.0 | 41 | 0.0 | 80 | 0.1 | 445 | 0.4 | 754 | 0.1 |
| Black Crappie | 15 | 0.0 | 20 | 0.0 | 39 | 0.1 | 46 | 0.1 | 6 | 0.0 | 45 | 0.0 | 51 | 0.1 | 19 | 0.0 | 42 | 0.0 | 45 | 0.0 | 328 | 0.0 |
| Rainbow <br> Trout | 23 | 0.1 | 219 | 0.4 | 20 | 0.0 | 2 | 0.0 | 5 | 0.0 | 70 | 0.1 | 9 | 0.0 | 14 | 0.0 | 10 | 0.0 | 4 | 0.0 | 376 | 0.0 |
| Sea Lamprey | 1 | 0.0 | 55 | 0.1 | 56 | 0.1 | 10 | 0.0 | 4 | 0.0 | 164 | 0.1 | 26 | 0.0 | 21 | 0.0 | 59 | 0.0 | 94 | 0.1 | 490 | 0.1 |
| Green Sunfish | 16 | 0.0 | 28 | 0.1 | 91 | 0.1 | 16 | 0.0 | 7 | 0.0 | 133 | 0.1 | 15 | 0.0 | 64 | 0.0 | 19 | 0.0 | 33 | 0.0 | 422 | 0.0 |
| Tiger Muskie | 34 | 0.1 | 53 | 0.1 | 56 | 0.1 | 16 | 0.0 | 10 | 0.0 | 73 | 0.1 | 35 | 0.0 | 30 | 0.0 | 20 | 0.0 | 33 | 0.0 | 360 | 0.0 |
| Muskellunge | 27 | 0.1 | 1 | 0.0 | 4 | 0.0 | -- |  | -- |  | 15 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | 48 | 0.0 |
| Flathead |  |  |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Shiners | -- |  | -- |  | 6 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 6 | 0.0 |
| Atlantic Needlefish | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 | -- |  | 4 | 0.0 |
| Brook Trout | 4 | 0.0 | 3 | 0.0 | 5 | 0.0 | 2 | 0.0 | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | 1 | 0.0 | 17 | 0.0 |
| Bluntnose Minnow | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 65 | 0.0 | -- |  | 65 | 0.0 |
| $\begin{array}{\|l\|} \hline \text { Northern } \\ \text { Hog Sucker } \\ \hline \end{array}$ | 13 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 | 4 | 0.0 | 1 | 0.0 | 1 | 0.0 | 22 | 0.0 |
| Goldfish | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 |  | 1 | 0.0 | 3 | 0.0 |
| Tessellated Darter | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 | 1 | 0.0 | -- |  | 5 | 0.0 |
| Logperch | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 1 | 0.0 | 1 | 0.0 | 2 | 0.0 | -- |  | 5 | 0.0 |
| Brook Trout x Lake Trout | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 | -- |  | 2 | 0.0 | 5 | 0.0 | -- |  | 1 | 0.0 | 10 | 0.0 |
| Northern Pike | 3 | 0.0 | -- |  | 5 | 0.0 | 1 | 0.0 | -- |  | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | 11 | 0.0 |
| Striped Bass <br> x White <br> Perch | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 10 | 0.0 | 19 | 0.0 | 1 | 0.0 | 3 | 0.0 | 33 | 0.0 |
| Creek Chubsucker | -- |  | 4 | 0.0 | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 | 1 | 0.0 | -- |  | 12 | 0.0 |


| Year | 1980 | 1980 | 1981 | 1981 | 1982 | 1982 | 1983 | 1983 | 1984 | 1984 | 1985 | 1985 | 1986 | 1986 | 1987 | 1987 | 1988 | 1988 | 1989 | 1989 | Totals 1980-1989 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 30 | Fish per lift | 37 |  | 44 |  | 29 |  | 34 |  | 55 |  | 59 |  | 60 |  | 63 |  | 51 |  | 462 |  |
| Lifts | 403 |  | 490 |  | 725 |  | 648 |  | 519 |  | 1118 |  | 831 |  | 1414 |  | 1330 |  | 1117 |  | 8595 |  |
| Est. Oper. Time(HR) | 221 |  | 275 |  | 502 |  | 299 |  | 251 |  | 542 |  | 546 |  | 639 |  | 637 |  | 539 |  | 4451 |  |
| Fishing Time (HR) | 117 |  | 178 |  | 336 |  | 224 |  | 192 |  | 421 |  | 449 |  | 532 |  | 513 |  | 457 |  | 3419 |  |
| \#Species | 42 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \hline \end{gathered}$ | 48 | Fish per lift | 46 | Fish per <br> lift | 40 | Fish per lift | 35 | Fish per lift | 41 | Fish per <br> lift | 43 | Fish per lift | 46 | Fish per lift | 49 | Fish per lift | 45 | Fish per lift |  | Fish per lift |
| Chain Pickerel | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 2 | 0.0 |
| Margined Madtom | -- |  | -- |  | 7 | 0.0 | -- |  | -- |  | -- |  | 3 |  | -- |  | 1 | 0.0 | -- |  | 11 | 0.0 |
| Banded Darter | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 |
| Rosyface Shiner | -- |  | -- |  | 8 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 8 | 0.0 |
| Swallowtail <br> Shiner | -- |  | 3 | 0.0 | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | 4 | 0.0 |
| Shield Darter | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Greenside Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Longnose <br> Dace | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Tadpole Madtom | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 2 | 0.0 |
| Trouts | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Sunfishes | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Trout | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Rainbow <br> Smelt | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 1 | 0.0 | -- |  | 2 | 0.0 |
| Blacknose Dace | -- |  | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Mummichog | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Lampreys | -- |  | -- |  | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Lake Herring | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Striped Mullet | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 | 2 | 0.0 |
| Sunfish Hybrids |  |  |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Palomino (Rainbow <br> Trout) | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 |
| Redfin Pickerel | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 |
| Carps and Minnows | -- |  | -- |  | 1 |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| River Chub | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Creek Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 |
| Madtoms | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| Bigmouth Buffalo | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 1 | 0.0 |


| Year | 1980 | 1980 | 1981 | 1981 | 1982 | 1982 | 1983 | 1983 | 1984 | 1984 | 1985 | 1985 | 1986 | 1986 | 1987 | 1987 | 1988 | 1988 | 1989 | 1989 | Totals 1980-1989 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 30 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 37 |  | 44 |  | 29 |  | 34 |  | 55 |  | 59 |  | 60 |  | 63 |  | 51 |  | 462 |  |
| Lifts | 403 |  | 490 |  | 725 |  | 648 |  | 519 |  | 1118 |  | 831 |  | 1414 |  | 1330 |  | 1117 |  | 8595 |  |
| $\begin{array}{\|l\|} \hline \text { Est. Oper. } \\ \text { Time(HR) } \\ \hline \end{array}$ | 221 |  | 275 |  | 502 |  | 299 |  | 251 |  | 542 |  | 546 |  | 639 |  | 637 |  | 539 |  | 4451 |  |
| Fishing Time (HR) | 117 |  | 178 |  | 336 |  | 224 |  | 192 |  | 421 |  | 449 |  | 532 |  | 513 |  | 457 |  | 3419 |  |
| \#Species | 42 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 48 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \\ & \hline \hline \end{aligned}$ | 46 | Fish per lift | 40 | Fish per lift | 35 | Fish per lift | 41 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \\ & \hline \hline \end{aligned}$ | 43 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \\ & \hline \end{aligned}$ | 46 | $\begin{aligned} & \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ & \hline \end{aligned}$ | 49 | $\begin{aligned} & \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ & \hline \end{aligned}$ | 45 | Fish per lift |  | Fish per lift |
| Total | 372379 | 924.0 | 1353310 | 2761.9 | 1403176 | 1935.4 | 1028092 | 1586.6 | 957821 | 1845.5 | 2317797 | 2073.2 | 1830569 | 2202.9 | 2593445 | 1834.1 | 1592938 | 1197.7 | 1035121 | 926.7 | 14484648 | 1685.2 |

## A-25 Cont.

| Year | 1990 | 1990 | 1991 | 1991 | 1992 | 1992 | 1993 | 1993 | 1994 | 1994 | 1995 | 1995 | 1996 | 1996 | 1997 | 1997 | 1998 | 1998 | 1999 | 1999 | Totals 1990-1999 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 64 |  | 63 |  | 64 |  | 45 |  | 47 |  | 68 |  | 28 |  | 44 |  | 41 |  | 43 |  | 507 |  |
| Lifts | 1363 |  | 1257 |  | 1559 |  | 1032 |  | 964 |  | 1245 |  | 464 |  | 611 |  | 476 |  | 709 |  | 9680 |  |
| Est. Oper. | 664 |  | 681 |  | 698 |  | 505.4 |  | 534.8 |  | 744.3 |  | 284.6 |  | 348.6 |  | 238.6 |  | 314.9 |  | 5014.2 |  |
| Fishing Time (HR) | 571 |  | 547 |  | 589 |  | 416.7 |  | 441.1 |  | 651.9 |  | 259.2 |  | 295.1 |  | 225.9 |  | 312.6 |  | 4309.5 |  |
| \#Species | 43 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 45 | $\begin{array}{\|l} \hline \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \end{array}$ | 46 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 37 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 46 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 44 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 39 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 34 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ |  | Fish per lift |
| Gizzard Shad | 1084073 | 795.4 | 433108 | 344.6 | 1450299 | 930.3 | 666010 | 645.4 | 511139 | 530.2 | 799694 | 642.3 | 196019 | 422.5 | 126570 | 207.2 | 497375 | 1044.9 | 652770 | 920.7 | 6417057 | 662.9 |
| White Perch | 24581 | 18.0 | 14996 | 11.9 | 37521 | 24.1 | 3892 | 3.8 | 9537 | 9.9 | 55719 | 44.8 | 4583 | 9.9 | 58685 | 96.0 | 32891 | 69.1 | 35357 | 49.9 | 277762 | 28.7 |
| Blueback Herring | 9658 | 7.1 | 15616 | 12.4 | 27533 | 17.7 | 4052 | 3.9 | 2603 | 2.7 | 93859 | 75.4 | 871 | 1.9 | 133257 | 218.1 | 5511 | 11.6 | 8546 | 12.1 | 301506 | 31.1 |
| Chanring Catfish | 8689 | 6.4 | 10252 | 8.2 | 7070 | 4.5 | 10841 | 10.5 | 3551 | 3.7 | 2432 | 2.0 | 5487 | 11.8 | 977 | 1.6 | 17250 | 36.2 | 2564 | 3.6 | 69113 | 7.1 |
| Common Carp | 2761 | 2.0 | 8257 | 6.6 | 4105 | 2.6 | 8488 | 8.2 | 7403 | 7.7 | 6209 | 5.0 | 5726 | 12.3 | 2281 | 3.7 | 8206 | 17.2 | 5124 | 7.2 | 58560 | 6.0 |
| American Eel | 224 | 0.2 | 213 | 0.2 | 2622 | 1.7 | 1487 | 1.4 | 128 | 0.1 | 204 | 0.2 | 640 | 1.4 | 110 | 0.2 | 89 | 0.2 | 234 | 0.3 | 5951 | 0.6 |
| Alewife | 425 | 0.3 | 2649 | 2.1 | 3344 | 2.1 | 572 | 0.6 | 70 | 0.1 | 5405 | 4.3 | , | 0.0 | 11 | 0.0 | 31 | 0.1 | 1795 | 2.5 | 14303 | 1.5 |
| $\begin{aligned} & \text { American } \\ & \text { Shad } \end{aligned}$ | 15719 | 11.5 | 13330 | 10.6 | 10335 | 6.6 | 5343 | 5.2 | 5615 | 5.8 | 15588 | 12.5 | 11473 | 24.7 | 12974 | 21.2 | 6577 | 13.8 | 9658 | 13.6 | 106612 | 11.0 |
| Comely Shiner | 5798 | 4.3 | 18356 | 14.6 | 8974 | 5.8 | 7358 | 7.1 | 13973 | 14.5 | 1746 | 1.4 | 2180 | 4.7 | 576 | 0.9 | 570 | 1.2 | 515 | 0.7 | 60046 | 6.2 |
| Quillback | 1270 | 0.9 | 2990 | 2.4 | 132 | 0.1 | 746 | 0.7 | 1576 | 1.6 | 981 | 0.8 | 583 | 1.3 | 780 | 1.3 | 280 | 0.6 | 823 | 1.2 | 10161 | 1.0 |
| Shorthead Redhorse | 4228 | 3.1 | 2871 | 2.3 | 1813 | 1.2 | 858 | 0.8 | 1994 | 2.1 | 2098 | 1.7 | 754 | 1.6 | 3134 | 5.1 | 357 | 0.8 | 1485 | 2.1 | 19592 | 2.0 |
| Spotfin Shiner | 135 | 0.1 | 2508 | 2.0 | 214 | 0.1 | 10 | 0.0 | 13 | 0.0 | 279 | 0.2 | 10 | 0.0 | 1 | 0.0 | 79 | 0.2 | -- |  | 3249 | 0.3 |
| Striped Bass | 1068 | 0.8 | 1682 | 1.3 | 2094 | 1.3 | 1595 | 1.5 | 4261 | 4.4 | 5467 | 4.4 | 1845 | 4.0 | 2665 | 4.4 | 2570 | 5.4 | 1001 | 1.4 | 24248 | 2.5 |
| Spottail Shiner | 112 | 0.1 | 635 | 0.5 | 156 | 0.1 |  |  | 6 | 0.0 | 249 | 0.2 | -- |  | 1041 | 1.7 | -- |  | -- |  | 2199 | 0.2 |
| Redbreast Sunfish | 187 | 0.1 | 281 | 0.2 | 154 | 0.1 | 170 | 0.2 | 165 | 0.2 | 1045 | 0.8 | 179 | 0.4 | 430 | 0.7 | 259 | 0.5 | 123 | 0.2 | 2993 | 0.3 |
| White Catfish | 560 | 0.4 | 1284 | 1.0 | 152 | 0.1 | 97 | 0.1 | 187 | 0.2 | 403 | 0.3 | 293 | 0.6 | 140 | 0.2 | 216 | 0.5 | -- |  | 3332 | 0.3 |
| Walleye | 460 | 0.3 | 411 | 0.3 | 203 | 0.1 | 217 | 0.2 | 653 | 0.7 | 1736 | 1.4 | 964 | 2.1 | 1063 | 1.7 | 827 | 1.7 | 547 | 0.8 | 7081 | 0.7 |
| Bluegill | 446 | 0.3 | 486 | 0.4 | 813 | 0.5 | 200 | 0.2 | 244 | 0.3 | 505 | 0.4 | 158 | 0.3 | 277 | 0.5 | 381 | 0.8 | 605 | 0.9 | 4115 | 0.4 |
| Striped Bass x White Bass | 1172 | 0.9 | 797 | 0.6 | 359 | 0.2 | 112 | 0.1 | 114 | 0.1 | 28 | 0.0 | 4 | 0.0 | 1 | 0.0 | 18 | 0.0 | 2 | 0.0 | 2607 | 0.3 |
| White Crappie | 33 | 0.0 | 106 | 0.1 | 74 | 0.0 | 62 | 0.1 | 36 | 0.0 | 33 | 0.0 | 25 | 0.1 | 30 | 0.0 | 119 | 0.3 | 72 | 0.1 | 590 | 0.1 |
| Yellow Perch | 124 | 0.1 | 502 | 0.4 | 127 | 0.1 | 318 | 0.3 | 48 | 0.0 | 462 | 0.4 | 180 | 0.4 | 102 | 0.2 | 109 | 0.2 | 134 | 0.2 | 2106 | 0.2 |
| Brown Bullhead | 108 | 0.1 | 260 | 0.2 | 107 | 0.1 | 73 | 0.1 | 9 | 0.0 | 281 | 0.2 | 54 | 0.1 | 27 | 0.0 | 398 | 0.8 | 8 | 0.0 | 1325 | 0.1 |
| Smallmouth <br> Bass | 424 | 0.3 | 704 | 0.6 | 411 | 0.3 | 227 | 0.2 | 132 | 0.1 | 362 | 0.3 | 232 | 0.5 | 251 | 0.4 | 812 | 1.7 | 1306 | 1.8 | 4861 | 0.5 |
| Pumpkinseed | 46 | ${ }_{0} 0.0$ | 48 | 0.0 | 118 | ${ }_{0}^{0.1}$ | $\stackrel{22}{59}$ | ${ }_{0} 0.0$ | 22 | 0.0 | 53 | ${ }_{0}^{0.0}$ | 11 | 0.0 | 51 | ${ }_{0}^{0.1}$ | 15 | 0.0 | 13 | 0.0 | 399 | 0.0 |
| White Sucker <br> Golden Shiner | 161 | 0.1 0.0 | 113 7 | 0.1 0.0 | $\frac{83}{11}$ | 0.1 0.0 | 59 | 0.1 | 36 | 0.0 0.0 | $\frac{105}{87}$ | 0.1 0.1 | 20 | 0.0 | 7 | 0.0 | 14 | 0.0 | 27 | 0.0 | 625 | 0.1 0.0 |


| Year | 1990 | 1990 | 1991 | 1991 | 1992 | 1992 | 1993 | 1993 | 1994 | 1994 | 1995 | 1995 | 1996 | 1996 | 1997 | 1997 | 1998 | 1998 | 1999 | 1999 | Totals 1990-1999 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 64 |  | 63 |  | 64 |  | 45 |  | 47 |  | 68 |  | 28 |  | 44 |  | 41 |  | 43 |  | 507 |  |
| Lifts | 1363 |  | 1257 |  | 1559 |  | 1032 |  | 964 |  | 1245 |  | 464 |  | 611 |  | 476 |  | 709 |  | 9680 |  |
| Est. Oper. Time(HR) | 664 |  | 681 |  | 698 |  | 505.4 |  | 534.8 |  | 744.3 |  | 284.6 |  | 348.6 |  | 238.6 |  | 314.9 |  | 5014.2 |  |
| Fishing Time (HR) | 571 |  | 547 |  | 589 |  | 416.7 |  | 441.1 |  | 651.9 |  | 259.2 |  | 295.1 |  | 225.9 |  | 312.6 |  | 4309.5 |  |
| \#Species | 43 | $\begin{aligned} & \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ & \hline \hline \end{aligned}$ | 45 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 46 | $\begin{array}{\|l} \hline \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \end{array}$ | 37 | $\begin{array}{\|l} \hline \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ \hline \end{array}$ | 46 | $\begin{aligned} & \begin{array}{l} \text { Fish per } \\ \text { lift } \end{array} \\ & \hline \hline \end{aligned}$ | 44 | $\begin{array}{\|l\|} \hline \text { Fish per } \\ \text { lift } \end{array}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \\ & \hline \hline \end{aligned}$ | 39 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \\ & \hline \end{aligned}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 34 | $\begin{aligned} & \hline \text { Fish per } \\ & \text { lift } \end{aligned}$ |  | Fish per lift |
| Brown Trout | 63 | 0.0 | 82 | 0.1 | 127 | 0.1 | 98 | 0.1 | 48 | 0.0 | 22 | 0.0 | 27 | 0.1 | 14 | 0.0 | 30 | 0.1 | 8 | 0.0 | 519 | 0.1 |
| Rock Bass | 39 | 0.0 | 53 | 0.0 | 106 | 0.1 | 90 | 0.1 | 34 | 0.0 | 83 | 0.1 | 149 | 0.3 | 280 | 0.5 | 126 | 0.3 | 68 | 0.1 | 1028 | 0.1 |
| Hickory Shad | 77 | 0.1 | 120 | 0.1 | 367 | 0.2 | -- |  | ${ }^{1}$ | 0.0 | 36 | 0.0 | -- |  | 118 | 0.2 | 6 | 0.0 | 32 | ${ }^{0.0}$ | 757 | 0.1 |
| Atlantic <br> Menhaden | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Largemouth | 48 | 0.0 | 176 | 0.1 | 211 | 0.1 | 84 | 0.1 | 78 | 0.1 | 174 | 0.1 | 23 | 0.0 | 55 | 0.1 | 49 | 0.1 | 78 | 0.1 | 976 | 0.1 |
| Yellow Bullhead | 32 | 0.0 | 25 | 0.0 | 23 | 0.0 | 19 | 0.0 | 17 | 0.0 | 16 | 0.0 | 22 | 0.0 | 37 | 0.1 | 19 | 0.0 | ${ }^{3}$ | 0.0 | 213 | 0.0 |
| Black Crappie | 22 | 0.0 | 22 | 0.0 | 23 | 0.0 | 7 | 0.0 | 8 | 0.0 | 24 | 0.0 | 9 | 0.0 | 19 | 0.0 | 6 | 0.0 | 10 | ${ }^{0.0}$ | 150 | 0.0 |
| Rainbow Trout | 14 | 0.0 | 13 | 0.0 | 12 | 0.0 | 4 | 0.0 | 3 | 0.0 | 6 | 0.0 | 12 | 0.0 | 1 | 0.0 | 4 | 0.0 | -- |  | 69 | 0.0 |
| Sea Lamprey | 38 | 0.0 | 34 | 0.0 | 42 | 0.0 | 5 | 0.0 | 11 | 0.0 | 7 | 0.0 | 10 | 0.0 | 2 | 0.0 | 7 | 0.0 | 24 | 0.0 | 180 | 0.0 |
| Green Sunfish | 17 | 0.0 | 22 | 0.0 | 35 | 0.0 | 10 | 0.0 | 11 | 0.0 | 20 | 0.0 | 18 | 0.0 | 6 | 0.0 | 10 | 0.0 | 7 | 0.0 | 156 | 0.0 |
| Tiger Muskie | 10 | 0.0 | 5 | ${ }^{0.0}$ | 3 | ${ }^{0.0}$ | 2 | 0.0 | 1 | 0.0 | 2 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | 24 | 0.0 |
| Muskellunge | 2 | ${ }^{0.0}$ | 2 | ${ }^{0.0}$ | 10 | ${ }^{0.0}$ | 9 | ${ }^{0.0}$ | 11 | ${ }^{0.0}$ | ${ }^{4}$ | ${ }^{0.0}$ | ${ }^{4}$ | ${ }^{0.0}$ | 2 | ${ }^{0.0}$ | 2 | ${ }^{0.0}$ | -- |  | 46 | 0.0 |
| $\begin{aligned} & \text { Flathead } \\ & \text { Catfish } \end{aligned}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Shiners | -- |  | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 |
| Atlantic Needlefish | ${ }^{5}$ | 0.0 | -- |  | 3 | 0.0 | -- |  | 8 | 0.0 | 1 | 0.0 | 1 | 0.0 | 1 | 0.0 | 4 | 0.0 | 2 | 0.0 | 25 | 0.0 |
| Brook Trout | -- |  | 7 | 0.0 | 5 | 0.0 | -- |  | 4 | 0.0 | 2 | 0.0 | 3 | 0.0 | -- |  | -- |  | -- |  | 21 | 0.0 |
| Bluntnose Minnow | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 0 | 0.0 |
| Northern Hog Sucker | 3 | 0.0 | -- |  | 5 | 0.0 | -- |  | 5 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | 13 | 0.0 |
| Goldfish | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Tessellated <br> Darter | -- |  | 6 | 0.0 | ${ }^{2}$ | 0.0 | -- |  | 1 | 0.0 | 5 | 0.0 |  |  | 5 | 0.0 |  |  | -- |  | 19 | 0.0 |
| Logperch | 2 | ${ }^{0.0}$ | 1 | ${ }^{0.0}$ | ${ }^{2}$ | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 |
| Brook Trout x Lake Trout | -- |  | -- |  | 1 | 0.0 | 5 | 0.0 | 2 | 0.0 |  |  | 4 | 0.0 | -- |  | -- |  | 2 | 0.0 | 14 | 0.0 |
| Northern Pike | -- |  | 5 | 0.0 | -- |  |  |  | 1 | 0.0 | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | 8 | 0.0 |
| Striped Bass X White Perch | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chubsucker | 1 | 0.0 | -- |  | 9 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 10 | 0.0 |


| Year | 1990 | 1990 | 1991 | 1991 | 1992 | 1992 | 1993 | 1993 | 1994 | 1994 | 1995 | 1995 | 1996 | 1996 | 1997 | 1997 | 1998 | 1998 | 1999 | 1999 | Totals 1990-1999 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 64 |  | 63 |  | 64 |  | 45 |  | 47 |  | 68 |  | 28 |  | 44 |  | 41 |  | 43 |  | 507 |  |
| Lifts | 1363 |  | 1257 |  | 1559 |  | 1032 |  | 964 |  | 1245 |  | 464 |  | 611 |  | 476 |  | 709 |  | 9680 |  |
| Est. Oper. Time(HR) | 664 |  | 681 |  | 698 |  | 505.4 |  | 534.8 |  | 744.3 |  | 284.6 |  | 348.6 |  | 238.6 |  | 314.9 |  | 5014.2 |  |
| Fishing Time (HR) | 571 |  | 547 |  | 589 |  | 416.7 |  | 44.1 |  | 651.9 |  | 259.2 |  | 295.1 |  | 22.9 |  | 312.6 |  | 4309.5 |  |
| \#Species | 43 | $\begin{array}{\|l} \hline \text { Fish per } \\ \text { lift } \end{array}$ | 45 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 46 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 37 | Fish per <br> lift | 46 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 44 | $\begin{array}{\|l\|l\|} \hline \text { Fish per } \\ \text { lift } \end{array}$ | 38 | Fish per <br> lift | 39 | Fish per <br> lift | 38 | Fish per lift | 34 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ |  | Fish per lift |
| Chain Pickerel | -- |  | 6 | 0.0 | 2 | 0.0 | -- |  | 3 | 0.0 | 4 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | 16 | 0.0 |
| Margined Madtom | -- |  | -- |  | -- |  | 12 | 0.0 | 3 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 16 | 0.0 |
| Banded Darter | 2 | 0.0 | 10 | 0.0 | -- |  | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | 14 | 0.0 |
| Rosyface <br> Shiner | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Swallowtail Shiner | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Shield Darter | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Greenside <br> Darter |  |  |  |  | -- |  | -- |  | -- |  | 4 | 0.0 | -- |  | 1 | 0.0 | -- |  | -- |  | 5 | 0.0 |
| $\begin{aligned} & \text { Longnose } \\ & \text { Dace } \end{aligned}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Tadpole Madtom | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 2 | 0.0 |
| Trouts | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Sunfishes | 2 | 0.0 | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 3 | 0.0 |
| Trout | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Rainbow <br> Smelt | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Blacknose Dace | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Mummichog | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Lampreys | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Lake Herring | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Striped | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Sunfish Hybrids | -- |  | -- |  | 1 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Palomino <br> (Rainbow <br> Trout) | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Redfin Pickerel | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{aligned} & \text { Carps and } \\ & \text { Minnows } \\ & \hline \end{aligned}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| River Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Madtoms | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Bigmouth Buffalo | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Total | 1162841 | 853.1 | 533052 | 424.1 | \#\#\#\#\# | 1000.5 | 713155 | 691.0 | 563773 | 584.8 | 995447 | 799.6 | 232615 | 501.3 | 345983 | 566.3 | 575220 | 1208.4 | 722945 | 1019.7 | 7404845 | 765.0 |


| Year | 2000 | 2000 | 2001 | 2001 | 2002 | 2002 | 2003 | 2003 | 2004 | 2004 | 2005 | 2005 | 2006 | 2006 | 2007 | 2007 | 2008 | 2008 | 2009 | 2009 | $\begin{array}{r} \text { Totals } \\ 2000-2009 \end{array}$ | $\begin{array}{r} \hline \text { Totals } \\ \text { 1972- } \\ 2009 \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 34 |  | 41 |  | 31 |  | 31 |  | 14 |  | 30 |  | 37 |  | 29 |  | 34 |  | 28 |  | 309 | 1695 |
| Lifts | 424 |  | 425 |  | 417 |  | 367 |  | 151 |  | 295 |  | 349 |  | 288 |  | 481 |  | 282 |  | 3479 | 27481 |
| $\begin{aligned} & \hline \text { Est. Oper. } \\ & \text { Time(HR) } \\ & \hline \end{aligned}$ |  |  |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 0 |  |
| Fishing Time (HR) | 206 |  | 195 |  | 147 |  | 171 |  | 74 |  | 166 |  | 215 |  | 135 |  | 174 |  | 144 |  | 1627 | 11436 |
| \#Species | 37 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 35 | $\begin{array}{r} \text { Fish per } \\ \text { lift } \end{array}$ | 30 | $\begin{array}{r} \text { Fish per } \\ \text { lift } \end{array}$ | 30 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 36 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 38 | $\begin{array}{r} \text { Fish per } \\ \text { lift } \end{array}$ | 35 | $\begin{array}{r} \text { Fish per } \\ \text { lift } \end{array}$ | 37 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 39 | $\begin{array}{r} \text { Fish per } \\ \text { lift } \\ \hline \end{array}$ |  |  |
| Gizzard Shad | 366099 | 863.4 | 218124 | 513.2 | 339292 | 813.6 | 118852 | 323.8 | 22899 | 151.6 | 82412 | 279.4 | 149250 | 427.7 | 146821 | 509.8 | 724737 | 1506.7 | 210633 | 746.9 | 2379119 | 683.9 |
| White Perch | 40318 | 95.1 | 44364 | 104.4 | 65031 | 155.9 | 14476 | 39.4 | 976 | 6.5 | 1102 | 3.7 | 1001 | 2.9 | 2276 | 7.9 | 2036 | 4.2 | 3095 | 11.0 | 174675 | 50.2 |
| Blueback Herring | 14326 | 33.8 | 16320 | 38.4 | 428 | 1.0 | 183 | 0.5 | 1 | 0.0 | 0 |  | 6 | 0.0 | 153 | 0.5 | 7 | 0.0 | 165 | 0.6 | 31589 | 9.1 |
| Channel | 8394 | 19.8 | 228 | 0.5 | 844 | 2.0 | 626 | 1.7 | 4839 | 32.0 | 1692 | 5.7 | 2880 | 8.3 | 1480 | 5.1 | 781 | 1.6 | 2393 | 8.5 | 24157 | 6.9 |
| $\begin{aligned} & \text { Common } \\ & \text { Carp } \end{aligned}$ | 3236 | 7.6 | 994 | 2.3 | 225 | 0.5 | 1110 | 3.0 | 2702 | 17.9 | 1179 | 4.0 | 716 | 2.1 | 372 | 1.3 | 400 | 0.8 | 399 | 1.4 | 11333 | 3.3 |
| $\begin{aligned} & \text { American } \\ & \text { Eel } \end{aligned}$ | 735 | 1.7 | 437 | 1.0 | 144 | 0.3 | 20 | 0.1 | 61 | 0.4 | 25 | 0.1 | 12 | 0.0 | 27 | 0.1 | 26 | 0.1 | 37 | 0.1 | 1524 | 0.4 |
| Alewife | 9189 | 21.7 | 7824 | 18.4 | 141 | 0.3 | 16 | 0.0 | -- |  | 0 |  | 2 | 0.0 | 7 | 0.0 | 2 | 0.0 | 20 | 0.1 | 17201 | 4.9 |
| American <br> Shad | 9785 | 23.1 | 10940 | 25.7 | 9347 | 22.4 | 9802 | 26.7 | 3426 | 22.7 | 3896 | 13.2 | 3970 | 11.4 | 4272 | 14.8 | 2627 | 5.5 | 6534 | 23.2 | 64599 | 18.6 |
| Comely <br> Shiner | 1 | 0.0 | 1228 | 2.9 | 2 | 0.0 | 22 | 0.1 | 67 | 0.4 | 226 | 0.8 | 548 | 1.6 | 45 | 0.2 | 27 | 0.1 | 1 | 0.0 | 2167 | 0.6 |
| Quillback | 154 | 0.4 | 76 | 0.2 | 13 | 0.0 | 91 | 0.2 | 52 | 0.3 | 848 | 2.9 | 289 | 0.8 | 73 | 0.3 | 52 | 0.1 | 2 | 0.0 | 1650 | 0.5 |
| Shorthead Redhorse | 1317 | 3.1 | 132 | 0.3 | 317 | 0.8 | 749 | 2.0 | 1 | 0.0 | 863 | 2.9 | 109 | 0.3 | 144 | 0.5 | 325 | 0.7 | 92 | 0.3 | 4049 | 1.2 |
| Spotfin Shiner | 32 | 0.1 | 237 | 0.6 | -- |  | -- |  | 15 | 0.1 | 230 | 0.8 | 1 | 0.0 | -- |  | 83 | 0.2 | 6 | 0.0 | 604 | 0.2 |
| Striped Bass | 2453 | 5.8 | 710 | 1.7 | 2086 | 5.0 | 703 | 1.9 | 458 | 3.0 | 489 | 1.7 | 383 | 1.1 | 263 | 0.9 | 42 | 0.1 | 179 | 0.6 | 7766 | 2.2 |
| $\begin{aligned} & \hline \text { Spotait } \\ & \text { SSiner } \end{aligned}$ | -- |  | 5833 | 13.7 | 3 | 0.0 | -- |  | -- |  | 5 | 0.0 | 15 | 0.0 | 986 | 3.4 | 76 | 0.2 | 2 | 0.0 | 6920 | 2.0 |
| Redbreast Sunfish | 123 | 0.3 | 783 | 1.8 | 179 | 0.4 | 19 | 0.1 | 70 | 0.5 | 80 | 0.3 | 148 | 0.4 | 53 | 0.2 | 21 | 0.0 | 71 | 0.3 | 1547 | 0.4 |
| White Catfish | 351 | 0.8 | 36 | 0.1 | 49 | 0.1 | 7 | 0.0 | 271 | 1.8 | 24 | 0.1 | 9 | 0.0 | 5 | 0.0 | 5 | 0.0 | 5 | 0.0 | 762 | 0.2 |
| Walleye | 177 | 0.4 | 274 | 0.6 | 79 | 0.2 | 68 | 0.2 | 57 | 0.4 | 217 | 0.7 | 1962 | 5.6 | 1776 | 6.2 | 1971 | 4.1 | 977 | 3.5 | 7558 | 2.2 |
| Bluegill | 292 | 0.7 | 260 | 0.6 | 155 | 0.4 | 45 | 0.1 | 15 | 0.1 | 14 | 0.0 | 145 | 0.4 | 85 | 0.3 | 60 | 0.1 | 313 | 1.1 | 1384 | 0.4 |
| Striped Bass x White Bass | 1 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 | 3 | 0.0 | 3 | 0.0 | 2 | 0.0 | -- |  | -- |  | 10 | 0.0 |
| White Crappie | 30 | 0.1 | 29 | 0.1 | 6 | 0.0 | 5 | 0.0 | 1 | 0.0 | 1 | 0.0 | 6 | 0.0 | -- |  | 3 | 0.0 | 8 | 0.0 | 89 | 0.0 |
| $\begin{aligned} & \text { Yellow } \\ & \text { Perch } \\ & \hline \end{aligned}$ | 161 | 0.4 | 150 | 0.4 | 122 | 0.3 | 102 | 0.3 | 8 | 0.1 | 33 | 0.1 | 54 | 0.2 | 49 | 0.2 | 26 | 0.1 | 55 | 0.2 | 760 | 0.2 |
| Brown Bullhead | 94 | 0.2 | 136 | 0.3 | 26 | 0.1 | 104 | 0.3 | 1599 | 10.6 | 713 | 2.4 | 1060 | 3.0 | 237 | 0.8 | 51 | 0.1 | 198 | 0.7 | 4218 | 1.2 |
| Smallmouth Bass | 764 | 1.8 | 309 | 0.7 | 390 | 0.9 | 232 | 0.6 | 33 | 0.2 | 560 | 1.9 | 306 | 0.9 | 140 | 0.5 | 95 | 0.2 | 109 | 0.4 | 2938 | 0.8 |
| Pumpkinsee <br> d | 13 | 0.0 | 27 | 0.1 | 26 | 0.1 | 3 | 0.0 | 3 | 0.0 | 2 | 0.0 | 23 | 0.1 | 8 | 0.0 | 8 | 0.0 | 33 | 0.1 | 146 | 0.0 |
| White | 44 | 0.1 | 12 | 0.0 | 8 | 0.0 | 1 | 0.0 | 3 | 0.0 | 1 | 0.0 | 7 | 0.0 | 4 | 0.0 | 1 | 0.0 | 7 | 0.0 | 88 | 0.0 |
| Golden Shiner | -- |  | -- |  | ${ }^{1}$ | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 | 3 | 0.0 |


| Year | 2000 | 2000 | 2001 | 2001 | 2002 | 2002 | 2003 | 2003 | 2004 | 2004 | 2005 | 2005 | 2006 | 2006 | 2007 | 2007 | 2008 | 2008 | 2009 | 2009 | Totals 2000-2009 | $\begin{array}{r}\text { Totals 1972- } \\ 2009 \\ \hline\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 34 |  | 41 |  | 31 |  | 31 |  | 14 |  | 30 |  | 37 |  | 29 |  | 34 |  | 28 |  | 309 | 1695 |
| Lifts | 424 |  | 425 |  | 417 |  | 367 |  | 151 |  | 295 |  | 349 |  | 288 |  | 481 |  | 282 |  | 3479 | 27481 |
| Est. Oper. Time(HR) |  |  |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 0 |  |
| Fishing Time (HR) | 206 |  | 195 |  | 147 |  | 171 |  | 74 |  | 166 |  | 215 |  | 135 |  | 174 |  | 144 |  | 1627 | 11436 |
| \#Species | 37 | $\begin{array}{\|c}  \\ \hline \text { Fish per } \\ \text { lift } \end{array}$ | 38 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 35 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 30 | $\begin{gathered} \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 30 | $\begin{array}{\|c} \text { Fish per } \\ \text { lift } \end{array}$ | 36 | $\begin{array}{\|c} \text { Fish per } \\ \text { lift } \end{array}$ | 38 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 35 | $\begin{gathered} \text { Fish per } \\ \text { lift } \end{gathered}$ | 37 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 39 | Fish per lift |  |  |
| Brown Trout | 8 | 0.0 | -- |  | 7 | 0.0 | -- |  | -- |  | 12 | 0.0 | 11 | 0.0 | 6 | ${ }^{0.0}$ | 2 | ${ }^{0.0}$ | 8 | 0.0 | 54 | 0.0 |
| Rock Bass | 119 | 0.3 | 188 | 0.4 | 65 | 0.2 | 100 | 0.3 | 9 | 0.1 | 35 | 0.1 | 51 | 0.1 | 30 | 0.1 | 18 | 0.0 | 116 | 0.4 | 731 | 0.2 |
| Hickory <br> Shad | 1 | 0.0 | 36 | 0.1 |  |  | 1 | 0.0 | -- |  | 0 |  | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 4 | 0.0 | 42 | 0.0 |
| Atlantic Menhaden | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{aligned} & \text { Largemouth } \\ & \text { Bass } \end{aligned}$ | 63 | 0.1 | 14 | 0.0 | 12 | 0.0 | 32 | 0.1 | 9 | 0.1 | 33 | 0.1 | 21 | 0.1 | 34 | 0.1 | 4 | 0.0 | 6 | 0.0 | 228 | 0.1 |
| Yellow Bullhead | 16 | 0.0 | 12 | 0.0 | 2 | 0.0 | 3 | 0.0 | -- |  | 1 | 0.0 | 29 | 0.1 | -- |  | -- |  | -- |  | 63 | 0.0 |
| Black Crappie | 8 | 0.0 | 5 | 0.0 | 1 | 0.0 | 3 | 0.0 | 1 | 0.0 | 4 | 0.0 | 10 | 0.0 | 2 | 0.0 | 7 | 0.0 | 19 | 0.1 | 60 | 0.0 |
| Rainbow <br> Trout | 5 | 0.0 | 1 | 0.0 | 3 | 0.0 | -- |  | -- |  | -- |  | 3 | 0.0 | 1 | 0.0 | 1 | 0.0 | 5 | 0.0 | 19 | 0.0 |
| Sea <br> Lamprey | 11 | 0.0 | 43 | 0.1 | 75 | 0.2 | 7 | 0.0 | -- |  | 10 | 0.0 | 43 | 0.1 | 6 | 0.0 | 28 | 0.1 | 72 | 0.3 | 295 | 0.1 |
| Green Sunfish | 17 | 0.0 | 28 | 0.1 | 4 | 0.0 | -- |  | 2 | 0.0 | -- |  | 2 | 0.0 | 1 | 0.0 | 1 | 0.0 | 9 | 0.0 | 64 | 0.0 |
| Tiger Muskie | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 |
| Muskellunge | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 3 | 0.0 | 2 | 0.0 | 5 | 0.0 |
| $\qquad$ | -- |  | -- |  | -- |  | -- |  | 7 | 0.0 | 9 | 0.0 | 42 | 0.1 | 13 | 0.0 | 22 | 0.0 | 196 | 0.7 | 289 | 0.1 |
| Shiners | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | ${ }^{2}$ | 0.0 |
| Atlantic Needlefish | 7 | 0.0 | -- |  | 16 | 0.0 | -- |  | -- |  | 31 | 0.1 | 9 | 0.0 | 3 | 0.0 | -- |  | 17 | 0.1 | 83 | 0.0 |
| Brook Trout | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 2 | ${ }^{0.0}$ | 13 | 0.0 | 2 | ${ }^{0.0}$ | 2 | ${ }^{0.0}$ | 20 | 0.0 |
| Bluntnose <br> Minnow | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Northern <br> Hog Sucker | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | 3 | 0.0 |
| Goldfish | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 |
| Tessellated Darter | -- |  | 6 | 0.0 | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | 8 | 0.0 |
| Logperch | -- |  | 2 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |
| Brook Trout x Lake Trout | 2 | 0.0 | 3 | 0.0 | 3 | 0.0 | -- |  | -- |  | 3 | 0.0 | 2 | 0.0 | -- |  | -- |  | -- |  | 13 | 0.0 |
| Northern Pike | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | 1 | 0.0 | 2 | 0.0 |
| Striped Bass x White Perch | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chubsucker | -- |  | 1 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 2 | 0.0 |


| Year | 2000 | 2000 | 2001 | 2001 | 2002 | 2002 | 2003 | 2003 | 2004 | 2004 | 2005 | 2005 | 2006 | 2006 | 2007 | 2007 | 2008 | 2008 | 2009 | 2009 | Totals 2000-2009 | Totals $1972-$ <br> 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 34 |  | 41 |  | 31 |  | 31 |  | 14 |  | 30 |  | 37 |  | 29 |  | 34 |  | 28 |  | 309 |  |
| Lifts | 424 |  | 425 |  | 417 |  | 367 |  | 151 |  | 295 |  | 349 |  | 288 |  | 481 |  | 282 |  | 3479 |  |
| Est. Oper. Time(HR) |  |  |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | 0 |  |
| $\begin{array}{\|r} \hline \text { Fishing } \\ \text { Time (HR) } \\ \hline \end{array}$ | 206 |  | 195 |  | 147 |  | 171 |  | 74 |  | 166 |  | 215 |  | 135 |  | 174 |  | 144 |  | 1627 |  |
| \#Species | 37 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \\ \hline \hline \end{array}$ | 38 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \\ \hline \hline \end{array}$ | 35 | $\begin{gathered} \hline \text { Fish per } \\ \text { lift } \\ \hline \end{gathered}$ | 30 | Fish per lift | 30 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \\ \hline \hline \end{array}$ | 36 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \\ \hline \hline \end{array}$ | 38 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \end{array}$ | 35 | $\begin{aligned} & \text { Fish per } \\ & \text { lift } \end{aligned}$ | 37 | $\begin{array}{r} \hline \text { Fish per } \\ \text { lift } \\ \hline \hline \end{array}$ | 39 | Fish per lift |  |  |
| Chain Pickerel | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Margined Madtom | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | -- |  | 1 | 0.0 |
| $\begin{array}{\|l\|l\|} \hline \text { Manduted } \\ \hline \text { Darter } \\ \hline \end{array}$ | -- |  | 1 | 0.0 | -- |  | 5 | 0.0 | 1 | 0.0 | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 8 | 0.0 |
| Rosyface Shiner | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Swallowtail Shiner | -- |  | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 | -- |  | -- |  | -- |  | -- |  | 5 | 0.0 |
| Shield | -- |  | -- |  | -- |  | -- |  | -- |  | 7 | 0.0 | -- |  | -- |  | -- |  | -- |  | 7 | 0.0 |
| Greenside <br> Darter | 1 | 0.0 | -- |  | -- |  | 1 | 0.0 | -- |  | -- |  | -- |  | -- |  | 1 |  | -- |  | 3 | 0.0 |
| $\begin{array}{\|l} \hline \begin{array}{l} \text { Longnose } \\ \text { Dace } \end{array} \\ \hline \end{array}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{array}{\|l\|} \hline \text { Tadpole } \\ \text { Madtom } \\ \hline \end{array}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Trouts |  |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | . |
| Sunfishes | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Trout | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | - |
| $\begin{aligned} & \text { Rainbow } \\ & \text { Smelt } \end{aligned}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{array}{\|l\|} \hline \text { Blacknose } \\ \text { Dace } \end{array}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Mummichog | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | - |
| Lampreys | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Lake Herring | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{aligned} & \text { Striped } \\ & \text { Mullet } \end{aligned}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Sunfish Hyrids | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{array}{\|l} \hline \text { Palomino } \\ \text { (Rainbow } \\ \text { Trout) } \\ \hline \end{array}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | - |
| Redfin Pickerel | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| $\begin{array}{\|l\|} \hline \text { Carps and } \\ \text { Minnows } \\ \hline \end{array}$ | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| River Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Creek Chub | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  |  |
| Madtoms | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | - |
| Bigmouth Buffalo | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  | -- |  |  | . |
| Total | 458349 | 1081.0 | 309804 | 729.0 | 419103 | 1005.0 | 147388 | 401.6 | 37589 | 248.9 | 94767 | 321.2 | 163131 | 467.4 | 159389 | 553.4 | 733553 | 1525.1 | 225794 | 800.7 | 2748867 | 790.1 |

APPENDIX B-EAST FISH LIFT DATA

B-1. CONOWINGO EAST FISH LIFT CPUE AND NUMBER OF LIFTS 1991 - 2009


B-2. SPECIES RICHNESS EAST FISH LIFT, NUMBER OF TAXA ANNUALLY 1991-2009


## B-3. PROPORTIONAL ABUNDANCE EAST FISH LIFT 1991-2009



B-4. PROPORTIONAL ABUNDANCE EAST FISH LIFT GIZZARD SHAD REMOVED 1991 - 2009


B-5. PROPORTIONAL ABUNDANCE EAST FISH LIFT 1991-1999


B-6. PROPORTIONAL ABUNDANCE EAST FISH LIFT 1991 - 1999 GIZZARD SHAD REMOVED


## B-7. PROPORTIONAL ABUNDANCE EAST FISH LIFT 2000 - 2009



B-8. PROPORTIONAL ABUNDANCE EAST FISH LIFT 2000 - 2009 GIZZARD SHAD REMOVED


B-9. GIZZARD SHAD (DOROSOMA CEPEDIANUM) EAST FISH LIFT, FISH PER LIFT


B-10. AMERICAN SHAD (ALOSA SAPIDISSIMA) EAST FISH LIFT, FISH PER LIFT


B-10

B-11. BLUEBACK HERRING (ALOSA AESTIVALIS) EAST FISH LIFT, FISH PER LIFT


B-11

B-12. WHITE PERCH (MORONE AMERICANA) EAST FISH LIFT, FISH PER LIFT


B-12

B-13. COMMON CARP (CYPRINUS CARPIO) EAST FISH LIFT, FISH PER LIFT


B-13

B-14. CHANNEL CATFISH (ICTALURUS PUNCTATUS) EAST FISH LIFT, FISH PER LIFT


B-14

B-16. WALLEYE (SANDER VITREUS) EAST FISH LIFT, FISH PER LIFT


B-15

B-16. ALEWIFE (ALOSA PSEUDOHARENGUS) EAST FISH LIFT, FISH PER LIFT


B-16

B-17. STRIPED BASS (MORONE SAXATILIS) EAST FISH LIFT, FISH PER LIFT


B-18. SHORTHEAD REDHORSE (MOXOSTOMA MACROLEPIDOTUM) EAST FISH LIFT, FISH PER LIFT


B-18

B-19. SMALLMOUTH BASS (MICROPTERUS DOLOMIEU) EAST FISH LIFT, FISH PER LIFT


B-19

B-20. LARGEMOUTH BASS (MICROPTERUS SALMOIDES) EAST FISH LIFT, FISH PER LIFT


B-21. WHITE CRAPPIE (POMOXIS ANNULARIS) EAST FISH LIFT, FISH PER LIFT


B-22. AMERICAN EEL (ANGUILLA ROSTRATA) EAST FISH LIFT, FISH PER LIFT


B-23. EAST FISH LIFT CATCH 1991 -2009

| Year | 1991 | 1991 | 1991 | 1992 | 1992 | 1992 | 1993 | 1993 | 1993 | 1994 | 1994 | 1994 | 1995 | 1995 | 1995 | 1996 | 1996 | 1996 | 1997 | 1997 | 1997 | 1998 | 1998 | 1998 | 1999 | 1999 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 60 |  |  | 49 |  |  | 42 |  |  | 55 |  |  | 68 |  |  | 49 |  |  | 64 |  |  | 50 |  |  | 53 |  |  |
| Lifts | 1168 |  |  | 599 |  |  | 848 |  |  | 955 |  |  | 986 |  |  | 599 |  |  | 652 |  |  | 652 |  |  | 610 |  |  |
| Est. Oper. Time(HR) | 647.2 |  |  | 454.1 |  |  | 463.5 |  |  | 574.8 |  |  | 706.2 |  |  | 454.1 |  |  | 640 |  |  | 640 |  |  | 467 |  |  |
| Fishing Time (HR) | 561.9 |  |  | 731.5 |  |  | 421.2 |  |  | 517.7 |  |  | 653.3 |  |  | 420.6 |  |  | -- |  |  |  |  |  | -- |  |  |
| \#Species | 42 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | $\begin{gathered} \% \\ \text { \% comp } \end{gathered}$ | 45 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | $\begin{array}{\|c} \hline \% \\ \text { comp } \\ \hline \end{array}$ | 29 | $\begin{array}{\|c} \hline \text { Fish } \\ \text { per lift } \end{array}$ | $\begin{gathered} \text { comp } \\ \text { comp } \end{gathered}$ | 36 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \\ \hline \end{array}$ | $\begin{gathered} \% \\ \text { \% comp } \end{gathered}$ | 36 | $\begin{gathered} \text { Fish } \\ \text { Fer lift } \end{gathered}$ | $\begin{array}{\|c\|} \hline \% \\ \text { comp } \end{array}$ | 35 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \\ \hline \end{array}$ | $\begin{gathered} \% \\ \text { \% } \\ \hline \end{gathered}$ | 36 | $\begin{gathered} \hline \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \end{gathered}$ | $\begin{gathered} \% \\ \text { \% } \\ \hline \end{gathered}$ | 33 | $\begin{aligned} & \text { Fish } \\ & \text { Fer lift } \end{aligned}$ | $\begin{gathered} \% \\ \text { \% } \\ \hline \end{gathered}$ | ${ }^{31}$ | $\begin{gathered} \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \\ \hline \hline \end{gathered}$ | $\begin{array}{\|c\|} \hline \% \\ \hline \text { comp } \\ \hline \end{array}$ |
| Gizzard Shad | 575505 | 492.73 | 88.4\% | 2351351 | 3925.4 | 98.2\% | 504116 | 594.48 | 95.2\% | 1025418 | 1073.7 | 96.5\% | 173785 | 1762.3 | 96.7\% | 455317 | 760.13 | 92.5\% | 34432 | 528.12 | 47.9\% | 654575 | 1003.9 | 91.8\% | 95050 | 1558.2 | 80.3\% |
| American Shad | 13897 | 11.90 | 2.1\% | 15386 | 25.69 | 0.6\% | 8203 | 9.67 | 1.5\% | 26715 | 27.97 | 2.5\% | 46062 | 46.72 | 2.6\% | 26040 | 43.47 | 5.3\% | 90971 | 139.53 | 12.6\% | 39904 | 61.20 | 5.6\% | 69712 | 114.28 | 5.9\% |
| Blueback Herring | 13149 | 11.26 | 2.0\% | 7347 | 12.27 | 0.3\% | 4574 | 5.39 | 0.9\% | 248 | 0.26 |  | 4004 | 4.06 | 0.2\% | 261 | 0.44 | 0.1\% | $\begin{gathered} 24281 \\ 5 \end{gathered}$ | 372.42 | 33.8\% | 700 | 1.07 | 0.1\% | $\begin{gathered} 13062 \\ 5 \end{gathered}$ | 214.14 | 11.0\% |
| White Perch | 2610 | 2.23 | 0.4\% | 8725 | 14.57 | 0.4\% | 215 | 0.25 |  | 133 | 0.14 |  | 528 | 0.54 |  | 49 | 0.08 |  | 27312 | 41.89 | 3.8\% | 2731 | 4.19 | 0.4\% | 27133 | 44.48 | 2.3\% |
| Common Carp | 23833 | 20.40 | 3.7\% | 6072 | 10.14 | 0.3\% | 6649 | 7.84 | 1.3\% | 5042 | 5.28 | 0.5\% | 3262 | 3.31 | 0.2\% | 4139 | 6.91 | 0.8\% | 3256 | 4.99 | 0.5\% | 6205 | 9.52 | 0.9\% | 2430 | 3.98 | 0.2\% |
| Quillback | 3220 | 2.76 | 0.5\% | 483 | 0.81 |  | 540 | 0.64 | 0.1\% | 2507 | 2.63 | 0.2\% | 2910 | 2.95 | 0.2\% | 3773 | 6.30 | 0.8\% | 2488 | 3.82 | 0.3\% | 218 | 0.33 |  | 144 | 0.24 |  |
| Comely Shiner | 11847 | 10.14 | 1.8\% | 650 | 1.09 |  | 3563 | 4.20 | 0.7\% | 433 | 0.45 |  | 163 | 0.17 |  | 117 | 0.20 |  | 140 | 0.21 |  | 164 | 0.25 |  |  |  |  |
| Channel Catfish | 321 | 0.27 |  | 1124 | 1.88 |  | 534 | 0.63 | 0.1\% | 544 | 0.57 | 0.1\% | 90 | 0.09 |  | 1037 | 1.73 | 0.2\% | 1178 | 1.81 | 0.2\% | 4135 | 6.34 | 0.6\% | 266 | 0.44 |  |
| Walleye | 335 | 0.29 | 0.1\% | 150 | 0.25 |  | 71 | 0.08 |  | 255 | 0.27 |  | 271 | 0.27 |  | 351 | 0.59 | 0.1\% | 2334 | 3.58 | 0.3\% | 685 | 1.05 | 0.1\% | 421 | 0.69 |  |
| Striped Bass | 581 | 0.50 | 0.1\% | 216 | 0.36 |  | 327 | 0.39 | 0.1\% | 506 | 0.53 |  | 505 | 0.51 |  | 276 | 0.46 | 0.1\% | 1015 | 1.56 | 0.1\% | 1467 | 2.25 | 0.2\% | 1231 | 2.02 | 0.1\% |
| Alewife | 323 | 0.28 |  | 285 | 0.48 |  | -- |  |  | 5 | 0.01 |  | 170 | 0.17 |  | , | 0.01 |  | 63 | 0.10 |  | 6 | 0.01 |  | 14 | 0.02 |  |
| Smallmouth Bass | 671 | 0.57 | 0.1\% | 494 | 0.82 |  | 185 | 0.22 |  | 212 | 0.22 |  | 120 | 0.12 |  | 531 | 0.89 | 0.1\% | 783 | 1.20 | 0.1\% | 508 | 0.78 | 0.1\% | 797 | 1.31 | 0.1\% |
| Shorthead Redhorse | 424 | 0.36 | 0.1\% | 278 | 0.46 |  | 184 | 0.22 |  | 242 | 0.25 |  | 118 | 0.12 |  | 228 | 0.38 |  | 1475 | 2.26 | 0.2\% | 885 | 1.36 | 0.1\% | 245 | 0.40 |  |
| Spotfin Shiner | 2647 | 2.27 | 0.4\% | 35 | 0.06 |  | -- |  |  | -- |  |  |  |  |  |  |  |  | 17 | 0.03 |  |  |  |  |  |  |  |
| Bluegill | 149 | 0.13 |  | 399 | 0.67 |  | 58 | 0.07 |  | 45 | 0.05 |  | 46 | 0.05 |  | 37 | 0.06 |  | 334 | 0.51 |  | 354 | 0.54 |  | 159 | 0.26 |  |
| Striped Bass x White Bass | 827 | 0.71 | 0.1\% | 413 | 0.69 |  | ${ }^{64}$ | 0.08 |  | 53 | 0.06 |  | 8 | 0.01 |  | 4 | 0.01 |  | 1 | 0.00 |  | 4 | 0.01 |  | 5 | 0.01 |  |
| Sea Lamprey | 19 | 0.02 |  | 17 | 0.03 |  | 4 | 0.00 |  | 5 | 0.01 |  | 4 | 0.00 |  | 9 | 0.02 |  | 30 | 0.05 |  | 11 | 0.02 |  | 27 | 0.04 |  |
| Redbreast Sunfish | 115 | 0.10 |  | 110 | 0.18 |  | 34 | 0.04 |  | 24 | 0.03 |  | 185 | 0.19 |  | 17 | 0.03 |  | 195 | 0.30 |  | 46 | 0.07 |  | 108 | 0.18 |  |
| Yellow Perch | 45 | 0.04 |  | 36 | 0.06 |  | 46 | 0.05 |  | 7 | 0.01 |  | 22 | 0.02 |  | 12 | 0.02 |  | 93 | 0.14 |  | 51 | 0.08 |  | 108 | 0.18 |  |
| White Sucker | 51 | 0.04 |  | 96 | 0.16 |  | 82 | 0.10 |  | 42 | 0.04 |  | 43 | 0.04 |  | 73 | 0.12 |  | 7 | 0.01 |  | 81 | 0.12 |  | 58 | 0.10 |  |
| Carps and Minnows | 100 | 0.09 |  | 554 | 0.92 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| American Eel | 103 | 0.09 |  | 119 | 0.20 |  | 49 | 0.06 |  | 54 | 0.06 |  | 162 | 0.16 |  | 39 | 0.07 |  | 13 | 0.02 |  | 5 | 0.01 |  | 3 | 0.00 |  |
| Largemouth Bass | 17 | 0.01 |  | 33 | 0.06 |  | 12 | 0.01 |  | 27 | 0.03 |  | 13 | 0.01 |  | 9 | 0.02 |  | 147 | 0.23 |  | 62 | 0.10 |  | 30 | 0.05 |  |
| Rock Bass | ? | 0.00 |  | 16 | 0.03 |  | 10 | 0.01 |  | 1 | 0.00 |  | 2 | 0.00 |  | 3 | 0.01 |  | 204 | 0.31 |  | 74 | 0.11 |  | 31 | 0.05 |  |
| Brown Bullhead | 3 | 0.00 |  | 3 | 0.01 |  | 1 | 0.00 |  | -- |  |  | 2 | 0.00 |  | 3 | 0.01 |  | 5 | 0.01 |  | 15 | 0.02 |  | 2 | 0.00 |  |
| Brown Trout | 44 | 0.04 |  | 76 | 0.13 |  | 53 | 0.06 |  | 42 | 0.04 |  | 13 | 0.01 |  | 22 | 0.04 |  | 19 | 0.03 |  | 61 | 0.09 |  | 9 | 0.01 |  |
| Spottail Shiner | 21 | 0.02 |  | 1 | 0.00 |  |  |  |  | -- |  |  | 37 | 0.04 |  |  | 0.00 |  | , | 0.00 |  |  |  |  |  |  |  |
| Rainbow Trout | 7 | 0.01 |  | 10 | 0.02 |  | 5 | 0.01 |  | 5 | 0.01 |  | 10 | 0.01 |  | 9 | 0.02 |  | 6 | 0.01 |  | 11 | 0.02 |  | 12 | 0.02 |  |
| Pumpkinseed | 16 | 0.01 |  | 13 | 0.02 |  | 2 | 0.00 |  | 3 | 0.00 |  |  |  |  | 1 | 0.00 |  | 36 | 0.06 |  | 4 | 0.01 |  | 1 | 0.00 |  |
| Green Sunfish | 10 | 0.01 |  | 12 | 0.02 |  | 1 | 0.00 |  | 8 | 0.01 |  | - | 0.00 |  | 4 | 0.01 |  | 6 | 0.01 |  | 2 | 0.00 |  | 13 | 0.02 |  |
| Golden Shiner | -- |  |  | -- |  |  |  |  |  | 45 | 0.05 |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Muskellunge | 3 | 0.00 |  | 10 | 0.02 |  | 9 | 0.01 |  | 5 | 0.01 |  |  |  |  | 2 | 0.00 |  | 5 | 0.01 |  | 2 | 0.00 |  |  |  |  |
| White Crappie | 7 | 0.01 |  | 4 | 0.01 |  | -- |  |  | -- |  |  | 1 | 0.00 |  | 1 | 0.00 |  |  | 0.00 |  | 12 | 0.02 |  | 3 | 0.00 |  |
| Hickory Shad | -- |  |  | 20 | 0.03 |  |  |  |  | 1 | 0.00 |  | 1 | 0.00 |  |  |  |  | -- |  |  |  |  |  |  |  |  |
| Yellow Bullhead | - |  |  | 1 | 0.00 |  | 2 | 0.00 |  | -- |  |  |  |  |  |  |  |  |  |  |  | 5 | 0.01 |  |  |  |  |
| Herrings | -- |  |  | 29 | 0.05 |  |  |  |  |  |  |  |  |  |  |  |  |  | -- |  |  |  |  |  |  |  |  |
| Tiger Muskie | 3 | 0.00 |  | 2 | 0.00 |  | 1 | 0.00 |  | -- |  |  | 5 | 0.01 |  | 4 | 0.01 |  |  |  |  |  |  |  |  |  |  |
| Black Crappie | 3 | 0.00 |  | 1 | 0.00 |  | -- |  |  | 1 | 0.00 |  | -- |  |  |  |  |  | 2 | 0.00 |  | 5 | 0.01 |  | 6 | 0.01 |  |
| Blacknose Dace | -- |  |  | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| White Catish | -- |  |  | -- |  |  | -- |  |  | -- |  |  |  |  |  | 4 | 0.01 |  |  |  |  | 3 | 0.00 |  | 2 | 0.00 |  |
| Tessellated Darter | 10 | 0.01 |  | 2 | 0.00 |  |  |  |  | -- |  |  | , | 0.01 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Atlantic Needlefish | 2 | 0.00 |  | -- |  |  |  |  |  | 1 | 0.00 |  | 1 | 0.00 |  |  |  |  |  |  |  | 2 |  |  | 4 | 0.01 |  |
| Banded Darter | 9 | 0.01 |  | 1 | 0.00 |  |  |  |  | -- |  |  | 2 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brook Trout x Lake Trout | 1 | 0.00 |  | 1 | 0.00 |  | -- |  |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brook Trout | 1 | 0.00 |  | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  |  |  |  | 2 | 0.00 |  |  |  |  |  |  |  | 2 | 0.00 |  |
| Margined Madtom | -- |  |  | -- |  |  | -- |  |  | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Notropis Sp. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 | 0.01 |  |  |  |  |  |  |  |
| Logperch | $\stackrel{-}{5}$ |  |  | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  | 3 | 0.00 |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |
| Hybrids | 5 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |



| Year | 2000 | 2000 | 2000 | 2001 | 2001 | 2001 | 2002 | 2002 | 2002 | 2003 | 2003 | 2003 | 2004 | 2004 | 2004 | 2005 | 2005 | 2005 | 2006 | 2006 | 2006 | 2007 | 2007 | 2007 | 2008 | 2008 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 45 |  |  | 43 |  |  | 51 |  |  | 44 |  |  | 44 |  |  | 52 |  |  | 61 |  |  | 39 |  |  | 51 |  |  |
| Liff | 570 |  |  | 559 |  |  | 560 |  |  | 645 |  |  | 590 |  |  | 541 |  |  | 619 |  |  | 479 |  |  | 483 |  |  |
| Est. Oper. Time(HR) | 367.76 |  |  | 359.8 |  |  | 440.7 |  |  | 416.6 |  |  | 390.3 |  |  | 434.32 |  |  | 429.8 |  |  | 335.25 |  |  | 409 |  |  |
| Fishing Time (HR) |  |  |  | - |  |  | -- |  |  | -- |  |  | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Species | 31 | Fish | \% | 30 | Fish | \% | 31 | Fish | \% | 25 | Fish | \% | 31 | Fish | \% | 28 | Fish | \% | 32 | Fish | \% | 31 | Fish | \% | 28 | Fish | \% |
|  |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |  | per lift | comp |
| Gizzard Shad | 317753 | 557.46 | 64.3\% | 429461 | 768.27 | 46.6\% | 513794 | 917.49 | 78.2\% | 459634 | 712.61 | 78.0\% | 602677 | 1021.49 | 84.2\% | 305378 | 564.47 | 80.8\% | 65990 | 1059.76 | 91.8\% | 508627 | 1061.85 | 94.3\% | 919975 | 1904.71 | 97.5\% |
| American Shad | 153546 | 269.38 | 31.1\% | 193574 | 346.29 | 21.0\% | 108001 | 192.86 | 16.4\% | 125135 | 194.01 | 21.2\% | 109360 | 185.36 | 15.3\% | 68926 | 127.40 | 18.2\% | 56899 | 91.92 | 8.0\% | 25464 | 53.16 | 4.7\% | 19914 | 41.23 | 2.1\% |
| Blueback Herring | 14963 | 26.25 | 3.0\% | 284921 | 509.70 | 30.9\% | 2037 | 3.64 | 0.3\% | 530 | 0.82 | 0.1\% | 101 | 0.17 |  | 4 | 0.01 |  |  |  |  | 460 | 0.96 | 0.1\% | 1 | 0.00 |  |
| White Perch | 4387 | 7.70 | 0.9\% | 2659 | 4.76 | 0.3\% | 29404 | 52.51 | 4.5\% | 1572 | 2.44 | 0.3\% | 512 | 0.87 | 0.1\% | 15 | 0.03 |  | 277 | 0.45 |  | 1434 | 2.99 | 0.3\% | 388 | 0.80 |  |
| Common Carp | 388 | 0.68 | 0.1\% | 1267 | 2.27 | 0.1\% | 172 | 0.31 |  | 561 | 0.87 | 0.1\% | 257 | 0.44 |  | 540 | 1.00 | 0.1\% | 108 | 0.17 |  | 107 | 0.22 |  | 199 | 0.41 |  |
| Quillback | 408 | 0.72 | 0.1\% | 241 | 0.43 |  | 400 | 0.71 | 0.1\% | 548 | 0.85 | 0.1\% | 308 | 0.52 |  | 2145 | 3.96 | 0.6\% | 407 | 0.66 | 0.1\% | 1236 | 2.58 | 0.2\% | 400 | 0.83 |  |
| Comely Shiner |  |  |  | 4 | 0.01 |  |  |  |  |  |  |  | 291 | 0.49 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Channel Catfish | 677 | 1.19 | 0.1\% | 29 | 0.05 |  | 199 | 0.36 |  | 57 | 0.09 |  | 928 | 1.57 | 0.1\% | 83 | 0.15 |  | 75 | 0.12 |  | 108 | 0.23 |  | 496 | 1.03 | 0.1\% |
| Walleye | 177 | 0.31 |  | 91 | 0.16 |  | 88 | 0.16 |  | 59 | 0.09 |  | 156 | 0.26 |  | 47 | 0.09 |  | 641 | 1.04 | 0.1\% | 695 | 1.45 | 0.1\% | 2088 | 4.32 | 0.2\% |
| Striped Bass | 802 | 1.41 | 0.2\% | 543 | 0.97 | 0.1\% | 913 | 1.63 | 0.1\% | 272 | 0.42 |  | 391 | 0.66 | 0.1\% | 89 | 0.16 |  | 73 | 0.12 |  | 127 | 0.27 |  | 20 | 0.04 |  |
| Alewife | 2 | 0.00 |  | 7458 | 13.34 | 0.8\% | 74 | 0.13 |  | 21 | 0.03 |  | 89 | 0.15 |  |  |  |  |  |  |  | 429 | 0.90 | 0.1\% | 4 | 0.01 |  |
| Smallmouth Bass | 427 | 0.75 | 0.1\% | 404 | 0.72 |  | 597 | 1.07 | 0.1\% | 247 | 0.38 |  | 172 | 0.29 |  | 256 | 0.47 | 0.1\% | 165 | 0.27 |  | 123 | 0.26 |  | 96 | 0.20 |  |
| Shorthead Redhorse | 91 | 0.16 |  | 382 | 0.68 |  | 292 | 0.52 |  | 304 | 0.47 | 0.1\% | 113 | 0.19 |  | 131 | 0.24 |  | 10 | 0.02 |  | 173 | 0.36 |  | 66 | 0.14 |  |
| Spotfin Shiner |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Bluegill | 96 | 0.17 |  | 55 | 0.10 |  | 130 | 0.23 |  | 37 | 0.06 |  | 19 | 0.03 |  | 10 | 0.02 |  | 25 | 0.04 |  | 27 | 0.06 |  | 65 | 0.13 |  |
| Striped Bass x White Bass | 2 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.00 |  | 4 | 0.01 |  | 6 | 0.01 |  | 1 | 0.00 |  |
| Sea Lamprey | 23 | 0.04 |  | 268 | 0.48 |  | 316 | 0.56 |  | 68 | 0.11 |  | 58 | 0.10 |  | 35 | 0.06 |  | 128 | 0.21 |  | 22 | 0.05 |  | 11 | 0.02 |  |
| Redbreast Sunfish | 20 | 0.04 |  | 46 | 0.08 |  | 51 | 0.09 |  | 4 | 0.01 |  | 8 | 0.01 |  | 19 | 0.04 |  | 5 | 0.01 |  |  |  |  | 3 | 0.01 |  |
| Yellow Perch | 31 | 0.05 |  | 33 | 0.06 |  | 258 | 0.46 |  | 37 | 0.06 |  | 5 | 0.01 |  | 24 | 0.04 |  | 26 | 0.04 |  | 5 | 0.01 |  | 4 | 0.01 |  |
| White Sucker | 29 | 0.05 |  | 66 | 0.12 |  | 16 | 0.03 |  | 10 | 0.02 |  | 11 | 0.02 |  | 8 | 0.01 |  | 3 | 0.00 |  | 13 | 0.03 |  |  |  |  |
| Carps and Minnows |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| American Eel |  |  |  | 3 | 0.01 |  |  |  |  |  |  |  |  |  |  | 5 | 0.01 |  | 11 | 0.02 |  |  |  |  |  |  |  |
| Largemouth Bass | 33 | 0.06 |  | 31 | 0.06 |  | 38 | 0.07 |  | 21 | 0.03 |  | 15 | 0.03 |  | 14 | 0.03 |  | 14 | 0.02 |  | 11 | 0.02 |  | 15 | 0.03 |  |
| Rock Bass | 25 | 0.04 |  | 33 | 0.06 |  | 41 | 0.07 |  | 18 | 0.03 |  | 7 | 0.01 |  | 5 | 0.01 |  | 14 | 0.02 |  | 17 | 0.04 |  | 14 | 0.03 |  |
| Brown Bullhead | 32 | 0.06 |  | 5 | 0.01 |  | 6 | 0.01 |  | 10 | 0.02 |  | 161 | 0.27 |  | 9 | 0.02 |  | 5 | 0.01 |  | 80 | 0.17 |  | 27 | 0.06 |  |
| Brown Trout | 8 | 0.01 |  | 8 | 0.01 |  | 12 | 0.02 |  |  |  |  | 3 | 0.01 |  | 4 | 0.01 |  | 6 | 0.01 |  | 17 | 0.04 |  | 6 | 0.01 |  |
| Spotail Shiner |  |  |  | 318 | 0.57 |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  | 2 | 0.00 |  |  |  |  | 2 | 0.00 |  |


| Year | 2000 | 2000 | 2000 | 2001 | 2001 | 2001 | 2002 | 2002 | 2002 | 2003 | 2003 | 2003 | 2004 | 2004 | 2004 | 2005 | 2005 | 2005 | 2006 | 2006 | 2006 | 2007 | 2007 | 2007 | 2008 | 2008 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 45 |  |  | 43 |  |  | 51 |  |  | 44 |  |  | 44 |  |  | 52 |  |  | 61 |  |  | 39 |  |  | 51 |  |  |
| Lifts | 570 |  |  | 559 |  |  | 560 |  |  | 645 |  |  | 590 |  |  | 541 |  |  | 619 |  |  | 479 |  |  | 483 |  |  |
| Est. Oper. Time(HR) | 367.76 |  |  | 359.8 |  |  | 440.7 |  |  | 416.6 |  |  | 390.3 |  |  | 434.32 |  |  | 429.8 |  |  | 335.25 |  |  | 409 |  |  |
| Fishing Time (HR) |  |  |  | -- |  |  | -- |  |  | -- |  |  | -- |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Species | 31 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \\ \hline \end{array}$ | $\begin{gathered} \% \\ \text { \% } \\ \text { comp } \end{gathered}$ | 30 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | $\begin{gathered} \hline \% \\ \text { comp } \end{gathered}$ | 31 | $\begin{array}{\|c} \hline \begin{array}{c} \text { Fish } \\ \text { per lift } \end{array} \end{array}$ | $\begin{gathered} \% \\ \text { \% } \\ \hline \end{gathered}$ | 25 | $\begin{array}{\|c} \hline \text { Fish } \\ \text { per lift } \end{array}$ |  | 31 | $\begin{array}{\|c} \hline \text { Fish } \\ \text { per lift } \end{array}$ | $\begin{gathered} \% \\ \text { comp } \\ \hline \end{gathered}$ | 28 | $\begin{array}{\|c} \hline \text { Fish } \\ \text { per lift } \end{array}$ | $\begin{gathered} \% \\ \text { comp } \end{gathered}$ | 32 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | $\begin{gathered} \% \\ \text { \% } \\ \hline \end{gathered}$ | 31 | $\begin{array}{\|c} \hline \text { Fish } \\ \text { per lift } \end{array}$ | $\begin{gathered} \% \\ \text { comp } \\ \hline \end{gathered}$ | 28 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | $\begin{gathered} \% \\ \text { comp } \end{gathered}$ |
| Rainbow Trout | 2 | 0.00 |  | 6 | 0.01 |  | 20 | 0.04 |  | 3 | 0.00 |  | 8 | 0.01 |  | 4 | 0.01 |  | 5 | 0.01 |  | 10 | 0.02 |  | 32 | 0.07 |  |
| Pumpkinseed | 3 | 0.01 |  |  |  |  | 14 | 0.03 |  |  |  |  | 1 | 0.00 |  | 1 | 0.00 |  | 4 | 0.01 |  | 2 | 0.00 |  | 4 | 0.01 |  |
| Green Sunfish |  | 0.00 |  | 6 | 0.01 |  | 5 | 0.01 |  | 2 | 0.00 |  | 1 | 0.00 |  | 1 | 0.00 |  |  |  |  | I | 0.00 |  |  |  |  |
| Golden Shiner |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 0.00 |  |  |  |  |  |  |  |
| Muskellunge | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.00 |  | 2 | 0.00 |  | 5 | 0.01 |  |
| White Crappie | 2 | 0.00 |  | 1 | 0.00 |  |  |  |  |  |  |  | 3 | 0.01 |  |  |  |  |  | 0.00 |  | 1 | 0.00 |  |  |  |  |
| Hickory Shad |  |  |  |  |  |  | 6 | 0.01 |  |  |  |  | 0 |  |  |  |  |  | 4 | 0.01 |  |  |  |  | 0 |  |  |
| Yellow Bullhead | 13 | 0.02 |  | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  | 2 | 0.00 |  |  |  |  | 1 | 0.00 |  | 2 | 0.00 |  |  |  |  |
| Herrings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Tiger Muskie |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 6 | 0.01 |  | 5 | 0.01 |  | 1 | 0.00 |  |  |  |  |
| Black Crappie |  | 0.00 |  | 1 | 0.00 |  |  |  |  |  |  |  | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  | 1 | 0.00 |  |  |  |  |
| Blacknose Dace |  |  |  |  |  |  |  |  |  | 25 | 0.04 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| White Catfish | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  | 1 | 0.00 |  | 5 | 0.01 |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |
| Tessellated Darter |  |  |  |  |  |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Atlantic Needlefish |  |  |  |  |  |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  | 6 | 0.01 |  |  |  |  | 1 |  |  |
| Banded Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brook Trout x Lake Trout |  |  |  | 1 | 0.00 |  | 4 | 0.01 |  |  |  |  | 1 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brook Trout |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.00 |  |
| Margined Madtom | 7 | ${ }_{0} 0.01$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Notropis Sp. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Logperch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Hybrids |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Northern Hog Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Creek Chubsucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Northern Pike |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Flathead Catfish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.00 |  |  |  |  | 1 | 0.00 |  |  |  |  |
| Creek Chub |  |  |  |  |  |  | 2 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Salmo Sp. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shield Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Longnose gar |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.00 |  |  |  |  |
| Chain Pickerel |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Esox Sp. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Longnose Dace <br> Sunfish Hybrids |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sunfish Hybrids |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 493953 | 86.58 |  | 921916 | 1,649.2 |  | 656894 | 1,173.0 |  | 589177 | 913.5 |  | 715664 | 1213.0 |  | 377762 | 698.3 |  | 714918 | 1155.0 |  | 539203 | 1125.7 |  | 943838 | 1954.1 |  |


| Year | 2009 | 2009 | 2009 | $\begin{aligned} & \text { Total } \\ & 1991- \\ & 2009 \end{aligned}$ | Total fish per lift 1991- 2009 | $\begin{array}{r} \text { \% Comp } \\ 1991-2009 \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 57 |  |  | 977 |  |  |
| Lifts | 618 |  |  | 12733 |  |  |
| Est. Oper. Time(HR) | 495.6 |  |  | 9126.03 |  |  |
| Fishing Time (HR) |  |  |  |  |  |  |
| \#Species | 30 | Fish | $\begin{gathered} \% \\ \text { comp } \end{gathered}$ | 63 | Fish per lift | \% comp |
| Gizzard Shad | 876412 | 1418.14 | 95.7\% | 14188500 | 1114.31 | 86.5\% |
| American Shad | 29272 | 47.37 | 3.2\% | 1226981 | 96.36 | 7.5\% |
| Blueback Herring | 71 | 0.11 |  | 706811 | 55.51 | 4.3\% |
| White Perch | 839 | 1.36 | 0.1\% | 110923 | 8.71 | 0.7\% |
| Common Carp | 886 | 1.43 | 0.1\% | 65373 | 5.13 | 0.4\% |
| Quillback | 899 | 1.45 | 0.1\% | 23275 | 1.83 | 0.1\% |
| Comely Shiner |  |  |  | 17372 | 1.36 | 0.1\% |
| Channel Catfish | 4201 | 6.80 | 0.5\% | 16082 | 1.26 | 0.1\% |
| Walleye | 1832 | 2.96 | 0.2\% | 10747 | 0.84 | 0.1\% |
| Striped Bass | 66 | 0.11 |  | 9420 | 0.74 | 0.1\% |
| Alewife | 160 | 0.26 |  | 9106 | 0.72 | 0.1\% |
| Smallmouth Bass | 113 | 0.18 |  | 6901 | 0.54 | 0.04\% |
| Shorthead Redhorse | 148 | 0.24 |  | 5789 | 0.45 | 0.04\% |
| Spotfin Shiner |  |  |  | 2699 | 0.21 | 0.02\% |
| Bluegill | 67 | 0.11 |  | 2112 | 0.17 | 0.01\% |
| Striped Bass x White | 1 | 0.00 |  | 1394 | 0.11 |  |
| Sass Lamprey | 190 | 0.31 |  | 1245 | 0.10 | 0.01\% |
| Redbreast Sunfish | 1 | 0.00 |  | 991 | 0.08 | 0.01\% |
| Yellow Perch | 23 | 0.04 |  | 866 | 0.07 | 0.01\% |
| White Sucker | 16 | 0.03 |  | 705 | 0.06 | 0.004\% |
| Carps and Minnows |  |  |  | 654 | 0.05 | 0.004\% |
| American Eel |  |  |  | 566 | 0.04 | 0.003\% |
| Largemouth Bass | 11 | 0.02 |  | 553 | 0.04 | 0.003\% |
| Rock Bass | 19 | 0.03 |  | 537 | 0.04 | 0.003\% |
| Brown Bullhead | 153 | 0.25 |  | 522 | 0.04 | 0.003\% |
| Brown Trout | 16 | 0.03 |  | 419 | 0.03 | 0.003\% |
| Spottail Shiner |  |  |  | 387 | 0.03 | 0.002\% |
| Rainbow Trout | 5 | 0.01 |  | 170 | 0.01 | 0.001\% |
| Pumpkinseed | 4 | 0.01 |  | 109 | 0.01 | 0.001\% |
| Green Sunfish | 2 | 0.00 |  | 78 | 0.01 | 0.0005\% |
| Golden Shiner |  |  |  | 48 | 0.00 | 0.0003\% |
| Muskellunge | 2 | 0.00 |  | 47 | 0.00 | 0.0003\% |
| White Crappie |  |  |  | 37 | 0.00 | 0.0002\% |
| Hickory Shad |  |  |  | 32 | 0.00 | 0.0002\% |
| Yellow Bullhead | 1 | 0.00 |  | 29 | 0.00 | 0.0002\% |
| Herrings |  |  |  | 29 | 0.00 | 0.0002\% |
| Tiger Muskie |  |  |  | 27 | 0.00 | 0.0002\% |
| Black Crappie | 2 | 0.00 |  | 26 | 0.00 | 0.0002\% |
| Blacknose Dace |  |  |  | 25 | 0.00 | 0.0002\% |
| White Catish | 4 | 0.01 |  | 22 | 0.00 | 0.0001\% |
| Tessellated Darter |  |  |  | 21 | 0.00 | 0.0001\% |
| Atlantic Needlefish |  |  |  | 18 | 0.00 | 0.0001\% |
| Banded Darter |  |  |  | 12 | 0.00 | 0.0001\% |
| Brook Trout x Lake <br> Trout |  |  |  | 9 | 0.00 | 0.0001\% |
| Brook Trout |  |  |  | 8 | 0.00 | 0.00005\% |
| Margined Madtom |  |  |  | 7 | 0.00 | 0.00004\% |
| Notropis Sp. |  |  |  | 7 | 0.00 | 0.00004\% |
| Logperch |  |  |  | 6 | 0.00 | 0.00004\% |
| Hybrids |  |  |  | 5 | 0.00 | 0.00003\% |
| Northern Hog Sucker |  |  |  | 4 | 0.00 | 0.00002\% |


| Year | 2009 | 2009 | 2009 | $\begin{aligned} & \hline \text { Total } \\ & 1991- \\ & 2009 \end{aligned}$ | Total <br> fish <br> per lift <br> $1991-$ <br> 2009 | \% Comp $1991-$ 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Days | 57 |  |  | 977 |  |  |
| Lifts | 618 |  |  | 12733 |  |  |
| Est. Oper. Time(HR) | 495.6 |  |  | 9126.03 |  |  |
| Fishing Time (HR) |  |  |  |  |  |  |
| \#Species | 30 | Fish per lift | $\begin{gathered} \% \\ \text { \% } \\ \text { comp } \end{gathered}$ | 63 | $\begin{gathered} \text { Fish } \\ \text { per lift } \end{gathered}$ | \% comp |
| Creek Chubsucker |  |  |  | 3 | 0.00 | 0.00002\% |
| Northern Pike | 1 | 0.00 |  | 2 | 0.00 | 0.00002\% |
| Flathead Catish |  |  |  | 2 | 0.00 | 0.00001\% |
| Creek Chub |  |  |  | 2 | 0.00 | 0.00001\% |
| Salmo Sp. |  |  |  | 2 | 0.00 | 0.00001\% |
| Shield Darter |  |  |  | 2 | 0.00 | 0.00001\% |
| Longnose gar |  |  |  | 1 | 0.00 | 0.00001\% |
| Chain Pickerel |  |  |  | 1 | 0.00 | 0.00001\% |
| Esox Sp. |  |  |  | 1 | 0.00 | 0.00001\% |
| Longnose Dace |  |  |  | 1 | 0.00 | 0.00001\% |
| Sunfish Hybrids |  |  |  | 1 | 0.00 | 0.00001\% |
| Bluntrose Minnow |  |  |  | 1 | 0.00 | 0.00001\% |
| Total | 915417 | 1481.3 | 10 | 16411728 | 1288.9 |  |

## APPENDIX C-ICHTHYOPLANKTON

C-1. WEEKLY SUMMARY OF ICHTHYOP1ANKTON (EGGS AND LARVAE) TAKEN BY 0.5 M PLANKTON NET IN THE SUSQUEHANNA RIVER BELOW CONOWINGO HYDROELECTRIC STATION, APRIL THROUGH JUNE 1983.

|  | A | A | A | M | M | M | M | M | J | J | J | J |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4/3/1983 | 4/10/1983 | 4/17/1983 | 5/1/1983 | 5/8/1983 | 5/15/1983 | 5/22/1983 | 5/29/1983 | 6/5/1983 | 6/12/1983 | 6/19/1983 | 6/26/1983 | Total |
| White perch | 8 | 3 | 359 | 2,372 | 2,583 | 30,752 | 14,751 | 47,660 | 10,992 | 2,744 | 22 | 3 | 112,249 |
| Alosa spp. |  |  | 15 | 2,168 | 1,746 | 6,921 | 9,187 | 6,143 | 601 | 45 |  | 1 | 26,827 |
| Gizzard Shad |  |  | 1 |  |  | 208 | 308 | 311 | 1,539 | 449 | 627 | 21 | 3,464 |
| Unid. Eggs |  |  | 4 | 82 | 55 | 125 | 73 | 48 | 34 | 9 | 1 |  | 431 |
| Carp |  |  |  |  |  | 1 | 4 | 12 | 277 | 10 | 2 | 1 | 307 |
| Quillback |  |  |  |  | 1 | 21 | 130 | 78 | 14 | 20 |  | 1 | 265 |
| Cyprinidae |  |  |  |  | 19 | 50 | 115 | 29 | 5 | 2 | 2 |  | 222 |
| American Shad |  |  |  | 1 |  | 1 |  | 135 |  | 1 |  |  | 138 |
| Percidae |  |  |  | 2 |  | 7 | 11 | 37 | 3 | 3 |  |  | 63 |
| Catostomidae |  |  |  |  |  | 2 | 14 | 6 | 10 | 4 |  |  | 36 |
| White crappie |  |  |  |  |  |  | 1 | 2 | 4 | 17 | 5 | 1 | 30 |
| Yellow perch |  |  |  | 1 |  | 1 | 15 | 3 |  | 1 |  |  | 21 |
| Creek chubsucker |  |  |  | 1 |  |  | 1 |  |  | 6 |  |  | 8 |
| Bluegill |  |  |  |  |  | 1 | 1 |  |  | 2 | 3 | 1 | 8 |
| Unid. Larvae |  |  |  |  |  | 5 |  |  |  |  |  |  | 5 |
| Northern hog sucker |  |  |  |  |  |  | 2 |  |  |  |  |  | 2 |
| Shorthead redhorse |  |  |  |  |  |  | 2 |  |  |  |  |  | 2 |
| Pumpkinseed |  |  |  |  |  |  |  |  |  | 2 |  |  | 2 |
| Centrarchidae |  |  |  |  |  | 1 |  |  |  |  |  |  | 1 |
| Largemouth bass |  |  |  |  |  |  |  |  |  |  | 1 |  | 1 |
| Tessellated darter |  |  |  | 1 |  |  |  |  |  |  |  |  | 1 |
| Walleye |  |  |  |  | 1 |  |  |  |  |  |  |  | 1 |
| TOTAL | 8 | 3 | 379 | 4,628 | 4,405 | 38,096 | 24,615 | 54,464 | 13,479 | 3,315 | 663 | 29 | 144,084 |
| No. of species | 1 | 1 | 3 | 7 | 5 | 12 | 14 | 11 | 9 | 14 | 7 | 7 | 20 |
| No. of samples | 1 | 1 | 1 | 2 | 1 | 5 | 6 | 7 | 7 | 6 | 1 | 1 | 39 |
| No./Sample Day | 8.0 | 3.0 | 379.0 | 2314.0 | 4405.0 | 7619.2 | 4102.5 | 7780.6 | 1925.6 | 552.5 | 663.0 | 29.0 |  |

C-2. WEEKLY SUMMARY OF ICHTHYOP1ANKTON (EGGS AND LARVAE) TAKEN BY 0.5 M PLANKTON NET IN THE SUSQUEHANNA RIVER BELOW CONOWINGO HYDROELECTRIC STATION, APRIL THROUGH JUNE 1984.

| Common Name | $\begin{array}{\|r\|} \hline \mathbf{A} \\ 4 / 8 / 1984 \end{array}$ | $\begin{array}{r} A \\ 4 / 15 / 1984 \end{array}$ | $\begin{array}{\|r\|} \hline \mathrm{A} \\ \text { 4/22/1984 } \end{array}$ | \|4/29/1984 | $\begin{gathered} \hline \text { M } \\ 5 / 6 / 1984 \end{gathered}$ | $\begin{array}{\|c\|} \hline \text { M } \\ 5 / 13 / 1984 \end{array}$ | $\begin{array}{\|c\|} \hline \text { M } \\ 5 / 20 / 1984 \end{array}$ | $\begin{array}{\|c\|} \hline \text { M } \\ 5 / 27 / 1984 \end{array}$ | $\begin{gathered} \hline \mathrm{J} \\ 6 / 3 / 1984 \end{gathered}$ | $\begin{gathered} \hline \mathbf{J} \\ 6 / 10 / 1984 \end{gathered}$ | $\begin{gathered} \hline \mathbf{J} \\ 6 / 17 / 1984 \end{gathered}$ | $\begin{array}{\|c\|} \hline J \\ \text { 6/24/1984 } \end{array}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| White perch | 457 | 1,866 | 2,436 | 4,797 | 6,260 | 13,726 | 9,215 | 2,726 | 11,639 | 7,159 | 883 | 425 | 61,589 |
| Alosa spp. |  | 186 | 33 | 2,950 | 1,026 | 11,794 | 5,724 | 1,911 | 1,821 | 1,193 | 30 | 9 | 26,677 |
| Gizzard shad |  | 1 | 1 | 5 | 14 | 12 | 28 | 45 | 16 | 665 | 819 | 280 | 1,886 |
| Yellow perch |  | 215 |  |  | 1 | 3 |  | 3 | 5 | 1 |  | 1 | 229 |
| Suckers |  |  |  |  | 8 |  | 25 | 16 | 63 | 37 | 23 | 9 | 181 |
| American shad |  |  |  |  | 3 |  |  | 2 | 170 |  | 4 |  | 179 |
| M innows |  |  |  | 1 | 1 | 20 | 27 | 29 | 17 | 32 | 37 | 9 | 173 |
| Carp |  |  |  |  |  |  |  | 2 | 7 | 9 | 99 | 1 | 118 |
| Unidentified (eggs) |  | 2 | 1 | 3 | 5 | 11 | 21 | 22 | 14 | 19 | 16 |  | 114 |
| Hickory shad |  |  | 3 | 6 | 11 | 1 | 1 |  | 3 |  |  |  | 25 |
| Quillback |  |  |  |  |  |  |  | 1 | 10 | 2 | 6 | 2 | 21 |
| Perches |  |  |  |  | 1 | 5 | 2 | 1 | 8 | 2 |  |  | 19 |
| Bluegill |  |  |  |  |  |  |  |  | 1 | 1 | 6 |  | 8 |
| Tessellated darter |  |  |  |  |  | 1 |  |  | 4 | 2 | 1 |  | 8 |
| Smallmouth bass |  |  |  |  |  |  |  |  |  |  | 6 |  | 6 |
| Creek chubsucker |  |  |  |  |  |  |  |  | 1 |  |  | 3 | 4 |
| Sunfish family |  |  |  |  |  |  |  |  | 2 | 2 |  |  | 4 |
| White crappie |  |  |  |  |  |  |  |  |  |  | 4 |  | 4 |
| Spotfin shiner |  |  |  |  |  |  |  |  | 1 |  | 1 |  | 2 |
| Spottail shiner |  |  |  |  | 1 | 1 |  |  |  |  |  |  | 2 |
| White sucker |  |  |  |  |  |  | 2 |  |  |  |  |  | 2 |
| Largemouth bass |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 |
| Pumpkinseed |  |  |  |  | 1 |  |  |  |  |  |  |  | 1 |
| Shorthead redhorse |  |  |  |  |  |  |  |  | 1 |  |  |  | 1 |
| Walleye |  |  |  |  | 1 |  |  |  |  |  |  |  | 1 |
| TOTAL | 457 | 2,270 | 2,474 | 7,762 | 7,333 | 25,574 | 15,045 | 4,758 | 13,783 | 9,124 | 1,935 | 740 | 91,255 |

C-3. NUMBER AND DENSITY (N/M3) OF AMERICAN SHAD EGGS TAKEN BY ANCHORED 0.5M PLANKTON NETS, FISHED NEAR THE BOTTOM, OFF THE NORTHEAST SHORE OF SPENCER LSLAND (X-1075; Y-6825), 14 MAY THROUGH 20 JUNE 1983.

| Date | No. Samples | Time Span | Total Sampling Time (min) | $\begin{gathered} \text { Approximate Flow X } \\ 10^{3} \text { Range (cfs)* } \\ \hline \hline \end{gathered}$ | $\begin{gathered} \text { Water } \\ \text { Temperature (F) } \\ \hline \hline \end{gathered}$ | Eggs Collected |  | Range of Egg Density$\left(\mathrm{n} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | No. | Average Density $\mathrm{n} / \mathrm{m}^{3}$ |  |
| 14-15 May | 3 | 0836-0943 | 46 | 10.6-20.7 | 60.8 | 2 | 0.03 | 0.00-0.04 |
| 16-17 May | 3 | 0638-0802 | 45 | 5.5-22.3 | 59.0 | 1 | 0.01 | 0.00-0.01 |
| 18-19 May | 3 | 0311-0711 | 45 | 16.5-40.7 | 61.4 | 10 | 0.12 | 0.06-0.31 |
| 20-21 May | 1 | 0752-0802 | 10 | 50.6 | 61.2 | 4 | 0.04 | 0.04 |
| 24-25 May | 1 | 0124-0139 | 15 | 73.0 | 62.6 | 2 | 0.02 | 0.02 |
| 26-27 May | 1 | 0236-0251 | 15 | 73.0 | 62.6 | - | 0.00 | 0.00 |
| 28-29 May | 1 | 0001-0016 | 15 | 66.1 | 63.4 | 3 | 0.06 | 0.06 |
| 1-2 Jun | 5 | 0145-0641 | 75 | 5.4-29.5 | 63.3 | 18 | 0.22 | 0.00-0.84 |
| 3-4 Jun | 3 | 2329-1051 | 45 | 10.6-53.4 | 66.0 | 32 | 0.26 | 0.06-0.31 |
| 5-6 Jun | 2 | 0718-0802 | 30 | 5.3-49.2 | 66.2 | 31 | 0.34 | 0.29-1.23 |
| 7-8 Jun | 2 | 0804-0858 | 30 | 5.4-30.1 | 68.4 | 1 | 0.01 | 0.00-0.01 |
| 9-10 Jun | 2 | 0736-0815 | 30 | 5.4-30.1 | 69.9 | 19 | 0.14 | 0.00-0.57 |
| 11-12 Jun | 1 | 1318-1333 | 15 | 10.7 | 72.5 | 1 | 0.24 | 0.24 |
| 12-13 Jun | 3 | 0827-1002 | 45 | 10.7-30.1 | 72.9 | 21 | 0.19 | 0.13-0.42 |
| 19-20 June | 2 | 1118-1203 | 30 | 29.6 | 78.8 | - | 0.00 | 0.00 |

* Flow's from Conowingo Hydroelectric Station approximately 1.5 hr prior to initiation of sampling

C-4. NUMBER AND DENSITY (N/M3) OF AMERICAN SHAD EGGS TAKEN BY ANCHORED 0.5 M PLANKTON NETS FISHED NEAR TELEMETERED AMERICAN SHAD AND/OR SPLASHING ACTIVITY ASSOCIATED WITH SPAWNING BELOW CONOWINGO DAM, 20 MAY THROUGH 6 JUNE 1983.

| Date* | Location |  | $\begin{gathered} \text { Water Temp } \\ \text { (F) } \end{gathered}$ | Depth |  | Time Span | Water Velocity Surface (ft/s) | Flow** | Eggs Collected |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | X | Y |  | Water Column | Net |  |  |  | No. | Density $\mathbf{n} / \mathbf{m}^{3}$ |
| 20-21 May | 975 | 7330 | 60.8 | 7 | - | 2210-2220 | 2.4 | 51.4 | 62 | 0.92 |
| 22-23 May | 1200 | 7000 | 61.2 | 9 | 9 | 0129-0144 | - | 62.5 | 3 | 0.04 |
| 28-29 May | 1100 | 8000 | 63.4 | 13 | 13 | 2115-2130 | 2.6 | 66.1 | 16 | 0.32 |
| 28-29 May | 400 | 900 | - | 4 | 4 | 2209-2224 | - | 66.1 | - | - |
| 28-29 May | -100 | 1500 | - | 4 | 4 | 2318-2324 | - | 66.1 | - | - |
| 29-30 May | 1150 | 9000 | 63.4 | 19 | Surface | 0135-0150 | 0.4 | 30 | - | - |
| 30-31 May | 970 | 7650 | 63.5 | 7 | Surface | 2134-2149 | 1.8 | 63.6 | 22 | 0.22 |
| 30-31 May | 970 | 7650 | 63.5 | 7 | 7 | 2154-2209 | 1.8 | 63.6 | 5 | 0.33 |
| 1-2 Jun | -50 | 1475 | 64.3 | 5 | 5 | 21J5-2150 | 1.6 | 62.9 | 2 | 0.03 |
| 1-2 Jun | -50 | 1475 | 64.3 | 4 | Surface | 2200-2215 | 0.7 | 39.1 | 123 | 2.79 |
| 1-2 Jun ${ }^{\text {a }}$ | 1000 | 7500 | 68 | - | Surface (tow) | 1940-1945 | - | 62.9 | 2 | 0.04 |
| 1-2 Jun ${ }^{\text {a }}$ | 1000 | 7500 | 68 | 9 | 9 | 2025-2037 | - | 62.9 | 24 | 0.58 |
| 1-2 Jun ${ }^{\text {a }}$ | 1000 | 7500 | 68 | 9 | 9 | 2048-2100 | - | 62.9 | 2 | 0.03 |
| 1-2 Jun ${ }^{\text {a }}$ | 1000 | 7850 | 68 | 10 | 10 | 2110-2129 | - | 62.9 | 84 | 2.63 |
| 1-2 Jun ${ }^{\text {a }}$ | 1000 | 7850 | 68 | 10 | Surface | 2133-2147 | - | 62.9 | 23 | 0.41 |
| 2-3 Jun | 900 | 7100 | 64.1 | - | Surface | 0212-0227 | - | 5.4 | - | - |
| 3-4 Jun | 260 | 1505 | 66.2 | 3 | Surface | 2113-2128 | 1.8 | 21.1 | 1 | 0.01 |
| 5-6 Jun | 1000 | 7750 | 66.2 | 8 | Surface | 2233-2248 | 1.6 | 64.4 | 127 | 1.34 |
| 5-6 Jun | 1000 | 7750 | 66.2 | 8 | Surface | 2256-2311 | 1.5 | 64.4 | 23 | 0.26 |
| 5-6 Jun | 1000 | 7800 | 66.2 | 8 | 8 | 0827-0842 | 0.7 | 49.2 | 7 | 0.99 |
| 5-6 Jun | 1000 | 7800 | 66.2 | 8 | 8 | 0848-0903 | 1 | 49.2 | 7 | 0.87 |

* Dates are listed by the night time period

Flows from Conowingo Hydroelectric Station approximately 1.5 -hr prior to initiation of sampling, except for samples taken where $\mathrm{Y}<$ or $=1550 \mathrm{~m}$ these are reported as current generation status
Collected by Delmarva Ecological Laboratory, Inc.

## APPENDIX D-ELECTROFISHING

## D-1. TOTAL HOURS ELECTROFISHING BY MONTH AND LOCATION

1982: Total Number Hours of Electrofishing: All Locations

| 1982 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| May | 1.1 | -- | -- | 1.0 | 2.1 |
| June | 1.5 | -- | -- | 2.0 | 3.5 |
| July | 1.8 | 2.0 | 2.0 | 2.0 | 7.8 |
| August | 3.8 | 1.5 | 2.0 | 4.9 | 12.3 |
| September | 3.8 | 1.8 | 2.0 | 2.0 | 9.6 |
| October | 3.8 | -- | -- | 4.9 | 8.7 |
| November | 1.9 | -- | -- | 2.0 | 3.9 |
| December | 1.9 | -- | -- | 2.0 | 3.9 |
| Total | 19.7 | 5.3 | 6.0 | 20.9 | 51.8 |

1983: Total Number Hours of Electrofishing: All Locations

| 1983 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| January | 1.7 |  |  | 1.8 | 3.6 |
| May | 0.8 |  |  | 2.0 | 2.8 |
| June | 2.0 |  |  | 2.0 | 4.0 |
| July | 2.0 | 1.8 | 2.0 | 2.5 | 8.3 |
| August | 2.0 | 2.0 | 2.0 | 2.5 | 8.5 |
| September | 2.0 | 2.0 | 2.0 | 2.5 | 8.5 |
| October | 2.0 |  |  | 2.5 | 4.5 |
| November | 2.0 |  |  | 2.0 | 4.0 |
| December |  |  |  |  |  |
| Total | 14.5 | 5.8 | 6.0 | 17.8 | 44.1 |

D-1. Cont.

1984: Total Number Hours of Electrofishing: All Locations

| 1984 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| January | 1.5 |  |  | 1.5 | 3.0 |
| March | 1.9 |  |  | 1.3 | 3.2 |
| May |  |  |  |  |  |
| June | 1.9 |  |  | 1.9 | 3.9 |
| July | 2.0 | 2.0 | 2.0 | 2.0 | 8.0 |
| August | 1.8 | 2.0 | 2.0 | 2.5 | 8.3 |
| September | 1.9 | 2.0 | 2.0 | 2.5 | 8.4 |
| October | 2.0 |  |  | 2.5 | 4.5 |
| November | 2.0 |  |  | 2.4 | 4.4 |
| December | 1.9 |  |  | 2.0 | 3.9 |
| Total | 17.0 | 6.0 | 6.0 | 18.6 | 47.5 |

1985: Total Number Hours of Electrofishing: All Locations

| 1985 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| January |  |  |  |  |  |
| March |  |  |  |  |  |
| May |  |  |  |  |  |
| June | 2.0 | 2.0 | 2.0 | 2.0 | 8.0 |
| July | 2.1 | 2.0 | 1.9 | 2.0 | 7.9 |
| August | 2.0 | 1.5 | 1.9 | 2.5 | 7.9 |
| September | 2.0 |  |  | 1.5 | 3.5 |
| October | 2.0 |  |  | 2.0 | 4.0 |
| November | 2.0 |  |  | 2.0 | 4.0 |
| December | 1.8 |  |  | 2.0 | 3.8 |
| Total | 13.8 | 5.5 | 5.8 | 14.0 | 39.1 |

D-1. Cont.

1986: Total Number Hours of Electrofishing: All Locations

| 1986 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| January |  |  |  |  |  |
| March | 1.7 |  |  |  | 1.7 |
| April | 1.9 |  |  |  | 1.9 |
| May | 2.0 |  |  | 2.0 | 3.9 |
| June | 2.0 | 2.0 | 1.8 | 2.5 | 8.3 |
| July | 2.0 | 2.0 | 2.0 | 2.0 | 8.0 |
| August | 1.9 | 2.0 | 1.9 | 2.5 | 8.3 |
| September | 1.9 | 2.0 | 1.8 | 2.5 | 8.3 |
| October | 2.0 |  |  | 2.0 | 4.0 |
| November | 1.8 |  |  | 1.3 | 3.1 |
| December | 1.9 |  |  | 1.3 | 3.3 |
| Total | 19.0 | 8.0 | 7.6 | 16.1 | 50.6 |

1987: Total Number Hours of Electrofishing: All Locations

| 1987 Month | Cono Tailrace | Lees Ferry | The Pool | Tidal <br> Zone | Total Hours Shocked |
| :---: | :---: | :---: | :---: | :---: | :---: |
| January |  |  |  |  |  |
| March |  |  |  |  |  |
| April |  |  |  |  |  |
| May |  |  |  |  |  |
| June |  |  |  | 2.0 | 4.0 |
| July | 2.0 |  |  | 2.0 | 4.0 |
| August | 2.0 |  |  | 1.8 | 2.2 |
| September | 2.2 |  |  |  | 3.8 |
| October | 2.0 |  |  | 5.8 |  |
| November |  |  |  |  | 13.9 |
| December |  |  |  |  |  |
| Total | 8.1 |  |  |  |  |

D-2. Electrofishing CPUE (fish per hr) by species and year hightest to lowest

|  | Species | January | $\begin{aligned} & \hline \hline \text { January } \\ & \text { CPUE } \end{aligned}$ | February | Febuary | March | March CPUE | April | $\begin{aligned} & \hline \hline \text { April } \\ & \text { CPUE } \end{aligned}$ | May | May | June | $\begin{aligned} & \hline \hline \text { June } \\ & \text { CPUE } \end{aligned}$ | July | July CPUE | August | August CPUE | September | September CPUE | October | October CPUE | Novmber | November CPUE | December | December <br> CPUE | Total | $\begin{aligned} & \hline \hline \text { Totat } \\ & \text { CPUE } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | $\begin{aligned} & \text { Gizzard } \\ & \text { Shad } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 70 | 35.5 | 75 | 37.5 | 20,264 | 9,425.1 | 252 | 126.6 |  |  |  |  | 20,661 | 2,547.6 |
| 1983 | $\begin{array}{\|l\|l\|} \hline \text { Gizzard } \\ \text { Shad } \end{array}$ | 16 | 9.4 |  |  |  |  |  |  | 149 | 179.5 | 377 | 193.3 | 315 | 157.5 | 206 | 103.0 | 143 | 71.5 | 171 | 85.5 | 5,475 | 2,737.5 |  |  | 6,852 | 472.9 |
| 1984 | $\begin{aligned} & \hline \text { Gizzard } \\ & \text { Shad } \end{aligned}$ | 114 | 75.2 |  |  | 3 | 1.6 |  |  |  |  | 1,384 | 715.9 | 46 | 23.0 | 117 | 63.8 | 311 | 162.3 | 1,103 | 560.8 | 3,703 | 1,851.5 | 20 | 10.3 | 6,801 | 400.5 |
| 1985 | $\begin{array}{\|l\|} \hline \text { Gizzard } \\ \text { Shad } \end{array}$ |  |  |  |  |  |  |  |  |  |  | 558 | 279.0 | 38 | 18.5 | 43 | 21.5 | 31 | 15.5 | 1,921 | 960.5 | 1,867 | 933.5 | 7 | 3.9 | 4,465 | 322.8 |
| 1982 | Carp |  |  |  |  |  |  |  |  | 3,565 | 32,90.9 | 105 | 68.5 | 60 | 33.3 | 251 | 65.8 | 157 | 41.1 | 75 | 19.7 | 64 | 34.3 | 84 | 43.8 | 4,361 | 221.9 |
| 1983 | American <br> Eel | 55 | 32.2 |  |  |  |  |  |  | 225 | 271.1 | 275 | 141.0 | 500 | 25 | 225 | 112.5 | 525 | 262.5 | 105 | 52.5 | 990 | 495.0 |  |  | 2,900 | 200.1 |
| 1987 | $\begin{aligned} & \text { American } \\ & \text { Eell } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 305 | 154.8 | 320 | 16 | 690 | 320.9 | 145 | 72.9 |  |  |  |  | 1,460 | 18 |
| 1982 | $\begin{array}{l\|l\|} \hline \text { Gizzard } \\ \text { Shad } \end{array}$ |  |  |  |  |  |  |  |  | 44 | 40.6 | 12 | 7.8 | 28 | 15.6 | 298 | 78.1 | 56 | 14.7 | 266 | 69.7 | 296 | 158.6 | 1,907 | 994.9 | 2,907 | 147.9 |
| 1984 | American <br> Eel | 125 | 82.4 |  |  | 80 | 42.5 |  |  |  |  | 330 | 170.7 | 300 | 15 | 260 | 141.8 | 425 | 221.7 | 560 | 284.7 | 250 | 125.0 | 96 | 49.7 | 2,426 | 142.8 |
| 1985 | White Perch |  |  |  |  |  |  |  |  |  |  | 411 | 205.5 | 618 | 301.5 | 301 | 150.5 | 477 | 238.5 | 155 | 77.5 | 7 | 3.5 |  |  | 1,969 | 142.4 |
| 1982 | $\begin{aligned} & \text { American } \\ & \text { Eel } \end{aligned}$ |  |  |  |  |  |  |  |  | 275 | 253.9 | 170 | 110.9 | 460 | 255.6 | 495 | 129.7 | 525 | 137.6 | 290 | 76.0 | 60 | 32.1 | 54 | 28.2 | 2,329 | 118.5 |
| 1985 | $\begin{aligned} & \text { American } \\ & \text { Eel } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  | 200 | 10 | 325 | 158.5 | 185 | 92.5 | 270 | 135.0 | 325 | 162.5 | 250 | 125.0 | 80 | 44.9 | 1,635 | 118.2 |
| 1986 | $\begin{array}{\|l\|} \hline \text { Gizzard } \\ \text { Shad } \end{array}$ |  |  |  |  | 1 | 0.6 | 418 | 223.5 | 729 | 370.1 | 22 | 11.0 | 11 | 5.5 | 48 | 25.0 | 50 | 26.0 | 305 | 156.4 | 423 | 239.0 | 3 | 1.6 | 2,010 | 106.0 |
| 1986 | American <br> Eel |  |  |  |  | 53 | 32.1 | 230 | 123.0 | 375 | 190.4 | 175 | 87.5 | 152 | 76.0 | 200 | 104.2 | 240 | 125.0 | 225 | 115.4 | 49 | 27.7 | 195 | 101.6 | 1,894 | 99.8 |
| 1985 | $\begin{aligned} & \hline \text { Yellow } \\ & \text { Perch } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  | 312 | 156.0 | 234 | 114.1 | 124 | 62.0 | 200 | 10 | 196 | 98.0 | 229 | 114.5 | 57 | 32.0 | 1,352 | 97.8 |
| 1983 | White Perch |  |  |  |  |  |  |  |  | 273 | 328.9 | 295 | 151.3 | 237 | 118.5 | 113 | 56.5 | 295 | 147.5 | 70 | 35.0 | 4 | 2.0 |  |  | 1,287 | 88.8 |
| 1987 | White Perch |  |  |  |  |  |  |  |  |  |  |  |  | 165 | 83.8 | 337 | 168.5 | 164 | 76.3 | 20 | 10.1 |  |  |  |  | 686 | 84.6 |
| 1983 | Yellow Perch | 27 | 15.8 |  |  |  |  |  |  | 45 | 54.2 | 189 | 96.9 | 89 | 44.5 | 75 | 37.5 | 81 | 40.5 | 210 | 105.0 | 419 | 209.5 |  |  | 1,135 | 78.3 |
| 1986 | $\begin{array}{\|l\|} \hline \text { Yellow } \\ \text { Perch } \end{array}$ |  |  |  |  | 5 | 3.0 | 93 | 49.7 | 513 | 260.4 | 185 | 92.5 | 219 | 109.5 | 117 | 60.9 | 85 | 44.3 | 103 | 52.8 | 44 | 24.9 | 6 | 3.1 | 1,370 | 72.2 |
| 1984 | White Perch |  |  |  |  |  |  |  |  |  |  | 360 | 186.2 | 372 | 186.0 | 252 | 137.5 | 103 | 53.7 | 93 | 47.3 | 7 | 3.5 |  |  | 1,187 | 69.9 |
| 1982 | $\begin{aligned} & \hline \text { Yellow } \\ & \text { Perch } \end{aligned}$ |  |  |  |  |  |  |  |  | 104 | 96.0 | 46 | 3 | 130 | 72.2 | 137 | 35.9 | 158 | 41.4 | 291 | 76.2 | 207 | 110.9 | 180 | 93.9 | 1,253 | 63.8 |
| 1986 | White Perch |  |  |  |  |  |  | 72 | 38.5 | 237 | 120.3 | 258 | 129.0 | 170 | 85.0 | 183 | 95.3 | 181 | 94.3 | 64 | 32.8 |  |  |  |  | 1,165 | 61.4 |
| 1982 | White Perch |  |  |  |  |  |  |  |  | 103 | 95.1 | 216 | 140.9 | 88 | 48.9 | 294 | 77.0 | 320 | 83.8 | 114 | 29.9 | 17 | 9.1 |  |  | 1,152 | 58.6 |
| 1983 | Pumpkinseed | 33 | 19.3 |  |  |  |  |  |  | 11 | 13.3 | 165 | 84.6 | 118 | 59.0 | 91 | 45.5 | 109 | 54.5 | 144 | 72.0 | 157 | 78.5 |  |  | 828 | 57.1 |
| 1983 | Bluegill |  |  |  |  |  |  |  |  | 35 | 42.2 | 90 | 46.2 | 200 | 10 | 84 | 42.0 | 187 | 93.5 | 189 | 94.5 | 34 | 17.0 |  |  | 819 | 56.5 |
| 1987 | Yellow <br> Perch |  |  |  |  |  |  |  |  |  |  |  |  | 59 | 29.9 | 91 | 45.5 | 123 | 57.2 | 96 | 48.2 |  |  |  |  | 369 | 45.5 |
| 1985 | $\begin{array}{\|l\|l\|} \hline \text { Channel } \\ \hline \end{array}$ |  |  |  |  |  |  |  |  |  |  | 28 | 14.0 | 75 | 36.6 | 106 | 53.0 | 155 | 77.5 | 100 | 5 | 76 | 38.0 | 83 | 46.6 | 623 | 45.0 |
| 1983 | $\begin{array}{\|l} \hline \text { Channel } \\ \text { Catfish } \\ \hline \end{array}$ | 26 | 15.2 |  |  |  |  |  |  | 70 | 84.3 | 88 | 45.1 | 50 | 25.0 | 51 | 25.5 | 77 | 38.5 | 135 | 67.5 | 143 | 71.5 |  |  | 640 | 44.2 |
| 1987 | Bluegill |  |  |  |  |  |  |  |  |  |  |  |  | 75 | 38.1 | 125 | 62.5 | 90 | 41.9 | 53 | 26.6 |  |  |  |  | 343 | 42.3 |
| 1982 | Channel |  |  |  |  |  |  |  |  | 30 | 27.7 | 48 | 31.3 | 23 | 12.8 | 41 | 10.7 | 290 | 76.0 | 193 | 50.6 | 75 | 40.2 | 115 | 6 | 815 | 41.5 |
| 1984 | Bluegill |  |  |  |  |  |  |  |  |  |  | 81 | 41.9 | 205 | 102.5 | 132 | 72.0 | 169 | 88.2 | 83 | 42.2 | 25 | 12.5 | 2 | 1.0 | 697 | 41.0 |
| 1985 | Bluegill |  |  |  |  |  |  |  |  |  |  | 122 | 61.0 | 103 | 50.2 | 105 | 52.5 | 212 | 106.0 | 13 | 6.5 | 6 | 3.0 | 2 | 1.1 | 563 | 40.7 |
| 1982 | Pumpkkinseed |  |  |  |  |  |  |  |  | 82 | 75.7 | 105 | 68.5 | 77 | 42.8 | 124 | 32.5 | 131 | 34.3 | 103 | 27.0 | 115 | 61.6 | 47 | 24.5 | 784 | 39.9 |
| 1984 | Yellow Perch |  |  |  |  | 10 | 5.3 |  |  |  |  | 136 | 70.3 | 55 | 27.5 | 42 | 22.9 | 57 | 29.7 | 235 | 119.5 | 110 | 55.0 | 9 | 4.7 | 654 | 38.5 |
| 1985 | Pumpkkinseed |  |  |  |  |  |  |  |  |  |  | 150 | 75.0 | 96 | 46.8 | 96 | 48.0 | 103 | 51.5 | 30 | 15.0 | 31 | 15.5 | 9 | 5.1 | 515 | 37.2 |

## D-2. Cont.

|  | Species | January | $\begin{aligned} & \text { January } \\ & \text { CPUE } \end{aligned}$ | February | ${ }^{\text {Febuary }}$ | March | $\begin{aligned} & \text { March } \\ & \text { CPUE } \end{aligned}$ | April | $\begin{array}{\|l\|l\|} \hline \text { April } \\ \text { CPUE } \end{array}$ | May | May <br> CPUE | June | ${ }^{\text {June }}$ | July | July | August | $\begin{aligned} & \hline \text { August } \\ & \text { CPUE } \end{aligned}$ | September | September CPUE | October | October CPUE | Novmber | November CPUE | December | December CPUE | Total | $\stackrel{\text { Totat }}{ }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | Bluegill |  |  |  |  |  |  |  |  | 19 | 17.5 | 69 | 45.0 | 92 | 51.1 | 156 | 40.9 | 158 | 41.4 | 134 | 35.1 | 29 | 15.5 | 1 | 0.5 | 658 | 33.5 |
| 1984 | Pumpkinseed | 6 | 4.0 |  |  | 32 | 17.0 |  |  |  |  | 119 | 61.6 | 174 | 87.0 | 74 | 40.4 | 57 | 29.7 | 74 | 37.6 | 21 | 10.5 | 9 | 4.7 | 566 | 33.3 |
| 1984 | Carp |  |  |  |  |  |  |  |  |  |  | 195 | 100.9 | 75 | 37.5 | 80 | 43.6 | 115 | 6 | 45 | 22.9 | 16 | 8.0 | 26 | 13.4 | 552 | 32.5 |
| 1987 | Striped Bass |  |  |  |  |  |  |  |  |  |  |  |  | 67 | 34.0 | 79 | 39.5 | 96 | 44.7 | 19 | 9.5 |  |  |  |  | 261 | 32.2 |
| 1984 | Channel Catfish |  |  |  |  | 103 | 54.7 |  |  |  |  | 100 | 51.7 | 40 | 2 | 15 | 8.2 | 37 | 19.3 | 81 | 41.2 | 70 | 35.0 | 89 | 46.0 | 535 | 31.5 |
| 1983 | Striped Bass |  |  |  |  |  |  |  |  | 2 | 2.4 | 325 | 166.7 | 61 | 30.5 | 27 | 13.5 | 9 | 4.5 | 25 | 12.5 | 1 | 0.5 |  |  | 450 | 31.1 |
|  | Channel |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Catfish |  |  |  |  | 8 | 4.8 | 38 | 20.3 | 44 | 22.3 | 45 | 22.5 | 57 | 28.5 | 43 | 22.4 | 90 | 46.9 | 100 | 51.3 | 54 | 30.5 | 41 | 21.4 | 520 | 27.4 |
| 1986 | Carp |  |  |  |  | 20 | 12.1 | 180 | 96.3 | 86 | 43.7 | 44 | 22.0 | 52 | 26.0 | 55 | 28.6 | 41 | 21.4 | 37 | 19.0 | 1 | 0.6 | 1 | 0.5 | 517 | 27.3 |
| 1983 | Green Sunfish | 23 | 13.5 |  |  |  |  |  |  | 2 | 2.4 | 25 | 12.8 | 34 | 17.0 | 17 | 8.5 | 20 | 1 | 121 | 60.5 | 146 | 73.0 |  |  | 388 | 26.8 |
| 1985 | Carp |  |  |  |  |  |  |  |  |  |  | 138 | 69.0 | 46 | 22.4 | 42 | 21.0 | 64 | 32.0 | 31 | 15.5 | 14 | 7.0 | 32 | 18.0 | 367 | 26.5 |
| 1986 | Redbreast Sunfish |  |  |  |  | 3 | 1.8 | 17 | 9.1 | 62 | 31.5 | 47 | 23.5 | 56 | 28.0 | 81 | 42.2 | 68 | 35.4 | 111 | 56.9 | 28 | 15.8 | 28 | 14.6 | 501 | 26.4 |
| 1983 | Carp | 60 | 35.1 |  |  |  |  |  |  | 80 | 96.4 | 56 | 28.7 | 70 | 35.0 | 28 | 14.0 | 32 | 16.0 | 32 | 16.0 | 20 | 1 |  |  | 378 | 26.1 |
| 1986 | Bluegill |  |  |  |  | 9 | 5.5 | 17 | 9.1 | 59 | 29.9 | 55 | 27.5 | 74 | 37.0 | 120 | 62.5 | 110 | 57.3 | 47 | 24.1 |  |  | 1 | 0.5 | 492 | 25.9 |
| 1987 | Smallmouth <br> Bass |  |  |  |  |  |  |  |  |  |  |  |  | 20 | 10.2 | 43 | 21.5 | 38 | 17.7 | 74 | 37.2 |  |  |  |  | 175 | 21.6 |
| 1987 | Carp |  |  |  |  |  |  |  |  |  |  |  |  | 29 | 14.7 | 40 | 2 | 73 | 34.0 | 26 | 13.1 |  |  |  |  | 168 | 20.7 |
| 1987 | Channel Catfish |  |  |  |  |  |  |  |  |  |  |  |  | 33 | 16.8 | 40 | 2 | 39 | 18.1 | 56 | 28.1 |  |  |  |  | 168 |  |
|  | Smallmouth |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 20.7 |
| 1983 | Bass |  |  |  |  |  |  |  |  | 36 | 43.4 | 46 | 23.6 | 31 | 15.5 | 28 | 14.0 | 47 | 23.5 | 54 | 27.0 | 12 | 6.0 |  |  | 254 | 17.5 |
| 1986 | Smallmouth <br> Bass |  |  |  |  | 1 | 0.6 | 80 | 42.8 | 21 | 10.7 | 13 | 6.5 | 19 | 9.5 | 39 | 20.3 | 63 | 32.8 | 64 | 32.8 | 13 | 7.3 | 13 | 6.8 | 326 | 17.2 |
| 1986 | Pumpkinseed |  |  |  |  | 13 | 7.9 | 20 | 10.7 | 46 | 23.4 | 71 | 35.5 | 43 | 21.5 | 33 | 17.2 | 36 | 18.8 | 36 | 18.5 | 11 | 6.2 | 15 | 7.8 | 324 | 17.1 |
| 1983 | Redbreast <br> Sunfish | 6 | 3.5 |  |  |  |  |  |  | 6 | 7.2 | 27 | 13.8 | 24 | 12.0 | 9 | 4.5 | 13 | 6.5 | 83 | 41.5 | 76 | 38.0 |  |  | 244 | 16.8 |
| 1987 | Redbreast Sunfish |  |  |  |  |  |  |  |  |  |  |  |  | 21 | 10.7 | 26 | 13.0 | 17 | 7.9 | 68 | 34.2 |  |  |  |  | 132 | 16.3 |
|  | Redbreast |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | Sunfish |  |  |  |  |  |  |  |  |  |  | 33 | 16.5 | 21 | 10.2 | 21 | 10.5 | 22 | 11.0 | 57 | 28.5 | 51 | 25.5 | 10 | 5.6 | 215 | 15.5 |
| 1982 | Smallmouth Bass |  |  |  |  |  |  |  |  | 32 | 29.5 | 48 | 31.3 | 16 | 8.9 | 49 | 12.8 | 65 | 17.0 | 45 | 11.8 | 33 | 17.7 | 5 | 2.6 | 293 | 14.9 |
|  | Green |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1982 | Sunfish |  |  |  |  |  |  |  |  | 6 | 5.5 | 2 | 1.3 | 4 | 2.2 | 7 | 1.8 | 89 | 23.3 | 97 | 25.4 | 43 | 23.0 | 34 | 17.7 | 282 | 14.4 |
| 1987 | Pumpkinseed |  |  |  |  |  |  |  |  |  |  |  |  | 40 | 20.3 | 39 | 19.5 | 13 | 6.0 | 24 | 12.1 |  |  |  |  | 116 | 14.3 |
| 1984 | Green Sunfish | 11 | 7.3 |  |  | 23 | 12.2 |  |  |  |  | 14 | 7.2 | 17 | 8.5 | 15 | 8.2 | 14 | 7.3 | 68 | 34.6 | 40 | 2 | 30 | 15.5 | 232 | 13.7 |
| 1985 | Striped Bass |  |  |  |  |  |  |  |  |  |  | 42 | 21.0 | 30 | 14.6 | 60 | 3 | 34 | 17.0 | 2 | 1.0 | 1 | 0.5 |  |  | 169 | 12.2 |
| 1984 | Smallmouth <br> Bass | 1 | 0.7 |  |  | 4 | 2.1 |  |  |  |  | 41 | 21.2 | 49 | 24.5 | 13 | 7.1 | 29 | 15.1 | 46 | 23.4 | 20 | 1 | 2 | 1.0 | 205 | 12.1 |
| 1986 | Striped Bass |  |  |  |  |  |  |  |  | 5 | 2.5 | 39 | 19.5 | 28 | 14.0 | 60 | 31.3 | 60 | 31.3 | 33 | 16.9 |  |  |  |  | 225 | 11.9 |
| 1982 | Striped Bass |  |  |  |  |  |  |  |  | 1 | 0.9 |  |  | 4 | 2.2 | 77 | 20.2 | 64 | 16.8 | 72 | 18.9 | 1 | 0.5 |  |  | 219 | 11.1 |
| 1984 | Redbreast Sunfish | 4 | 2.6 |  |  | 4 | 2.1 |  |  |  |  | 25 | 12.9 | 19 | 9.5 | 12 | 6.5 | 17 | 8.9 | 62 | 31.5 | 35 | 17.5 | 4 | 2.1 | 182 | 10.7 |
| 1982 | Redbreast Sunfish |  |  |  |  |  |  |  |  | 27 | 24.9 | 11 | 7.2 | 18 | 1 | 24 | 6.3 | 30 | 7.9 | 67 | 17.6 | 22 | 11.8 | 8 | 4.2 | 207 | 10.5 |
| 1984 | Striped Bass |  |  |  |  |  |  |  |  |  |  | 25 | 12.9 | 71 | 35.5 | 33 | 18.0 | 16 | 8.3 | 13 | 6.6 | 1 | 0.5 |  |  | 159 | 9.4 |
| 1986 | Green Sunfish |  |  |  |  |  |  | 6 | 3.2 | 19 | 9.6 | 23 | 11.5 | 31 | 15.5 | 16 | 8.3 | 20 | 10.4 | 41 | 21.0 | 7 | 4.0 | 14 | 7.3 | 177 | 9.3 |
|  | Spottail |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Shiner |  |  |  |  | 3 | 1.8 | 23 | 12.3 | 31 | 15.7 | 15 | 7.5 |  |  |  |  | 12 | 6.3 | 43 | 22.1 | 28 | 15.8 | 21 | 10.9 | 176 | 9.3 |
| 1985 | Spottail Shiner |  |  |  |  |  |  |  |  |  |  | 15 | 7.5 | 5 | 2.4 | 15 | 7.5 |  |  | 22 | 11.0 | 20 | 1 | 39 | 21.9 | 116 | 8.4 |

D-2. Cont.

|  | Species | January | $\begin{array}{\|l\|l\|} \hline \text { Jauary } \\ \text { CPUE } \\ \hline \end{array}$ | February | $\begin{aligned} & \text { Febuary } \\ & \text { CPUEE } \end{aligned}$ | March | $\begin{aligned} & \hline \text { March } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | April | $\begin{aligned} & \text { April } \\ & \text { CPUE } \end{aligned}$ | May | $\begin{aligned} & \hline \text { May } \\ & \text { CPUE } \end{aligned}$ | June | $\begin{aligned} & \text { June } \\ & \text { CPUE } \end{aligned}$ | July | $\begin{aligned} & \text { July } \\ & \text { CPUE } \end{aligned}$ | August | $\begin{aligned} & \text { August } \\ & \text { CPUE } \end{aligned}$ | September | September CPUE | October | $\begin{aligned} & \text { October } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | Novmber | $\begin{aligned} & \text { November } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | December | $\begin{array}{\|l} \hline \text { December } \\ \text { CPUE } \\ \hline \end{array}$ | Total | $\begin{aligned} & \text { Totat } \\ & \text { CPUUE } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Comely |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Shiner |  |  |  |  | 19 | 11.5 |  |  | 20 | 10.2 |  |  |  |  | 2 | 1.0 |  |  | 20 | 10.3 | 18 | 10.2 | 79 | 41.1 | 158 | 8.3 |
| 1983 | Crappie | 1 | 0.6 |  |  |  |  |  |  | 4 | 4.8 | 28 | 14.4 | 6 | 3.0 | 1 | 0.5 | 22 | 11.0 | 39 | 19.5 | 18 | 9.0 |  |  | 119 | 8.2 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Spottail | 49 | 28.7 |  |  |  |  |  |  | 10 | 12.0 |  |  |  |  |  |  | 2 | 1.0 | 22 | 11.0 | 26 | 13.0 |  |  | 109 | 7.5 |
| 1983 | Rock Bass |  |  |  |  |  |  |  |  | 13 | 15.7 | 10 | 5.1 | 11 | 5.5 | 8 | 4.0 | 8 | 4.0 | 15 | 7.5 | 29 | 14.5 |  |  | 94 | 6.5 |
| 1984 | $\begin{aligned} & \begin{array}{l} \text { Spotail } \\ \text { Shinei } \end{array} \end{aligned}$ | 9 | 5.9 |  |  | 10 | 5.3 |  |  |  |  |  |  | 15 | 7.5 | 1 | 0.5 | 3 | 1.6 | 20 | 10.2 | 31 | 15.5 | 19 | 9.8 | 108 | 6.4 |
| 1985 | Smallmouth <br> Bass |  |  |  |  |  |  |  |  |  |  | 16 | 8.0 | 9 | 4.4 | 6 | 3.0 | 12 | 6.0 | 14 | 7.0 | 13 | 6.5 | 9 | 5.1 | 79 | 5.7 |
| 1984 | White Crappie |  |  |  |  |  |  |  |  |  |  | 6 | 3.1 | 45 | 22.5 | 19 | 10.4 | 8 | 4.2 | 6 | 3.1 | 5 | 2.5 | 8 | 4.1 | 97 | 5.7 |
| 1983 | Quillback |  |  |  |  |  |  |  |  | 75 | 90.4 | 4 | 2.1 |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 80 | 5.5 |
| 1982 | Rock Bass |  |  |  |  |  |  |  |  | 8 | 7.4 | 9 | 5.9 | 5 | 2.8 | 9 | 2.4 | 20 | 5.2 | 27 | 7.1 | 18 | 9.6 | 8 | 4.2 | 104 | 5.3 |
| 1985 | Green Sunfish |  |  |  |  |  |  |  |  |  |  | 16 | 8.0 | 5 | 2.4 | 9 | 4.5 | 11 | 5.5 | 12 | 6.0 | 14 | 7.0 | 6 | 3.4 | 73 | 5.3 |
| 1986 | Shad sp |  |  |  |  |  |  | 100 | 53.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 100 | 5.3 |
| 1987 | Rock Bass |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 7 | 3.5 | 9 | 4.2 | 24 | 12.1 |  |  |  |  | 42 | 5.2 |
| 1985 | Rock Bass |  |  |  |  |  |  |  |  |  |  | 13 | 6.5 | 5 | 2.4 | 10 | 5.0 | 10 | 5.0 | 13 | 6.5 | 18 | 9.0 | 2 | 1.1 | 71 | 5.1 |
| 1986 | Alewife |  |  |  |  |  |  | 96 | 51.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 96 | 5.1 |
| 1982 | Largemouth Bass |  |  |  |  |  |  |  |  | 2 | 1.8 | 16 | 10.4 | 13 | 7.2 | 15 | 3.9 | 24 | 6.3 | 16 | 4.2 | 11 | 5.9 | 1 | 0.5 | 98 | 5.0 |
| 1983 | Largemouth |  |  |  |  |  |  |  |  | 3 | 3.6 | 13 | 6.7 | 19 | 9.5 | 9 | 4.5 | 3 | 1.5 | 13 | 6.5 | 12 | 6.0 |  |  | 72 | 5.0 |
| 1982 | Spotifin Shiner |  |  |  |  |  |  |  |  | 7 | 6.5 | 47 | 30.7 | 10 | 5.6 | 6 | 1.6 | 4 | 1.0 | 7 | 1.8 | 3 | 1.6 | 13 | 6.8 | 97 | 4.9 |
| 1986 | Rock Bass |  |  |  |  |  |  | 12 | 6.4 | 27 | 13.7 | 9 | 4.5 | 5 | 2.5 | 9 | 4.7 | 14 | 7.3 | 11 | 5.6 | 3 | 1.7 | 2 | 1.0 | 92 | 4.8 |
|  | Brown Bullhead |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Bullhead | 4 | 2.3 |  |  |  |  |  |  | 7 | 8.4 | 4 | 2.1 | 6 | 3.0 | 13 | 6.5 | 7 | 3.5 | 10 | 5.0 | 19 | 9.5 |  |  | 70 | 4.8 |
| 1987 | $\begin{aligned} & \text { Green } \\ & \text { Sunfis } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 7 | 3.6 | 12 | 6.0 | 6 | 2.8 | 14 | 7.0 |  |  |  |  | 39 | 4.8 |
| 1982 | Spottail |  |  |  |  |  |  |  |  | 3 | 28 | 14 | 91 | 8 | 44 | 10 | 26 | 6 | 16 | 22 | 5.8 | 8 | 43 | 20 | 10.4 | 91 | 46 |
|  | White |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1982 | Crappie |  |  |  |  |  |  |  |  | 4 | 3.7 | 30 | 19.6 | 8 | 4.4 | 22 | 5.8 | 6 | 1.6 | 15 | 3.9 | 5 | 2.7 | 1 | 0.5 | 91 | 4.6 |
| 1982 | Brown Bullhead |  |  |  |  |  |  |  |  | 5 | 4.6 | 1 | 0.7 | 11 | 6.1 | 16 | 4.2 | 14 | 3.7 | 14 | 3.7 | 17 | 9.1 | 7 | 3.7 | 85 | 4.3 |
|  | Comely |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | Shiner | 1 | 0.7 |  |  | 52 | 27.6 |  |  |  |  | 5 | 2.6 | 1 | 0.5 |  |  |  |  |  |  |  |  | 13 | 6.7 | 72 | 4.2 |
| 1984 | Largemouth <br> Bass |  |  |  |  |  |  |  |  |  |  | 4 | 2.1 | 20 | 1 | 25 | 13.6 | 13 | 6.8 | 6 | 3.1 | 1 | 0.5 |  |  | 69 | 4.1 |
| 1984 | Rock Bass |  |  |  |  | 3 | 1.6 |  |  |  |  | 9 | 4.7 | 4 | 2.0 | 2 | 1.1 | 6 | 3.1 | 28 | 14.2 | 14 | 7.0 | 3 | 1.6 | 69 | 4.1 |
| 1982 | Comely <br> Shiner |  |  |  |  |  |  |  |  |  |  | 23 | 15.0 |  |  | 2 | 0.5 | 2 | 0.5 | 2 | 0.5 | 8 | 4.3 | 34 | 177 | 71 | 3.6 |
|  | Comely |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Shiner | 20 | 11.7 |  |  |  |  |  |  | 12 | 14.5 |  |  | 6 | 3.0 |  |  |  |  | 12 | 6.0 | 2 | 1.0 |  |  | 52 | 3.6 |
| 1983 | $\begin{aligned} & \text { Spotfin } \\ & \text { chine } \end{aligned}$ | 1 | 0.6 |  |  |  |  |  |  | 27 | 32.5 | 14 | 7.2 | 10 | 5.0 |  |  |  |  |  |  |  |  |  |  | 52 | 3.6 |
| 1987 | Hybrid |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4 |  |  |  |  |  |  |  |  |  |  | 36 |
|  | Stiped Bass |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Shiner |  |  |  |  | 3 | 1.8 |  |  | 17 | 8.6 | 14 | 7.0 | 14 | 7.0 |  |  | 2 | 1.0 | 6 | 3.1 | 2 | 1.1 | 5 | 2.6 | 63 | 3.3 |
| 1985 | Largemouth Bass |  |  |  |  |  |  |  |  |  |  | 10 | 5.0 | 11 | 5.4 | 9 | 4.5 | 11 | 5.5 | 2 | 1.0 | 2 | 1.0 |  |  | 45 | 3.3 |

D-2. Cont.

|  | Species | January | $\begin{array}{\|l\|} \hline \begin{array}{l} \text { January } \\ \text { CPUE } \end{array} \\ \hline \end{array}$ | February | Febuary CPUE | March | $\begin{aligned} & \hline \text { March } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | April | $\begin{aligned} & \hline \text { April } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | May | $\begin{aligned} & \text { May } \\ & \text { CPUE } \end{aligned}$ | June | $\begin{array}{\|l\|l\|} \hline \text { June } \\ \text { CPUUE } \\ \hline \end{array}$ | July | $\begin{aligned} & \begin{array}{l} \text { July } \\ \text { CPUU } \end{array} \end{aligned}$ | August | $\begin{aligned} & \text { August } \\ & \text { CPUE } \end{aligned}$ | September | September CPUE | October | $\begin{aligned} & \text { October } \\ & \text { CPIII } \end{aligned}$ | Novmber | November CPUE | December | December CPUE | Total | $\begin{aligned} & \hline \text { Totat } \\ & \text { CPUE } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | White |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Sucker |  |  |  |  | 21 | 12.7 | 5 | 2.7 | 3 | 1.5 | 1 | 0.5 |  |  |  |  | 1 | 0.5 | 11 | 5.6 | 12 | 6.8 | 7 | 3.6 | 61 | 3.2 |
| 1985 | White Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 8 | 4.0 | 29 | 163 | 38 | 2.7 |
|  | Brown |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | Bullhead |  |  |  |  |  |  |  |  |  |  | 4 | 2.0 | 6 | 2.9 | 3 | 1.5 | 8 | 4.0 | 6 | 3.0 | 6 | 3.0 | 1 | 0.6 | 34 | 2.5 |
| 1983 | Yellow <br> Bullhead | 1 | 0.6 |  |  |  |  |  |  | 1 | 1.2 |  |  | 4 | 2.0 | 4 | 2.0 | 3 | 1.5 | 10 | 5.0 | 12 | 6.0 |  |  | 35 | 2.4 |
| 1984 | Hybrid Striped Bass |  |  |  |  |  |  |  |  |  |  | 8 | 4.1 | 2 | 1.0 | 9 | 49 | 5 | 26 | 9 | 4.6 | 7 | 3.5 |  |  | 40 | 24 |
| 1983 | Walleye | 5 | 2.9 |  |  |  |  |  |  | 14 | 16.9 | 5 | 2.6 | 2 | 1.0 | 3 | 1.5 | 4 | 2.0 |  |  | 1 | 0.5 |  |  | 34 | 2.3 |
| 1987 | Largemouth |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 25 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Yellow |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2.3 |
| 1982 | Bullhead |  |  |  |  |  |  |  |  | 2 | 1.8 | 2 | 1.3 |  |  | 6 | 1.6 | 13 | 3.4 | 14 | 3.7 | 5 | 2.7 | 2 | 1.0 | 44 | 2.2 |
| 1984 | Spotfin | 3 | 2.0 |  |  | 12 | 6.4 |  |  |  |  | 1 | 0.5 | 19 | 9.5 | 1 | 0.5 | 1 | 0.5 | 1 | 0.5 |  |  |  |  | 38 | 2.2 |
|  | Largemouth |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Bass |  |  |  |  |  |  | 19 | 10.2 | 6 | 3.0 | 5 | 2.5 | 3 | 1.5 | 3 | 1.6 | 2 | 1.0 | 2 | 1.0 | 1 | 0.6 |  |  | 41 | 2.2 |
| 1982 | Walleye |  |  |  |  |  |  |  |  | 3 | 2.8 | 5 | 3.3 | 1 | 0.6 | 1 | 0.3 | 4 | 1.0 | 4 | 1.0 | 14 | 7.5 | 10 | 5.2 | 42 | 2.1 |
| 1985 | Spotfin Shiner |  |  |  |  |  |  |  |  |  |  | 19 | 9.5 | 4 | 2.0 | 3 | 1.5 | 1 | 0.5 |  |  | 1 | 0.5 | 1 | 0.6 | 29 | 2.1 |
| 1987 | Spottail |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 17 | 8.5 |  |  |  |  | 17 | 2.1 |
|  | Comely |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | Shiner |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 |  |  |  |  |  |  | 1 | 0.5 |  |  | 24 | 13.5 | 28 | 2.0 |
| 1987 | Walleye |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 3 | 1.5 | 9 | 4.2 | 1 | 0.5 |  |  |  |  | 15 | 1.8 |
|  | Blueback |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Herring |  |  |  |  |  |  |  |  | 25 | 30.1 |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  | 26 | 1.8 |
| 1986 | Blueback <br> Herring |  |  |  |  |  |  | 34 | 18.2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 34 | 1.8 |
| 1985 | Hybrid |  |  |  |  |  |  |  |  |  |  | 8 | 40 | 4 | 20 | 2 | 10 | 8 | 40 | 2 | 10 |  |  |  |  | 24 | 17 |
| 1986 | Walleye |  |  |  |  | 1 | 0.6 | 12 | 6.4 | 7 | 3.6 | 4 | 2.0 | 1 | 0.5 | 5 | 2.6 |  |  |  |  | 1 | 0.6 |  |  | 31 | 1.6 |
| 1985 | Walleye |  |  |  |  |  |  |  |  |  |  | 16 | 8.0 | 2 | 1.0 |  |  |  |  |  |  | 3 | 1.5 |  |  | 21 | 1.5 |
|  | Brown |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Bullhead |  |  |  |  |  |  | 9 | 4.8 | 8 | 4.1 | 3 | 1.5 | 1 | 0.5 | 2 | 1.0 |  |  | 3 | 1.5 | 1 | 0.6 | 1 | 0.5 | 28 | 1.5 |
| 1982 | Tiger Muskie |  |  |  |  |  |  |  |  | 4 | 3.7 |  |  | 4 | 2.2 | 1 | 0.3 | 4 | 1.0 | 9 | 2.4 | 6 | 3.2 |  |  | 28 | 1.4 |
|  | Hybrid |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Striped Bass |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 | 11 | 5.5 | 1 | 0.5 | 3 | 1.6 | 1 | 0.5 | 8 | 4.1 |  |  |  |  | 26 | 1.4 |
| 1987 | Yellow <br> Bullhead |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 5 | 2.5 |  |  | 4 | 2.0 |  |  |  |  | 11 | 1.4 |
|  | Black |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Crappie |  |  |  |  |  |  |  |  | 2 | 2.4 | 5 | 2.6 | 2 | 1.0 | 1 | 0.5 | 3 | 1.5 | 4 | 2.0 | 2 | 1.0 |  |  | 19 | 1.3 |
| 1986 | Shorthead Redhorse |  |  |  |  |  |  | 3 | 1.6 | 1 | 0.5 | 2 | 1.0 | 1 | 0.5 |  |  | 5 | 2.6 | 4 | 2.1 | 2 | 1.1 | 6 | 3.1 | 24 | 1.3 |
|  | Yellow |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Bullhead |  |  |  |  | 1 | 0.6 | 2 | 1.1 | 6 | 3.0 | 1 | 0.5 | 3 | 1.5 | 3 | 1.6 | 3 | 1.6 | 1 | 0.5 | 2 | 1.1 | 2 | 1.0 | 24 | 1.3 |
| 1985 | Golden Shiner |  |  |  |  |  |  |  |  |  |  | 7 | 3.5 | 3 | 1.5 | 7 | 3.5 |  |  |  |  |  |  |  |  | 17 | 1.2 |
| 1986 | Tessellated <br> Darter |  |  |  |  | 4 | 2.4 |  |  |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.5 |  |  |  |  | 16 | 8.3 | 22 | 1.2 |
| 1985 | White Catfish |  |  |  |  |  |  |  |  |  |  | 6 | 3.0 | 1 | 0.5 |  |  | 3 | 1.5 |  |  | 1 | 0.5 | 5 | 2.8 | 16 | 1.2 |
|  | Black |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1982 | Crappie |  |  |  |  |  |  |  |  | 4 | 3.7 | 7 | 4.6 |  |  | 2 | 0.5 | 2 | 0.5 | 4 | 1.0 | 3 | 1.6 |  |  | 22 | 1.1 |
| 1983 | Golden Shiner |  |  |  |  |  |  |  |  |  |  | 5 | 2.6 | 5 | 2.5 | 2 | 1.0 |  |  | 2 | 1.0 | 2 | 1.0 |  |  | 16 | 1.1 |

D-2. Cont.

|  | Species | January | January CPUE | February | Febuary CPUE | March | $\begin{array}{\|l\|l\|} \hline \text { March } \\ \text { CPUE } \end{array}$ | April | $\begin{aligned} & \hline \text { April } \\ & \text { CPUE } \\ & \hline \end{aligned}$ | May | May CPUE | June | $\begin{array}{\|l} \hline \text { June } \\ \text { CPUE } \end{array}$ | July | July CPUE | August | August CPUE | September | September CPUE | October | October CPUE | Novmber | November CPUE | December | December CPUE | Total | $\begin{aligned} & \hline \text { Totat } \\ & \text { CPUE } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Brown |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | Bullhead |  |  |  |  | 3 | 1.6 |  |  |  |  | 2 | 1.0 | 2 | 1.0 | 2 | 1.1 | 2 | 1.0 | 1 | 0.5 | 6 | 3.0 |  |  | 18 | 1.1 |
| 1984 | Walleye |  |  |  |  | 1 | 0.5 |  |  |  |  | 6 | 3.1 | 1 | 0.5 | 3 | 1.6 | 1 | 0.5 | 2 | 1.0 | 2 | 1.0 | 2 | 1.0 | 18 | 1.1 |
| 1986 | White Catfish |  |  |  |  |  |  | 5 | 2.7 | 1 | 0.5 | 3 | 1.5 | 1 | 0.5 | 6 | 3.1 | 2 | 1.0 |  |  | 2 | 1.1 |  |  | 20 | 1.1 |
| 1983 | White Catish | 1 | 0.6 |  |  |  |  |  |  | 2 | 2.4 | 2 | 1.0 | 2 | 1.0 | 1 | 0.5 | 3 | 1.5 | 2 | 1.0 | 2 | 1.0 |  |  | 15 | 1.0 |
| 1985 | Yellow |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 3 | 1.5 | 1 | 0.5 | 1 | 0.5 | 3 | 1.5 | 3 | 1.5 | 1 | 0.6 | 14 | 1.0 |
| 1985 | White Crappie |  |  |  |  |  |  |  |  |  |  | 6 | 3.0 | 1 | 0.5 | 1 | 0.5 |  |  | 1 | 0.5 | 1 | 0.5 | 3 | 1.7 | 13 | 0.9 |
| 1982 | Golden Shiner |  |  |  |  |  |  |  |  | 2 | 1.8 | 4 | 2.6 | 2 | 1.1 | 6 | 1.6 | 2 | 0.5 | 2 | 0.5 |  |  |  |  | 18 | 0.9 |
| 1987 | Logperch |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 2.5 | 1 | 0.5 | 1 | 0.5 |  |  |  |  | 7 | 0.9 |
| 1987 | Northern Hog Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 | 3.5 |  |  |  |  | 7 | 0.9 |
| 1986 | White Crappie |  |  |  |  |  |  | 2 | 1.1 |  |  | 2 | 1.0 | 2 | 1.0 | 1 | 0.5 | 3 | 1.6 | 4 | 2.1 | 2 | 1.1 |  |  | 16 | 0.8 |
| 1982 | White Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  | 15 | 7.8 | 16 | 0.8 |
| 1985 | Logperch |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 3 | 1.5 | 7 | 3.5 |  |  |  |  |  |  | 11 | 0.8 |
| 1984 | Golden Shiner |  |  |  |  |  |  |  |  |  |  | 3 | 1.6 | 6 | 3.0 | 3 | 1.6 |  |  | 1 | 0.5 |  |  |  |  | 13 | 0.8 |
| 1984 | White Catfish | 1 | 0.7 |  |  |  |  |  |  |  |  | 3 | 1.6 | 2 | 1.0 | 1 | 0.5 |  |  | 1 | 0.5 | 2 | 1.0 | 3 | 1.6 | 13 | 0.8 |
| 1984 | Yellow | 1 | 07 |  |  | 1 | 05 |  |  |  |  | 2 | 10 |  |  |  |  | 1 | 0.5 | 5 | 25 | 2 | 10 | 1 | 0.5 | 13 | 08 |
|  | Brown |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987 | Bullhead |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 4 | 1.9 | 1 | 0.5 |  |  |  |  | 6 | 0.7 |
| 1984 | Tessellated Darter | 7 | 4.6 |  |  | 4 | 2.1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 12 | 0.7 |
| 1982 | White Catish |  |  |  |  |  |  |  |  | 1 | 0.9 | 3 | 2.0 | 3 | 1.7 |  |  | 3 | 0.8 |  |  | 1 | 0.5 | 2 | 1.0 | 13 | 0.7 |
| 1984 | Shorthead |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 10 |  |  |  |  |  |  | 1 | 05 | 8 | 41 | 11 | 06 |
| 1986 | Golden Shiner |  |  |  |  |  |  | 2 | 1.1 | 8 | 4.1 |  |  | 1 | 0.5 |  |  | 1 | 0.5 |  |  |  |  |  |  | 12 | 0.6 |
| 1983 | Mummichog |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 2.5 | 4 | 2.0 |  |  | 9 | 0.6 |
| 1987 | Spotfin Shiner |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 |  |  | 2 | 0.9 |  |  |  |  |  |  | 5 | 0.6 |
| 1982 | Logperch |  |  |  |  |  |  |  |  | 1 | 0.9 |  |  |  |  | 1 | 0.3 | 8 | 2.1 | 1 | 0.3 | 1 | 0.5 |  |  | 12 | 0.6 |
| 1984 | Quillback |  |  |  |  |  |  |  |  |  |  | 10 | 5.2 |  |  |  |  |  |  |  |  |  |  |  |  | 10 | 0.6 |
| 1985 | Tessellated Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 7 | 3.9 | 8 | 0.6 |
| 1986 | Quillback |  |  |  |  |  |  |  |  | 10 | 5.1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 10 | 0.5 |
|  | Atlantic |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987 | Menhaden |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 1 | 0.5 | 1 | 0.5 |  |  |  |  |  |  | 4 | 0.5 |
| 1987 | Comely Shiner |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 2.0 |  |  |  |  | 4 | 0.5 |
| 1987 | Golden Shiner |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 |  |  | 1 | 0.5 |  |  |  |  |  |  | 4 | 0.5 |
| 1987 | White Catfish |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 3 | 1.4 |  |  |  |  |  |  | 4 | 0.5 |
| 1983 | Tiger Muskie |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  | 3 | 1.5 | 2 | 1.0 |  |  | 7 | 0.5 |
| 1983 | White Sucker | 4 | 2.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 1 | 0.5 |  |  | 7 | 0.5 |
| 1984 | Blueback Herring |  |  |  |  |  |  |  |  |  |  |  |  | 7 | 3.5 |  |  |  |  | 1 | 0.5 |  |  |  |  | 8 | 0.5 |

D-2. Cont.

|  | Species | January | January CPUE | February | Febuary CPUE | March | March CPUE | April | $\begin{aligned} & \text { April } \\ & \text { CPIIF } \end{aligned}$ | May | May CPUE | June | June CPUE | July | $\begin{aligned} & \text { July } \\ & \text { CPUE } \end{aligned}$ | August | $\begin{aligned} & \text { August } \\ & \text { CPUE } \end{aligned}$ | September | September CPUE | October | October | Novmber | November CPUE | December | December CPUE | Total | Totat CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | Mummichog |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 | 2 | 0.5 | 4 | 1.0 |  |  | 2 | 1.0 | 9 | 0.5 |
| 1985 | Quillback |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 2.4 |  |  |  |  |  |  |  |  | 1 | 0.6 | 6 | 0.4 |
|  | Shorthead |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Redhorse | 2 | 1.2 |  |  |  |  |  |  | 4 | 4.8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 6 | 0.4 |
| 1983 | Tessellated | 2 | 1.2 |  |  |  |  |  |  | 1 | 1.2 | 2 | 1.0 |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 6 | 0.4 |
| 1984 | Brown Trout |  |  |  |  |  |  |  |  |  |  | 7 | 3.6 |  |  |  |  |  |  |  |  |  |  |  |  | 7 | 0.4 |
|  | Shorthead |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987 | Redhorse |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 |  |  |  |  | 3 | 0.4 |
| 1985 | Atlantic Menhaden |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 | 1 | 0.5 |  |  |  |  |  |  | 1 | 0.5 |  |  | 5 | 0.4 |
|  | Tiger |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | Muskie |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 3 | 1.5 |  |  | 1 | 0.6 | 5 | 0.4 |
| 1982 | Rosyside Dace |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 6 | 1.6 |  |  |  |  | 1 | 0.5 | 7 | 0.4 |
| 1982 | Shorthead Redhorse |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 10 | 2 |  | , |  |  |  |
|  | Black |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.5 |  | 0.4 |
| 1984 | Crappie |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 2 | 1.0 |  |  |  |  |  |  | 2 | 1.0 |  |  | 6 | 0.4 |
|  | Hybrid |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Striped Bass |  |  |  |  |  |  |  |  |  |  | 5 | 2.6 |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 0.3 |
| 1982 | Shield Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0.8 | 1 | 0.3 | 1 | 0.5 | 1 | 0.5 | 6 | 0.3 |
| 1985 | Black Crappie |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  | 4 | 0.3 |
|  | Blueback |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | Herring |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 3 | 1.5 |  |  |  |  |  |  |  |  |  |  | 4 | 0.3 |
| 1985 | Margined Madtom |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.5 |  |  | 4 | 0.3 |
|  | Tessellated |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1982 | Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0.8 | 1 | 0.3 |  |  | 1 | 0.5 | 5 | 0.3 |
| 1987 | White |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  |  |  |  |  |  |  | 2 | 02 |
|  | White |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987 | Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.5 |  |  |  |  | 2 | 0.2 |
| 1984 | White Sucker |  |  |  |  | 2 | 1.1 |  |  |  |  | 1 | 0.5 |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 4 | 0.2 |
| 1985 | Brown Trout |  |  |  |  |  |  |  |  |  |  | 3 | 1.5 |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0.2 |
| 1985 | ${ }^{\text {Cutlips }}$ |  |  |  |  |  |  |  |  |  |  | 1 | 05 |  |  | 1 | 05 |  |  |  |  |  |  | 1 | 06 | 3 |  |
|  | Minnow |  |  |  |  |  |  |  |  |  |  |  | 0.5 |  |  |  | 0.5 |  |  |  |  |  |  |  | 0.6 |  | 0.2 |
| 1986 | Crappie |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  |  |  | 4 | 0.2 |
|  | Commom |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | Shiner |  |  |  |  | 1 | 0.6 |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.5 | 4 | 0.2 |
| 1986 | Northern Hog Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.6 | 3 | 1.6 | 4 | 0.2 |
|  | Margined |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1982 | Madtom |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 | 2 | 0.5 |  |  | 1 | 0.5 | 4 | 0.2 |
| 1984 | Logperch |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  | 3 | 0.2 |
| 1984 | Tiger Muskie |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 2 | 1.0 |  |  |  |  | 3 | 0.2 |
| 1986 | American <br> Shad |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  |  |  | 3 | 0.2 |
| 1986 | Creek Chub |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.6 |  |  | 3 | 0.2 |
| 1986 | Cutlips Minnow |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  | 1 | 0.5 | 3 | 0.2 |
| 1986 | Hickory Shad |  |  |  |  |  |  | 3 | 1.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 0.2 |
| 1986 | River Chub |  |  |  |  | 2 | 1.2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 3 | 0.2 |

D-2. Cont.

|  | Species | January | January | February | Febuary CPUE | March | March CPUE | April | $\begin{gathered} \text { April } \\ \text { CPIFIF } \end{gathered}$ | May | May CPUE | June | June CPUE | July | July CPUE | August | $\begin{aligned} & \text { August } \\ & \text { CPIUF } \end{aligned}$ | September | September CPUE | October | October CPUE | Novmber | November <br> CPUE | December | December CPUE | Total | Totat CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | Alewife |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  | 2 | 0.1 |
| 1985 | Blue Tilapia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  | 2 | 0.1 |
| 1983 | Cutlips Minnow | 1 | 0.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 2 | 0.1 |
| 1983 | Logperch | 1 | 0.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 2 | 0.1 |
| 1983 | Margined <br> Madtom | 1 | 0.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 2 | 0.1 |
| 1987 | Tessellated <br> Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 1 | 0.1 |
| 1987 | Tidewater Silverside |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 1 | 0.1 |
| 1984 | Alewife |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  |  |  |  |  |  |  | 2 | 0.1 |
| 1984 | Margined Madtom |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 | 2 | 0.1 |
| 1984 | Mummichog |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 |  |  |  |  | 2 | 0.1 |
| 1984 | White Mullet |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  | 2 | 0.1 |
| 1986 | Bluntnose Minnow |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1.0 | 2 | 0.1 |
| 1986 | Brown Trout |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 2 | 0.1 |
| 1986 | Mummichog |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 2 | 0.1 |
| 1986 | Tiger Muskie |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.5 |  |  |  |  | 2 | 0.1 |
| 1982 | Brown Trout |  |  |  |  |  |  |  |  | 1 | 0.9 | 1 | 0.7 |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 0.1 |
| 1982 | Creek Chub |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  | 1 | 0.5 | 2 | 0.1 |
| 1982 | Cutlips Minnow |  |  |  |  |  |  |  |  |  |  | 1 | 0.7 |  |  |  |  | 1 | 0.3 |  |  |  |  |  |  | 2 | 0.1 |
| 1982 | Hybrid Striped Bass |  |  |  |  |  |  |  |  | 1 | 0.9 |  |  |  |  |  |  | 1 | 0.3 |  |  |  |  |  |  | 2 | 0.1 |
| 1982 | Northern Hog Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  | 1 | 0.5 | 2 | 0.1 |
| 1985 | Mummichog |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.1 |
|  | Atlantic |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | Menhaden |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1983 | Goldfish | 1 | 0.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1983 | River Chub |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1983 | Shield Darter | 1 | 0.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1984 | American <br> Shad |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.1 |
| 1984 | Blue Tilapia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 1 | 0.1 |
| 1984 | Creek Chub | 1 | 0.7 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1984 | Cutlips Minnow |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 1 | 0.1 |
|  | Northern |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | Hog Sucker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.1 |
| 1986 | Goldfish |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1986 | Hogchoker |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 1 | 0.1 |
| 1986 | Logperch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  | 1 | 0.1 |
| 1986 | Margined |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  | 1 | 0.1 |
| 1986 | Rosyside |  |  |  |  | 1 | 06 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |  |
| 1986 | Shield Darter |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.1 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

D2. Cont

|  | Species | January | January CPUE | February | Febuary CPUE | March | March CPUE | April | April CPUE | May | May CPUE | June | June CPUE | July | July CPUE | August | August CPUE | September | September CPUE | October | October CPUE | Novmber | November CPUE | December | December CPUE | Total | Totat CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | $\begin{array}{\|l\|} \hline \text { Striped Bass } \\ \text { x White } \\ \text { Deat } \end{array}$ <br> Perch |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1982 | $\begin{array}{\|l\|} \hline \text { Atlantic } \\ \text { Menhaden } \\ \hline \end{array}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1982 | Goldfish |  |  |  |  |  |  |  |  |  |  | 1 | 0.7 |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1982 | $\begin{aligned} & \text { Northern } \\ & \text { Pike } \end{aligned}$ Pike |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 |  |  | 1 | 0.1 |
| 1982 | Quillback |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  |  |  |  |  |  |  | 1 | 0.1 |
| 1982 | River Chub |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  |  |  | 1 | 0.1 |
| 1982 | $\begin{array}{\|l} \hline \text { Silvery } \\ \text { Minnow } \end{array}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.5 | 1 | 0.1 |
| 1982 | White Mullet |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 0.3 |  |  |  |  | 1 | 0.1 |
| 1982 | Total | 0 |  | 0 |  | 0 |  | 0 |  | 4,336 | 4,002.6 | 996 | 649.6 | 1,065 | 591.7 | 2,053 | 537.9 | 2,172 | 569.1 | 1,901 | 498.1 | 1,066 | 571.1 | 2,558 | 1,334.6 | 16,147 |  |
| 1983 | Total | 341 | 199.4 | 0 |  | 0 |  | 0 |  | 1,134 | 1,366.3 | 2,056 | 1,054.4 | 1,804 | 902.0 | 998 | 499.0 | 1,594 | 797.0 | 1,480 | 74 | 7,611 | 3,805.5 | 0 |  | 17,018 |  |
| 1984 | Total | 284 | 187.3 | 0 |  | 347 | 184.2 | 0 |  | 0 |  | 2,880 | 1,489.7 | 1,554 | 777.0 | 1,111 | 606.0 | 1,392 | 726.3 | 2,552 | 1,297.6 | 4374 | 2,187.0 | 347 | 179.5 | 14,841 |  |
| 1985 | Total | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 2,149 | 1,074.5 | 1,656 | 807.8 | 1,155 | 577.5 | 1,642 | 821.0 | 2,912 | 1,456.0 | 2625 | 1,312.5 | 410 | 230.3 | 12,549 |  |
| 1986 | Total | 0 |  | 0 |  | 169 | 102.4 | 1,502 | 803.2 | 2,345 | 1,190.4 | 1,048 | 524.0 | 948 | 474.0 | 1,029 | 535.9 | 1,099 | 572.4 | 1,285 | 659.0 | 706 | 398.9 | 466 | 242.7 | 10,597 |  |
| 1987 | Total | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 917 | 465.5 | 1,265 | 632.5 | 21,664 | 10,076.3 | 915 | 459.8 | 0 |  | 0 |  | 24,761 |  |

D-3. YEARLY CPUE AND HOURS SAMPLING SUSQUEHANNA RIVER BELOW CONOWINGO DAM 1982-1987


D-12

D-4 YEARLY OVERALL CPUE1982 - 1987 AT THE CONOWINGO TAILRACE, LEES FERRY, THE POOL AND TIDAL ZONE.

## Conowingo Tailrace



The Pool


Lees Ferry


Tidal Zone


D-5. YEARLY OVERALL CPUE 1982 - 1987 AT THE CONOWINGO TAILRACE, LEES FERRY, THE POOL AND TIDAL ZONE.


D-6. LOWER SUSQUEHANNA SPECIES COMPOSITION, PROPORTIONAL ABUNDANCE AND SEASONALITY IN THE CONOWINGO TAILRACE, BASED ON CPUE OF FIVE MOST FREQUENTLY COLLECTED SPECIES IN EACH YEAR 1982-87


D-7. LOWER SUSQUEHANNA SPECIES COMPOSITION, PROPORTIONAL ABUNDANCE AND SEASONALITY AT LEES FERRY, BASED ON CPUE OF FIVE MOST FREQUENTLY COLLECTED SPECIES IN EACH YEAR 1982-87


D-8. LOWER SUSQUEHANNA SPECIES COMPOSITION, PROPORTIONAL ABUNDANCE AND SEASONALITY AT THE POOL, BASED ON CPUE OF FIVE MOST FREQUENTLY COLLECTED SPECIES IN EACH YEAR 1982-87


D-9. LOWER SUSQUEHANNA SPECIES COMPOSITION SEASONALITY IN THE TIDAL ZONE, BASED ON CPUE OF FIVE MOST FREQUENTLY COLLECTED SPECIES IN EACH YEAR 1982-87


APPENDIX E-GILL NETS

E-1. LOWER SUSQUEHANNA GILL NET CPUE (FISH/NET-NIGHT) IN THE CONOWINGO TAILRACE 1982-84


E-2. LOWER SUSQUEHANNA GILL NET CPUE(FISH/NET-NIGHT) AT LEES FERRY 1982-84


E-3. LOWER SUSQUEHANNA GILL NET CPUE AT (FISH/NET-NIGHT) THE POOL 1982-83


E-4. LOWER SUSQUEHANNA GILL NET CPUE (FISH/NET-NIGHT) AT THE TIDAL ZONE 1982-83


E-5. LOWER SUSQUEHANNA ALL AREAS TOTAL GILL NET CPUE (FISH/NET-NIGHT) 1982-84


E-6. WHITE PERCH CPUE FISH/NET-NIGHT AT THE CONOWINGO TAILRACE 1982-1984.


E-7. WHITE PERCH CPUE FISH/NET-NIGHT AT LEE'S FERRY 1982-1984.


E-8. WHITE PERCH CPUE FISH/NET-NIGHT AT THE TIDAL ZONE 1982 - 1984.


E-9. GIZZARD SHAD CPUE AT CONOWINGO TAILRACE 1982-1984


E-10. GIZZARD SHAD CPUE AT LEE'S FERRY 1982 - 1984


E-11. GIZZARD SHAD CPUE AT THE TIDAL ZONE 1982-1984


E-12. GIZZARD SHAD CPUE AT CONOWINGO TAILRACE 1982-1984.


E-13. GIZZARD SHAD CPUE AT CONOWINGO TAILRACE, LEE'S FERRY AND THE TIDAL ZONE 1982 - 1984.


E-14. GIZZARD SHAD CPUE AT CONOWINGO TAILRACE, LEE'S FERRY AND THE TIDAL ZONE 1982 - 1984.


E-15. NUMBER COLLECTED, PERCENT COMPOSITION, AND CATCH PER EFFORT (NO. PER NET NIGHT) FOR FISHES COLLECTED BY EXPERIMENTAL GILL NETS IN THE LOWER SUSQUEHANNA RIVER, JULY-OCTOBER 1982.

| Area | Tailrace |  |  | Lee's Ferry |  |  | Pool |  |  | Tidal |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Nets Set | 8 |  |  | 11 |  |  | 13 |  |  | 16 |  |  | 48 |  |  |
|  | No. | \% | $\text { CPUE }^{1}$ | No. | \% | $\text { CPUE }^{1}$ | No. | \% | $\text { CPUE }^{1}$ | No. | \% | $\text { CPUE }^{1}$ | No. | \% | $\text { CPUE }^{1}$ |
| Channel catfish | 491 | 58 | 61 | 139 | 58 | 13 | 81 | 44 | 6 | 3 | 4 | 0.2 | 714 | 55\% | 15 |
| White perch | 191 | 23 | 24 | 60 | 25 | 5 | 57 | 32 | 4 | 8 | 11 | 1 | 316 | 24\% | 7 |
| Gizzard shad | 125 | 15 | 16 | 5 | 3 | 0 | 15 | 10 | 1 | 6 | 74 | 0.4 | 151 | 12\% | 3 |
| Shorthead redhorse |  |  |  | 17 | 8 | 2 | 22 | 15 | 2 | 3 | 4 | 0.2 | 42 | 3\% | 1 |
| Carp | 11 | 2 | 1 | 8 | 3 | 1 | 5 | 3 | 0.4 |  |  |  | 24 | 2\% | 1 |
| Hybrid Striped | 7 | 1 | 1 |  |  |  | 1 | 1 | 0.1 |  |  |  | 8 | 1\% | 0.2 |
| Spottail shiner |  |  |  |  |  |  |  |  |  | 8 | 11 | 0.5 | 8 | 1\% | 0.2 |
| Striped bass | 3 | 1 | 0.4 | 4 | 2 | 0.4 | 1 | 1 | 0.1 |  |  |  | 8 | 1\% | 0.2 |
| Atlantic menhaden | 5 | 1 | 1 | 1 | 3 | 0.1 |  |  |  |  |  |  | 6 | 0.5\% | 0.1 |
| White catfish | 1 | 0.4 | 0.1 | 4 | 2 | 0.4 |  |  |  |  |  |  | 5 | 0.4\% | 0.1 |
| White crappie | 3 | 1 | 0.4 |  |  |  |  |  |  |  |  |  | 3 | 0.2\% | 0.1 |
| Blueback herring | 1 | 0.3 | 0.1 |  |  |  | 2 | 4 | 0.2 |  |  |  | 3 | 0.2\% | 0.1 |
| Spot |  |  |  |  |  |  |  |  |  | 2 | 3 | 0.1 | 2 | 0.2\% | 0.04 |
| White sucker |  |  |  |  |  |  | 1 | 1 | 0.1 | 1 | 1 | 0.1 | 2 | 0.2\% | 0.04 |
| Yellow perch | 2 | 1 | 0.3 |  |  |  |  |  |  |  |  |  | 2 | 0.2\% | 0.04 |
| American shad |  |  |  |  |  |  | 1 | 2 | 0.1 |  |  |  | 1 | 0.1\% | 0.02 |
| Largemouth bass |  |  |  |  |  |  |  |  |  | 1 | 1 | 0.1 | 1 | 0.1\% | 0.02 |
| Walleye | 1 | 0.3 | 0.1 |  |  |  |  |  |  |  |  |  | 1 | 0.1\% | 0.02 |
| Yellow bullhead | 1 | 0.4 | 0.1 |  |  |  |  |  |  |  |  |  | 1 | 0.1\% | 0.02 |
| Totals | 842 |  | 105 | 238 |  | 22 | 186 |  | 14 | 32 |  | 2 | 1,298 |  | 27 |

${ }^{1}$ In certain instances the CPUE presented here has been altered from data originally reported in "Annual Report (Article 34; Objective 5): 1982 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" to standardize with 1983 and 1984 data where CPUE is determined as No. of Fish Collected / effort (No. nets set) and includes 'zero catches' as part of the effort. In 1982 if a gill net yielded zero fish of a particular species the effort was not included in determining the species specific CPUE. All data presented here as CPUE is number of fish / number of nets set. The 1983 and 1984 data were derived from "Annual Report (Article 34; Objective 5): 1983 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" and "Annual Report (Article 34; Objective 5): 1984 Fisheries Studies for Determination of Flow Needs for Protection and Enhancement of Fish Populations Below Conowingo" respectively.

E-16. CONT. NUMBER, PERCENT COMPOSITION (\%). AND CATCH PER EFFORT (CPE: NO. PER NET NIGHT) OF FISHES COLLECTED BY EXPERIMENTAL GILL NET IN THE LOWER SUSQUEHANNA RIVER. JULY THROUGH NOVEMBER 1983.

| Area | Tailrace |  |  | Lee's Ferry |  |  | The Pool |  |  | Tidal Zone |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. of Nets Set | 9 |  |  | 12 |  |  | 12 |  |  | 12 |  |  | 45 |  |  |
|  | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| Atlantic menhaden | 4 | 0.4 | 0.4 |  |  |  |  |  |  | 27 | 9 | 2 | 31 | 2 | 1 |
| Gizzard shad | 281 | 30 | 31.2 | 49 | 19 | 4 | 2 | 3 | 0.2 | 120 | 40 | 10 | 452 | 29 | 10 |
| Carp | 10 | 1 | 1.1 | 10 | 4 | 1 |  |  |  | 5 | 2 | 0.4 | 25 | 2 | 1 |
| Shorhead redhorse |  |  |  | 32 | 13 | 3 | 4 | 6 | 0.3 | 10 | 3 | 1 | 46 | 3 | 1 |
| Channel catfish | 381 | 41 | 42.3 | 60 | 24 | 5 | 43 | 60 | 4 | 56 | 18 | 5 | 540 | 35 | 12 |
| White perch | 168 | 18 | 18.7 | 95 | 38 | 8 | 18 | 25 | 2 | 80 | 26 | 7 | 361 | 23 | 8 |
| Striped bass | 65 | 7 | 7.2 | 2 | 1 | 0.2 | 2 | 3 | 0.2 | 2 | 1 | 0.2 | 71 | 5 | 2 |
| Hybrid Striped bass | 11 | 1 | 1.2 |  |  |  | 1 | 1 | 0.1 |  |  |  | 12 | 1 | 0.3 |
| Other Fishes ${ }^{1}$ | 14 | 2 | 1.6 | 5 | 2 | 0.4 | 2 | 3 | 0.2 | 4 | 1.3 | 0.3 | 25 | 2 | 1 |
| Totals | 934 |  | 104 | 253 |  | 21 | 72 |  | 6 | 304 |  | 25 | 1563 |  | 35 |

${ }^{1}$ Other species collected were combined under the category 'Other Fishes' in 1983; the species denoted account for the known minimum species richness

E-17. SPECIES COMPOSITION OF FISHES CAUGHT IN THE EXPERIMENTAL GILL NET PROGRAM FROM THE LOWER SUSQUEHANNA RIVER 19841

| Area | Conowingo Tailrace |  |  | Lees Ferry |  |  | Tidal Zone |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. Nets | 10 |  |  | 5 |  |  | 10 |  |  | 25 |  |  |
|  | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| Alewife |  |  |  |  |  |  | 1 | 0.4 | 0.1 | 1 | 0.4 | 0.04 |
| American shad | 1 | 0.1 | 0.1 |  |  |  | 1 | 0.4 | 0.1 | 2 | 0.2 | 0.1 |
| Atlantic Menhaden |  |  |  |  |  |  | 9 | 4 | 1 | 9 | 4 | 0.4 |
| Black Crappie | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Carp | 39 | 5 | 4 | 8 | 6 | 1.6 | 5 | 2 | 1 | 52 | 4 | 2 |
| Channel Catfish | 279 | 34 | 28 | 44 | 35 | 8.8 | 114 | 45 | 11 | 437 | 37 | 17 |
| Gizzard shad | 220 | 27 | 22 | 6 | 5 | 1.2 | 66 | 26 | 7 | 292 | 25 | 12 |
| Hickory shad |  |  |  |  |  |  | 1 | 0.4 | 0.1 | 1 | 0.4 | 0.04 |
| Hybrid Striped Bass | 64 | 8 | 6 | 2 | 2 | 0.4 | 1 | 0.4 | 0.1 | 67 | 6 | 3 |
| Largemouth Bass | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Quillback | 2 | 0.2 | 0.2 |  |  |  | 1 | 0.4 | 0.1 | 3 | 0.3 | 0.1 |
| Redbreast Sunfish |  |  |  |  |  |  | 1 | 0.4 | 0.1 | 1 | 0.4 | 0.04 |
| Shad |  |  |  |  |  |  | 2 | 1 | 0.2 | 2 | 0.2 | 0.1 |
| Shorthead Redhorse |  |  |  | 14 | 11 | 2.8 | 6 | 2 | 1 | 20 | 5 | 1 |
| Smallmouth Bass | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Spot |  |  |  |  |  |  | 9 | 4 | 1 | 9 | 4 | 0.4 |
| Striped Bass | 19 | 2 | 2 | 1 | 1 | 0.2 | 8 | 3 | 1 | 28 | 2 | 1 |
| Tiger Musky | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Walleye | 3 | 0.4 | 0.3 |  |  |  | 1 | 0.4 | 0.1 | 4 | 0.4 | 0.2 |
| White catfish |  |  |  |  |  |  | 3 | 1 | 0.3 | 3 | 1 | 0.1 |
| White Crappie | 8 | 1 | 1 |  |  |  |  |  |  | 8 | 1 | 0.3 |
| White Perch | 174 | 21 | 17 | 50 | 40 | 10 | 22 | 9 | 2 | 246 | 21 | 10 |
| White sucker | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Yellow bullhead | 1 | 0.1 | 0.1 |  |  |  |  |  |  | 1 | 0.1 | 0.04 |
| Yellow perch | 1 | 0.1 | 0.1 | 1 | 1 | 0.2 |  |  |  | 2 | 0.2 | 0.1 |
| Totals | 816 |  | 82 | 126 |  | 25 | 251 |  | 25 | 1193 |  | 48 |

${ }^{(1)}$ Because of high river flows in 1984 only one collection was made in the Pool, thus data from this location has been excluded from the analysis.

E-18. ANNUAL CATCH PER EFFORT (NUMBER PER SET) OF ALL FISH CAUGHT IN THE EXPERIMENTAL GILL NET PROGRAM FROM THE LOWER SUSQUEHANNA RIVER.

| LOCATION | CPUE |  |  |
| :--- | ---: | ---: | ---: |
|  | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ |
| Conowingo Tailrace | 105 | 103 | 82 |
| Lees Ferry | 22 | 21 | 25 |
| Tidal Zone | 37 | 25 | 25 |
| All Locations | 55 | 45 | 48 |

APPENDIX F-AGE AND GROWTH PLOTS FROM RMC 1985A, B, C. AND LENGTH FREQUENCY CHARTS FROM THE WEST FISH LIFT 2010

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH


Figure 6.1.2

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH


Figure 6.1.A

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH



Figure6.15

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH
 SECIESPPUPRKINSEEO YERR=1982


Figure 6.1.6

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH

SPECTES-REDBREAST SUMFISH TEAR-1962


Figure 6.1.7

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH
subarourto or sprino(A), somere(b), and fact-witer(c)



Figure 6.18

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH

SPECIES=SIRPPED BASS YEAR-1982


Figure 6.19

## PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH

SUBGROURED BY SPRTMO(A), sumeth (G), and fall-winter(c) SPECIES=YMLLEYE YEAR=1902


Figure 6.1.10

## PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH


SPECILSELAROTNOUTH DKSS


Figure 6.1.11

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH
suberoupeo ay 5palua(A), sunuen(0), aho tall-minter(0)
spteitseatutolt Yeatalgas
PCRCRTAGE HLOCK CHART


Figure6.1.12

PERCENTAGE BY AREA OF LENGTH FREQUENCY OF ELECTROSHOCKED FISH
 SPECIES=EREEN SUMF I SH YEAR=1982


Figure 6.1.13


FIGURE 5.1-2,
Seasonal length frequency distribution (expressed as a percentage) of white perch taken by electrofishing below Conowingo Dam, 1983. Lee's Ferry and the Pool areas were combined (River) due to small sample size.


FIGURE 5,1-3.
Seasonal length frequency distribution (expressed as a percentage) of channel catfish taken by electrofishing below Conowingo Dam, 1983. Lee's Ferry and the Pool areas were combined (River) due to small sample size.


FIGURE 5.1-4.
Seasonal length frequency distribution (expressed as a percentage) of
gizzard shad taken by electrotishing below Conowingo Dam, 1983. Lee's
Ferry and the Pool areas were combined (River) due to small sample size.



figure 5.1-7.
Seasonal length frequency distribution (expressed as a percentage) of striped bass taken by electrofishing below Conowingo Dam, 1983. Lee's Ferry and the Pool areas were combined (River) due to small sample size.

## 1



FIGURE 5.1-8.
Seasonal length frequency distribution (expressed as a percentage) of
walleye taken by electrofishj.ng below Conowingo Dant, 1983. Lee's
Ferry and the Pool areas were combined (River) due to small sample size.


Length frequencies of striped bass, expressed as a percentage of the total number caught, collected by electrofishing in of the total number caught, collected by electrof ishing in Tidal Zone (lower) in 1982 (2) and 1983 (3).


FIGURE 5.1-11.
Length frequencies of white perch, expressed
as a percentage of the total number caught,
coilected by electrofishing in the Conowing
coliected by electrof siding in the Conowing
Tain race (upper) and T.
1982 (2) and 1983 ( 3 ).



FIGURE 4.1-7.
Seasonal length frequency distribution of channel catfish collected at the Conowingo Dam Fish Passage Facility in 1984.



FIGURE 4.1-9.
Seasonal length frequency distribution of walleye collected at the Conowingo Dam Fish Passage Facility in 1984.


## FIGURE 4.1-10.

ancy distribution of striped bass collected at the Conowingo Dam Fish Passage Facility in 1984.
Seasonal length frequency distribution of striped bass collected at the Conowing


FIGURE 4.1-11.
Seasonal length frequency distribution of striped bass x white bass hybrid collected at the Conowingo Dam Fish Passage Facility in 1984.




FIGURE 5.2-5.
Percent composition of length groups of striped bass collected by experimental gill net in the lower Susquehanna River, July through October, 1984. ( $\mathrm{N}=29$ ).

F-30. CHANNEL CATFISH WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative No. of Fish |
| :---: | :---: | :---: |
| 71 to 170 mm | 9 | $3 \%$ |
| 171 to 210 mm | 9 | $3 \%$ |
| 271 to 370 mm | 103 | $30 \%$ |
| 371 to 470 mm | 191 | $56 \%$ |
| 471 to 570 mm | 26 | $8 \%$ |
| 571 to 670 mm | 1 | $0 \%$ |

Length Frequency Distribution Channel Catfish


## F-31. RED BREAST SUNFISH WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative <br> No. of <br> Fish |
| :---: | :---: | :---: |
| 100 to 125 mm | 0 | $0 \%$ |
| 125 to 150 mm | 0 | $0 \%$ |
| 150 to 175 mm | 4 | $5 \%$ |
| 175 to 200 mm | 57 | $68 \%$ |
| 200 to 225 mm | 23 | $27 \%$ |
| 225 to 250 mm | 0 | $0 \%$ |

Length Frequency Distribution Red Breast Sunfish


F-32. BLUEGILL WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative <br> No. of <br> Fish |
| :---: | :---: | :---: |
| 100 to 125 mm | 1 | $1 \%$ |
| 125 to 150 mm | 17 | $23 \%$ |
| 150 to 175 mm | 34 | $47 \%$ |
| 175 to 200 mm | 28 | $38 \%$ |
| 200 to 225 mm | 11 | $15 \%$ |
| 225 to 250 mm | 0 | $0 \%$ |

Length Frequency Distribution Bluegill


F-33. SMALLMOUTH BASS WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative <br> No. of <br> Fish |
| :---: | :---: | :---: |
| 141 to 230 mm | 0 | $0 \%$ |
| 231 to 320 mm | 1 | $2 \%$ |
| 321 to 410 mm | 31 | $52 \%$ |
| 411 to 500 mm | 28 | $47 \%$ |

Length Frequency Distribution Smallmouth Bass


F-34. LARGEMOUTH BASS WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative <br> No. of <br> Fish |
| :---: | :---: | :---: |
| 250 to 275 mm | 0 | $0 \%$ |
| 275 to 300 mm | 2 | $11 \%$ |
| 300 to 325 mm | 5 | $28 \%$ |
| 325 to 350 mm | 2 | $11 \%$ |
| 350 to 375 mm | 4 | $22 \%$ |
| 375 to 400 mm | 7 | $39 \%$ |
| 400 to 425 mm | 5 | $28 \%$ |
| 425 to 450 mm | 0 | $0 \%$ |

## Length Frequency Distribution Largemouth Bass



F-35. YELLOW PERCH WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. <br> Fish | Relative No. of Fish |
| :---: | :---: | :---: |
| 75 to 100 mm | 0 | $0 \%$ |
| 100 to 125 mm | 1 | $5 \%$ |
| 125 to 150 mm | 1 | $5 \%$ |
| 150 to 175 mm | 4 | $21 \%$ |
| 175 to 200 mm | 7 | $37 \%$ |
| 200 to 225 mm | 4 | $21 \%$ |
| 225 to 250 mm | 2 | $11 \%$ |
| 250 to 275 mm | 1 | $5 \%$ |
| 275 to 300 mm | 1 | $5 \%$ |
| 300 to 325 mm | 0 | $0 \%$ |

Length Frequency Distribution Yellow Perch


F-36. WALLEYE WEST FISH LIFT LENGTH FREQUENCY 2010

| Length Class | No. Fish | Relative <br> No. of <br> Fish |
| :---: | :---: | :---: |
| 171 to 240 mm | 1 | $1 \%$ |
| 241 to 310 mm | 21 | $11 \%$ |
| 311 to 380 mm | 16 | $8 \%$ |
| 381 to 450 mm | 79 | $42 \%$ |
| 451 to 520 mm | 66 | $35 \%$ |
| 521 to 590 mm | 7 | $4 \%$ |

## Length Frequency Distribution Walleye



APPENDIX G-LENGTH AND WEIGHT RELATIONSHIPS

G-1 WALLEYE LENGTH X WEIGHT REGRESSION COMPARISON 1982-1987 TO 2010


G-2. CHANNEL CATFISH LENGTH X WEIGHT REGRESSION COMPARISON 1982-1987 TO 2010


G-3. SMALLMOUTH BASS LENGTH X WEIGHT REGRESSION COMPARISON 1982-1987 TO 2010


G-4. LARGEMOUTH BASS LENGTH X WEIGHT REGRESSION COMPARISON 1982-1987 TO 2010


G-5. YELLOW PERCH LENGTH X WEIGHT REGRESSION COMPARISON 1982-1987 TO 2010



[^0]:    (2) Dominant taxa ( $>10.0 \%$ of the total in one or more collections) are entered in bold print

